GRAIP_Lite: A System for Road Impact Assessment

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Abstract

GRAIP_Lite is a system of tools developed for ArcGIS that is used to model road-related sediment impacts to stream habitats. GRAIP_Lite uses a topographic model, along with other inputs, to create road segments, applies average vegetation parameters and calculates sediment production from individual road segments, uses a local polynomial fit to describe stream connection probabilities and fractional sediment delivery based on flow distance to streams, and accumulates routed sediment throughout the modelled stream network. Road-related sediment impact is described using specific sediment (Mg/yr/km2) in the modelled stream network. This metric can easily be used to determine areas where roads present a higher risk to stream habitats when prioritizing areas for restoration or remediation efforts. When used for alternatives analysis, GRAIP Lite allows the user to specify various treatment options for individual roads and then models the road-related sediment conditions at the initial condition (before work begins), disturbed condition (immediately post-work or during haul), and the recovered condition (once vegetation has recovered to normal values). GRAIP Lite also has reporting tools that generate basic maps for use in reports based on the model results.

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Introduction

Fine sediment accumulations in streams reduce the quality of critical stream habitats by impairing spawning gravels and reducing productivity. Forest roads are common sources of chronic fine sediment delivery in otherwise clear mountain streams. While the amounts delivered may be small compared to long-term sedimentation rates and sediment transport capacity (Goode et al., 2012), short-term roadrelated sediment can be significant and represent a large portion of the stream's short-term sediment load. Furthermore, the chronic nature of road erosion inputs may have important consequences for fish habitat (Maturana et al., 2014).

Road density is one tool that has been used to estimate road-related sediment risks to streams. It is generally assumed that where there are higher densities of roads there should also be greater roadrelated sedimentation, and to some extent this is true. The issue is that this approach treats all roads and environments as equals, even though anyone who has spent time exploring or working on forest roads knows that not all roads are created equal; there is substantial variability in road design, location, and maintenance practices (e.g. Luce et al., 2001). The one benefit of the road density approach is that it is quick, easy, and inexpensive to apply.

Road inventories are a fairly direct approach to identifying the unique qualities of each road. For instance, the Geomorphic Roads Analysis and Inventory Package (GRAIP, Black et al., 2012, Cissel et al., 2012) uses a GPS inventory to map the road network hydrology at the scale of individual road segments and drainpoints. An empirical model is used to estimate road sediment production on individual road segments, and the estimated sediment production is routed to the observed drainpoints and to the stream network, if the drainpoint was observed to be hydrologically connected to the stream network. This allows GRAIP to provide a detailed map of road sediment production, delivery, and accumulation in streams within a watershed. GRAIP also addresses a number of other potential road related risks, including landslide and gully initiation risks and stream crossing problems. The detailed information comes at greater cost than road density information; however such costs are a tiny fraction of actual road treatment costs, and targeting treatments to where they will have the greatest outcomes can more than offset inventory expenses. The amount of information provided by field surveys is suitable for project planning and some design level decisions. Such detail may be much greater than is necessary for many broad planning-level exercises.

Practical, intermediate information-content tools would be useful to managers. GRAIP_Lite leverages empirical patterns and information from several watershed studies done using GRAIP inventories and analyses to substantially refine the information that can be extracted from GIS map-level information. In keeping with a modeling philosophy that seeks to match decision maker information needs to modeling effort (Figure 1), GRAIP Lite can be implemented using only existing GIS data, or it can be supported using a field-collected calibration data set to inform a statistical estimation of vegetation cover and delivery curves. Existing GRAIP data can be used to supply this calibration data set or new calibration data can be collected in areas where existing GRAIP data is not available or applicable. We also offer a calibration survey design that takes less time and expense than complete inventories. Again, the intent is to provide a spectrum of tools to allow managers to match their needs for precision and accuracy to the level of effort expended (Figure 1).

Density	Uncalibrated	Calibrated	GRAIP
Line presence	GRAIP_Lite	GRAIP Lite	GPS Roads
	Road Line Data DEM Slope Streams	Road Line Data DEM Slope Streams Improved Vegetation & Delivery	Surfacing Existence Condition GPS Drainpoints Type Location Condition Connection DEM Slope Streams Erodibility

Figure 1: The information-effort scale for road erosion estimation.

Calculations

Supporting Data

GRAIP_Lite was developed from data associated with several large GRAIP watershed inventories, which provided the means of determining the underlying relationships between sediment delivery and flow distance to streams. Over 77,000 drainpoint observations were used to test various predictive variables and model relationships. Road segment data was used to derive average vegetation states for different surface type and maintenance level classes. The more detailed GRAIP model provided the basic structure, with the GRAIP_Lite model being similarly composed of separate components for sediment production, sediment delivery, and sediment accumulation.

GRAIP_Lite uses a single flow path road model where each road segment drains completely at a single drainpoint located at the downhill end of the road segment. This is a simplification from the GRAIP model, which allows two flow paths and more complex flow routing through multiple road segments, but without the complete road inventory there is no way to predict such routing. What can be done, and what GRAIP Lite does, is to use topography and maintenance level as the bases to break up the road network into individual road segments, each of which can be assigned a drainpoint, model sediment production from that road segment based on several factors, model sediment delivery based on connection probabilities, and then accumulate the delivered sediment through the stream network to assess sediment loads and risks.

Components

GRAIP_Lite estimates sediment production and delivery separately. The first component calculates an estimate of the amount of sediment produced from each segment of, and the second component then models how much of that sediment is actually delivered to the stream network, where it can impact aquatic habitat.

Sediment Production

Sediment production is the amount of sediment eroded from the road surface, and is estimated for each road segment as

E = *BRSV*

where *E* is the total sediment production for the road segment, *B* is the baserate, *R* is the elevation difference between road segment ends, *S* is the surface type factor, and *V* is the vegetation factor. These components are shown in Figure 2.

Figure 2: Components of the Sediment Production equation.

Baserates

The baserate is used to describe the erodibility of the road surface and the power of rainfall to detach and transport sediment, and is used to take into account different variables including local geologic and climatic factors. In most cases, baserates are measured using plot studies, as described in Black and Luce (2013). In some cases, baserates have been inferred from other methods, including WEPP model runs (e.g. Tysdal et al., 1999). The "base" condition is a graveled road segment 80 meters long with no vegetation low traffic and a 6% slope (so R is about 5m). The default baserate is 79 kg/yr/m vertical

drop along the road based on a field site in western Oregon (Luce and Black, 1999). Baserates for other calibrations can be found in Appendix A.

Surface Factors

The surface factors used in GRAIP_Lite are dependent on the surface type and the traffic level of the road being modelled. This allows GRAIP Lite to account for differences observed between roads with varying amounts of traffic and with different surface types. The surface type "Not a road" is used to model planned roads that do not yet exist as part of how GRAIP Lite models different treatment alternatives. These factors are shown in Table 1. The entries for the "Low" traffic level and varying surfacing are based on field data in western Oregon (Luce and Black, 1999). There have been a limited number of studies examining the effect of vehicle traffic on road surface erosion. These studies have typically measured erosion at the bottom of a road segment under different traffic levels with different levels of road maintenance that generally accompany increased road use. Kochenderfer and Helvey 1987, Bilby et al. 1989, Fahey and Coker 1989, Foltz 1999, and Sheridan et al 2006 found that a change from light vehicle use to more than 5 heavy truck passes a day resulted in an increase by a factor of 2-5 in the observed erosion. This traffic change is typical of a US Forest Service maintenance level 2 or 3 road that is being used as a haul route for a timber sale. Observations on erosion from closed, gated and unused roads are rare. In a study on the Lolo National Forest (Black et al. unpublished) we found that erosion on unused road with vegetation on the surface was less than 10% of that on similar roads with occasional light vehicle use.

Table 1: Surface Factors by Surface Type and Traffic Level.

Vegetation Factors

Vegetation growing in the flowpath on the road's surface acts to reduce sediment production by increasing surface roughness and by anchoring material. The vegetation factor is calculated as

V = 1 – 0.86 *x*

where *x* is the fraction of the road segments where the flowpath vegetation cover is greater than 25%, as observed in a calibration data set. This is calculated for each combination of surface type and maintenance level and represents the average vegetation on those resulting classes of roads.

Sediment Delivery

Data from GRAIP inventories shows that not all sediment eroded from the road surface is delivered to streams, and that the majority of the sediment that is delivered to the stream network is delivered by a small portion of the total road network. With the GRAIP model, sediment delivery is based on a direct field observation indicating if each individual drain point is hydrologically connected to the stream network. Since GRAIP_Lite does not use those direct observations, a different method of modelling sediment delivery is required.

Stream Connection Probability as Fractional Delivery

GRAIP_Lite uses a fractional delivery model instead of the yes/no model that GRAIP uses. Calibration data is used to define a set of curves describing the probability that a drainpoint would be observed to be stream connected based on the modeled flow distance to the modeled stream network and the length of the road segment it drains (Figure 3). We define the fractional sediment delivery to be equal to the probability of observed stream connection, and use a local polynomial regression (Loader, 2013) to estimate those probabilities conditioned on distance to stream and road segment length class. Given that a probability is mapped to fractional delivery, this approximation is only applicable when averaging across a number of potential delivery sites. It is an application of risk assessment, where the cost is the amount of sediment and the probability of delivery is the hazard.

Figure 3: A comparison of the different calibration curves used by two local calibrations.

Sediment Accumulation

The very fine sediments originating in road surface erosion move fairly readily through steeper headwater streams and it is reasonable to ignore Sediment accumulation within the stream network is calculated by the ArcHydro tools in the same way that GRAIP uses the TauDEM tools to route sediment through the stream network and calculate contributing areas and specific sediment loads.

Calibration

Calibration data is collected in the field using GPS and data entry at each site. At each sample drainpoint, the crew collects data on drain type, stream connection, road surface type, recent road maintenance, and flowpath vegetation and records the data in the database. Excepting road maintenance, GRAIP data can be used to derive a calibration data set if there is pre-existing GRAIP data available. The observations of stream connection provide the basis for local polynomial regression curves used to predict the probability of stream connection as a function of stream distance. Stream distance is the modelled flow distance from the observed drainpoint and the stream network modelled by the ArcHydro tools. The road surface type, maintenance level, and flowpath vegetation records are used to calculate the vegetation factor for road surface type and maintenance level combinations.

If you are using one of the included calibrations, you will see a message in green text advising you that the delivery probability, vegetation factor, and baserate tables are missing (Figure 4). The table values are already in the tables and so this is not an issue when using these calibrations.

Importing calibration zone... Missing calibration table DeliveryProbability_ Basalt_PayetteNF. Missing calibration table DeliveryProbability_ Granite_ BoiseNF. Missing custom calibration table VegFactor. The default table will be used. Make sure to upload your calibration zones. Missing custom calibration table BaseRate. The default table will be used. Make sure to upload your calibration zones.

Figure 4: Messages advising that calibration tables are missing.

Prerequisites and Installation

To install GRAIP_Lite, make sure that the computer has ArcGIS 10.3.1 or higher, with the Spatial Analyst and 3D Analyst extensions and the Advanced License, then install the most recent version of Arc Hydro Tools (10.3.172 or later). GRAIP_Lite is part of the Arc Hydro Tools package, available at [http://downloads.esri.com/archydro/archydro/setup/.](http://downloads.esri.com/archydro/archydro/setup/)

Further information on GRAIP_Lite can be found at https://www.fs.usda.gov[/GRAIP/GRAIP_Lite.ht](https://www.fs.usda.gov/GRAIP/GRAIP_Lite.html)ml.

ArcMap setup.

There are a couple things in the ArcMap settings that are necessary to let GRAIP_Lite run smoothly. These should be set before starting the GRAIP_Lite tools; once set, they should remain set, but they are good places to start when troubleshooting.

First, using the Geoprocessing menu, open Geoprocessing Options (Figure 5). Make sure that the boxes are checked for "Overwrite the outputs of geoprocessing operations" and "Add results of geoprocessing operations to the display" that Background Processing is not enabled. This should eliminate errors where GRAIP_Lite cannot overwrite certain temporary files (like the out_splitlineatpoint feature classes). Background processing is generally less stable than foreground processing.

Figure 5: ArcMap Geoprocessing settings. These should also be set in ArcCatalog.

Another useful setting, especially on slower computers, is found in the ArcMap Options window (Figure 6) under the Customize menu. If you uncheck the box by "Make newly added layers visible by default",

ArcMap will not try to draw each layer over and over as different layers are added to the map document. This can significantly speed up slower machines.

Figure 6: ArcMap Options settings.

ArcHydro and GRAIP_Lite make use of two extensions in ArcGIS: 3D Analyst and Spatial Analyst. Spatial Analyst is also very useful for setting up your GRAIP_Lite projects. You can verify that these are set up

and available by going to Customize -> Extensions and making sure the appropriate boxes are checked (Figure 7).

Figure 7: The Extensions window accessed in the Customize menu.

Input Data Requirements

GRAIP_Lite requires that the input data meet certain requirements. While in some cases GRAIP_Lite may still run if these requirements are not met, the model results may not be accurate. GRAIP_Lite requires a rectangular projection with meters as the linear unit. Since input data is often in different projections, or in unprojected geographic coordinate systems, all input data should be checked and projected into a suitable coordinate system. Universal Transverse Mercator (UTM) systems are ideal.

Roads Layer

The roads layer input into GRAIP_Lite also needs to meet specific requirements as to the attributes present in the data set. GRAIP_Lite was designed around the U.S. Forest Service's INFRA road layers, which contain information about road surfacing and maintenance that is used by the model to estimate parameters related to sediment production.

Both the attribute field names and the values contained in those fields needs to fit the specific form that GRAIP_Lite expects. The three required fields for any dataset are the route status (ROUTE_STAT or ROUTE_STATUS), surface type (SUFACE_TY or SURFACE_TYPE), and operational maintenance level (OPER_MAINT or OPER_MAINT_LEVEL) fields; short names are for shapefiles and long names are for geodatabase feature classes.

Where data are missing from these fields, GRAIP_Lite attempts to fill it in by making certain assumptions. If data are missing from the surface type field, the model will assume a native road surface; if data are missing from the maintenance level field, then the model assumes the road is a maintenance level 2 road, meaning that it has minimal design and is intended for slow, high clearance vehicles. Often, INFRA road layers are missing this information from non-Forest Service System roads, which may include private or county roads and state or federal highways. In most such cases, especially with county, state, or federal roads, the surface type may be known and it is just the maintenance level that is missing. Ideally, all missing data will be accurately filled in. Again, the design allows for the manger's discretion with respect to effort versus accuracy. In most cases, the default sets the condition to the highest risk so that analysts may reasonably claim that uncertainty was set to a worst case analysis.

We have seen GRAIP_Lite give an error when the data in Oper_Maint is not in the correct format; for example, if it is set to "D – Decommissioned" it will cause an error because D is not a numeric value and GRAIP_Lite does not know what to do with the value. Such values need to be corrected before the database will initialize by changing the value to one of the expected values; "D – Decommissioned" gets changed to "1 – BASIC CUSTODIAL CARE (CLOSED)" and "NA – NOT APPLICABLE" should be changed to match the road characteristics, if known, or deleted so there is no value for those records.

Digital Elevation Model

Calibration sets included with GRAIP Lite were developed using DEMs from the $1/3rd$ arc-second National Elevation Dataset (NED) with a nominal 30m resolution, accessible at [https://viewer.nationalmap.gov/basic/.](https://viewer.nationalmap.gov/basic/) In practice, these DEMs run between about 20m and 28m resolutions when projected into a UTM coordinate space; we usually round the cell size off to the nearest meter for simplicity. When using one of the installed calibrations, including the default calibration, a nominal 30m DEM has been assumed and should be used.

Because Arc Hydro uses a threshold approach to determining stream head locations and that contributing area threshold is determined by the number of contributing cells rather than the contributing area, using a higher resolution DEM without also changing the threshold cell count results in a denser modelled stream network (Figure 8), which in turn results in shorter modelled flow distances to streams and differences in road connectivity and sediment delivery. DEM resolution must match that of the calibration being used, or care must be used to set the *Number of Cells* threshold when running the *DEM Processing* tool to match the contributing area threshold, and therefore modelled drainage density, of the calibration. The contributing area threshold is described by

 $C L² = A$

Figure 8: Relationships between DEM resolution, contributing cell threshold, and resulting contributing area thresholds in terms of resulting modelled drainage density and stream connection probabilities.

It is also important to keep in mind that the streams in GRAIP_Lite are modeled streams at all steps beyond the collection of the field calibration data. This means that there will be discrepancies between these modeled streams and the actual streams on the ground when it comes to the location and extent of the stream networks. Local controls on steam head locations may not be apparent when modeling stream networks from a DEM when using a simple threshold cell-count approach.

Using Calibration Zones

Calibration zones are defined by a polygon shapefile or feature class with an extent equal to that of the DEM. These define areas where characteristics are relatively similar (e.g. climate and geology) and the

necessary model parameters are known. A given run of GRAIP-Lite can use one or more zones. Each defined zone requires a field called "Name" that contains the name of the calibration for that area. Included calibrations can be used by assigning the names of the desired calibrations (Table 2). Managers seeking to increase the accuracy and precision for their project area can complete surveys to create a custom calibration (effort-accuracy balance). Custom calibrations require that the appropriate parameter tables are included in the same directory as the calibration zone shapefile or feature class.

Table 2: Included GRAIP_Lite calibrations.

Known Drainpoints*-optional*

The known drainpoints layer is a point shapefile or feature class containing the locations of known, surveyed drainpoints along the modeled roads. GRAIP_Lite makes use of points within 10m of the roads layer by snapping those points to the road layer and using them as split points when breaking up the road layer into road segments. This allows the user to force GRAIP_Lite to create split points at engineered drainage features in addition to relying on topographic and maintenance level based split point predictions.

Input Data from Other Sources

Road layers from other sources may be used, but some pre-processing will likely be necessary to make such layers compatible. It will likely be necessary to add and populate the required fields, which are the route status (ROUTE_STAT or ROUTE_STATUS), surface type (SUFACE_TY or SURFACE_TYPE), and operational maintenance level (OPER_MAINT or OPER_MAINT_LEVEL) fields; you can use either long or short names depending on the data type. All of these fields are text fields.

To populate the fields with appropriate data, you will need to know something about your roads and know what values to input (Table 3) so GRAIP_Lite understands what your roads are like. The easiest field is the route status field; this field is used primarily to guess the maintenance level if the maintenance level field is blank but should still be filled in. The surface type field should also be fairly easy to assign. Maintenance levels are probably the trickier values to assign. For more information on Forest Service road maintenance levels see Apadoca et. al. 2012.

Table 3: Input values for GRAIP_Lite.

Applications and Uses

GRAIP_Lite has several applications and uses for management purposes. These applications extend from forest-wide watershed condition assessments to project scale NEPA work. Different types of analysis have some different assumptions and limitations.

Model validations between GRAIP_Lite and GRAIP reveal typical patterns of potential errors created by the differences between the two models, largely due to GRAIP_Lite using a probabilistic fractional delivery model and GRAIP using binary (yes or no) observations of delivery. When there is a small amount of road being modeled, there is a greater chance that GRAIP_Lite sediment predictions for individual stream reaches may be significantly different from those predicted by GRAIP (Figure 9A). At the same time, the predictions showing the greatest error are also likely to be relatively smaller masses of sediment (Figure 9B). In both cases, the mean errors, and even the standard deviations, are small even with small amounts of road and smaller amounts of modeled sediment. Some of the errors are also due to GRAIP's ability to describe road segments and drainpoints that do not fit average observed conditions, either by being in better shape than average or by being in worse shape than average.

Another important factor is the accuracy of the input information, especially the data in the road layer. GRAIP has an advantage in this case because of the intensive field inventory; all of GRAIP's input data is effectively field verified and can vary at the scale of individual model road segments. GRAIP_Lite is dependent on the accuracy of the road layer input; if the data in the road layer is inaccurate or out of date, the model results will be as well, and where data are missing, for example where surface type or maintenance level attributes are not complete, GRAIP_Lite must assume parameters which may not reflect reality. Any model is, at best, only as good as the data driving it. A range of options are available in GRAIP_Lite for accepting lower quality input data while still using some understanding of road systems to give a reasonable estimate.

GRAIP_Lite is an ideal tool for use in watershed condition assessments that may be done across one or more forests at a time. In this type of analysis, GRAIP Lite is used to assess road surface-related sedimentation within areas ranging from $6th$ -code HUCs to about half a $6th$ -code HUC across the total landscape. At these scales, GRAIP_Lite yields similar sediment accumulation and delivery values to GRAIP, and absolute values may be used with confidence, insofar as the road data in INFRA are accurate. The goal of this kind of assessment is to prioritize $6th$ -code or half $6th$ -code sized HUCs in regards to roadrelated impact so that future work can be focused in areas that address the greater needs where higher road-related impacts exist. This allows more effective use of project dollars. This application is fast and easy enough to implement that it can replace road density as a sort of basic index. GRAIP Lite essentially adds that steeper roads, roads closer to streams, and roads with more traffic and less surface preparation pose greater risk from the perspective of road surface erosion. While such concepts are generally well understood, their implementation in GIS has previously been burdensome.

On the other end of the scale spectrum, GRAIP_Lite is used to assess how different proposed road treatments may affect road-related sediment risks in support of a NEPA proposal. This type of analysis usually covers a smaller area, being run on a project scale, and is assessing a smaller set of roads. As a result, this becomes an analysis of relative risk rather than of absolute values. Keeping this in mind, GRAIP_Lite uses the Alternatives module to assess how road-related sediment impacts are likely to change depending on the specified current conditions and treatments applied, modeling the treated road network at an initial condition before any work has been done, at a disturbed condition when the work is recently completed, and at a recovered condition when road surface vegetation and traffic levels have recovered to an equilibrium condition after the work has been completed.

Figure 9: Typical error distributions when validating GRAIP_Lite against the GRAIP model, based on accumulated road length and accumulated sediment.

Model Run Tutorials

Several tutorial data sets have been prepared to illustrate potential applications like those outlined above. A "basic" run may be used to assess the relative risk from road surface erosion across subwatersheds. It is ideal for prioritization of watershed restoration efforts and is a substantial improvement over road density. A "calibration zone" example is included to show how a more heterogeneous area, say with variation in geology, can be modeled. It has similar purposes to the basic run, but accounts for environmental variation in the calculus of determining which basins have greater risk. Finally, an example of how to apply GRAIP_Lite for alternative comparison, as one would do for project planning under NEPA, is included. Here an input system describing most treatments has been

set up to make treatment specification easier, and execution of the multiple runs and preparation of output graphics has been automated to speed analyses. The three tutorial examples have been set up as a progression to let users become familiar with easier applications before taking on more advanced tasks.

Basic

This tutorial shows how to use the tools in GRAIP_Lite's Basic Run toolbox. The data set uses INFRA roads sourced from a geodatabase, in this case for the North Fork Boise River drainage on the Boise National Forest, current as of 8 January 2016. The Basic Run is intended to provide the user with a quick view of higher risk areas within the road network and uses the default model calibration. The tool provides a simple analysis and includes a reporting tool to create basic maps for use in reports or presentations.

The first step in any GRAIP_Lite model run, including a basic run, is to save the map document in the project folder. This sets the default locations for files created and accessed by GRAIP_Lite and names the geodatabase in which GRAIP_Lite stores the various feature classes created. With ArcMap open with a blank document, go to *File -> Save As*, navigate to your project folder, and save the document, in this case as *BasicRun.mxd* (Figure 10).

Figure 10: Saving the map document.

Once the map document is saved, it is time to open the Basic Run tool. Open the ArcToolbox window if it is not yet open. The tool is located in ArcToolbox->GRAIP Lite->Basic Run-> 01. Basic Run - Road and DEM only. Double-click to open the tool. Notice that the Target Geodatabase Directory and the Target Geodatabase Name fields are automatically filled in based on the location and name used to save the map document (Figure 11).

Figure 11: Basic Run tool opened.

With the tool open, you will need to populate the Input Road and Input DEM fields. The browse buttons allow you to select the appropriate inputs. For this tutorial, the Input Road is the INFRA Roads feature class in the BasicInputs.gdb geodatabase (Figure 12) and the Input DEM is nfbc in the Layers folder (Figure 13).

Figure 12: Selecting the Input Road feature class.

Figure 13: Selecting the Input DEM.

With the road and DEM inputs specified, double check that the box next to QC Road is checked (Figure 14); this ensures that the model will do some basic checks on the input road layer, including looking for loops and geometric duplicates. Click OK and the tool starts running (Figure 15).

Figure 14: Basic Run tool ready to run.

Figure 15: Basic Run tool running.

Once the tool has completed (Figure 16), you can close the tool dialog and view the results in the map window (Figure 17). The model has completed its run at this point and the data is now available for use. One of the primary model outputs is Specific Sediment Delivery, which is shown in the DrainageLine feature class. This is the road surface-related sediment load per unit contributing area, and provides a good measure of road-related sediment impact on aquatic habitats. In Figure 17, we can see streams highlighted in red or orange where the local impact is high; yellows and greens are used for moderate to low impacts, and blue is used where no sediment is delivered to streams.

Figure 16: Basic Run tool has completed the model run.

Figure 17: Results from the Basic Run tool Specific Sediment Delivery along Drainage Lines.

The second tool in the Basic Run toolbox is the Basic Report tool. This tool automatically generates a series of maps in both .jpg and .pdf formats; the .jpg files are intended to be used directly in documents and presentations while the .pdf documents provide an easy way of communicating those maps to the public or to various other partners.

Open the tool and note that the Input GRAIP Lite Workspace is already populated (Figure 18). More advanced users may be able to use custom templates, but the defaults are preferred here so there is no need to populate the Input Template Directory field. Hit OK to run the tool.

Figure 18: The Basic Report tool ready to run.

As the tool runs, it lists off each map it is exporting and where it is stored. The tool stores these reports in a Reports subfolder within the project folder, and within this creates a unique folder each time the tool is run. Close the tool when it has completed (Figure 19). Example maps from the Basic Report tool will be included in Appendix D.

Figure 19: Basic Report tool complete.

Calibrated

The tools in the Processing toolbox can be used to customize the model run and take advantage of additional capabilities. The most important of these is the ability to use different model calibrations to more closely describe the sediment production and delivery characteristics of the project area. As with the Default Run, the first step is saving the map document (Figure 20).

Figure 20: Saving the map document.

The next step is to run the *01. Initialize GRAIP Lite Database* tool (Figure 21). This tool populates the database used to store the inputs for and outputs from the model. The Target Geodatabase Directory and the Target Geodatabase Name fields are automatically filled out based on how you saved the map document; the directory is where you stored it and the name matches the name of the map document.

The Input Road and the Input DEM are required fields and are the same as those for the Basic Run; for this example we want to use PNF_INFRA_Clip.shp for the Input Road and wr as the Input DEM. Using just these two inputs is the same as doing the Basic Run, although it does allow more fine-tuning in some of the steps, namely in the DEM processing steps to come. The two optional fields are the Input Observed DrainPoint field and the Input Calibration Zone field. The Input Observed DrainPoint field, which is not used in this example, is used to provide the model with the locations of known drainage feature from a culvert or other inventory; the model then uses these features as breaks when creating road segments for the model. The Input Calibration Zone field is used to delineate areas that should be modeled using different model calibrations; for this example we are using CalibrationZones.shp as the input. Leave the box for QC Road checked.

Figure 21: The Initialize GRAIP Lite Database tool.

Next is the *02. DEM Processing* tool (Figure 22). This is the tool that generates the stream network and the grids necessary for routing sediment within the model. The Number of cells field is used to calibrate the number of accumulated gridcells, or contributing area, necessary for generation of a stream head; however, this needs to match what was used for creating the individual calibrations. All of the included calibrations, which we are using in this example, were created using the default threshold of 100. All of the fields should be automatically populated.

Figure 22: DEM Processing tool.

The next step is the *03. Calculate Stream Distance* tool (Figure 23). This tool calculates the flow distance to the nearest downhill stream and stores it as a raster; the data gets used later to help calculate sediment delivery. All of the fields should be pre-populated and ready to run.

Figure 23: Calculate Distance from Stream tool.

Next is the 04. Create Road Segments tool (Figure 24). This tool takes the input roads features and splits them into GRAIP_Lite road segments using catchment boundaries, stream crossings, calibration zones, known drainpoints, and pre-determined maximum distances. All of the fields should be pre-populated.

Figure 24: Create Road Segments tool.

The next step is to run the 05. Calculate Road Segment Sediment Production tool (Figure 25). This tool calculates the sediment production expected from each road segment based on the difference in endpoint elevations, the surface type, the expected vegetation, and the expected traffic. Different calibrations provide different baserates and vegetation factors to customize the calculations. Again, all fields should be pre-populated.

Figure 25: Calculate Road Segment Sediment Production tool.

The 06. Create Road Segment Drain Points tool (Figure 26) is used to select the end of each road segment which is at the lowest elevation and create drain point features at those location. The tool also appends the flow distance to the nearest stream from the DisttoStr raster.

Figure 26: Create Road Segment Drain Points tool.

The 07. Calculate Sediment Delivery tool (Figure 27) uses the flow distance at each drainpoint and the delivery probability table to calculate the amount of sediment expected to be delivered to the stream network at each drainpoint. This step usually has the longest run time. The Delivery Probability table has values describing different curves for each calibration zone allowing the sediment delivery characteristics of an area to be described as part of the calibration.

Figure 27: Calculate Sediment Delivery tool.

The 08. Route Sediment to Streams tool (Figure 28) is next on the list. This tool uses some of the rasters created by the DEM Processing tool to route sediment downhill from the drainpoints to the stream network and then add up the total amount of expected road sediment in the stream network. It also routes and sums the connected road length to create a tally of how much road has affected the stream network. It also normalizes both the accumulated sediment data and the accumulated connected road length data by the contributing area to produce rasters describing specific sediment (accumulated sediment divided by contributing area) and connected road density.

Figure 28: Route Sediment to Streams tool.

The 09. Report Parameters of Drainage Line tool (Figure 29) takes the data recorded in the rasters created in the previous step and appends that data to the features in the Drainage Line feature class, which stores the stream network data. This makes it much easier to present the data in map form, and allows the data to more easily be summarized.

Figure 29: Report Parameters on Drainage Line tool.

At this point, the Basic Report can be run, which will provide the same set of maps as those produced for the Basic Run.

Alternatives

GRAIP_Lite is also designed as a tool for analyzing different potential treatment options. In order to do this, it provides a way of analyzing different alternative at three time steps. Each alternative is modeled at the initial condition time step, a disturbed time step, and a recovered time step. GRAIP_Lite has a dialog box that is used to create each alternative and set the treatments applied with that alternative; multiple alternatives can be created and run as part of a GRAIP_Lite model. This tutorial is intended to highlight the main ways in which treatments are specified for the GRAIP_Lite model.

There are also a few additional data considerations when modeling alternatives that are not as big a deal when just modeling current conditions. Since road treatments may only be applied to portions of certain roads, those portions need to be separate from the untreated portions of the road network. Since in many cases these treatment portions are defined based on intersections with other roads, one easy way to deal with this is to planarize the road network prior to beginning the model run; this method is demonstrated in this tutorial. If a treatment boundary does not correspond to an intersection, the road section must be split at the appropriate point in order to correctly model the differing sections.

As always, the first step is to save the map document in the project folder, in this case as EFWR.mxd (Figure 30). It is also a good idea to save the map document frequently.

Figure 30: Saving the map document.

The next step is to add the INFRA.shp file to the map document (Figure 31). This file is the shapefile containing the road information for the project. Click on the Add Data button to access the dialogue box and add the shapefile.

Figure 31: Add INFRA shapefile.

Next, make sure the Editor toolbar is visible, click on Editor to open the menu, and then click on Start Editing. This will activate all of the tools in the Editor and Advanced Editing toolbars (Figure 32).

Figure 32: Start Editing.

Next, use the Selection tool to select all roads in the INFRA shapefile (Figure 33). The easiest way to do this is to click and drag a box around all of the roads shown in the map window.

Figure 33: Select all road features.

Once all of the roads are selected, click on the Planarize Lines tool in the Advanced Editing toolbar (Figure 34). Specify a Cluster Tolerance of 0.001 meters, or 1 mm, and click OK. This tool splits all of the lines at each intersection point, resulting in separate features on either side of the intersection. All attributes are preserved during the process.

Figure 34: Planarize Lines tool.

Once the tool has finished, click on Save Edits and then Stop Editing, both located in the Editor menu (Figure 35). At this point, all road sections should be deselected. If there are places where the road needs to be split between intersections, use the Split tool, located in the main Editor toolbar, to do so while in editing mode, making sure to save your edits; in this tutorial this is not necessary.

Figure 35: Save edits and stop editing.

Now that the data is in the format we need so that we can properly designate which portions of the road network will be treated, it is time to start the GRAIP_Lite model run. Before we can model the alternatives, we need to give GRAIP Lite the underlying structure. This can be done with the Basic Run tool if there is no calibration or known drainpoints to be included, but if either or both of these are available it is best to do a full run using the Processing tools. The first step is the 01. Initialize GRAIP Lite Database tool (Figure 36). This tool tells GRAIP Lite where our data is and creates the geodatabase to keep track of the results.

Figure 36: Initialize GRAIP Lite Database tool.

The next tool, *02. DEM Processing*, generates the stream network and the raster datasets needed for routing sediment in the GRAIP_Lite model (Figure 37).

Figure 37: DEM Processing tool.

Next is the 03. Calculate Distance from Stream tool (Figure 38), which calculates the distance along the flowpath to the nearest downhill stream.

Figure 38: Calculate Distance from Stream tool.

The next tool is the *04. Create Road Segments* tool (Figure 39). This tool takes the input road layer and splits it up into GRAIP_Lite road segments based on intersections with calibration zone and catchment boundaries, streams (drainage lines), observed drainpoints, and a maximum road segment length determined by the road segment maintenance level.

Figure 39: Create Road Segments tool.

Road-surface sediment production is then calculated using the 05. Calculate Road Segment Sediment Production tool (Figure 40).

Figure 40: Calculate Road Segment Sediment Production tool.

The next tool, the 06. Create Road Segment Drain Points tool, creates drain points at the low end of each road segment and appends the flow distance to the stream (Figure 41).

Figure 41: Create Road Segment Drain Points tool.

Next is the 07. Calculate Sediment Delivery tool (Figure 42) that calculates how much of the produced sediment from each road segment will be delivered to the stream network.

Figure 42: Calculate Sediment Delivery tool.

The next step is to route the delivered sediment into the stream network using the 08. Route Sediment to Streams tool (Figure 43).

Figure 43: Route Sediment to Streams tool.

Finally, run the 09. Report Parameters on Drainage Line tool to add the accumulated sediment metrics to the drainage line feature class (Figure 44). This completes the initial part of the model run and establishes the model parameters that will be used to model the alternatives.

Figure 44: Report Parameters on Drainage Line tool.

Before the treatments can be modeled, GRAIP_Lite needs to know what the treatments are and what road segments are being treated. This is done using the GRAIP_Lite window, which is opened from the GRAIP_Lite toolbar by clicking on GRAIP Lite (Figure 45). The window can be docked or moved to a convenient location on your screen.

Figure 45: Open GRAIP Lite window.

Another common, and annoying, error we have been seeing involves the ApUtilities setting the target locations incorrectly, usually to one of the previously opened map documents. This results in errors when starting the GRAIP Lite window in preparation for defining treatments to be modeled as alternatives. When the GRAIP Lite window is opened from the GRAIP Lite toolbar, three tables are supposed to be copied into the GRAIP_Lite project database. The red box in the Figure 46 shows the three tables; the picture shows them in the correct location, as noted by the path highlighted in blue (pointing to the project geodatabase).

Figure 46: The tables created when opening the GRAIP_Lite window should appear in this location in your project geodatabase.

If the tables show up where they are supposed to, then all is good. If the tables show up in any other path, close ArcMap **WITHOUT** saving the document. Re-open ArcMap, then re-open the map document, and try again. In our testing, it should work at this point, but double check that the tables are being copied to the correct location before saving the map document.

One way we have found to avoid this is to save, close, and then re-open the map document after running the *Initialize GRAIP Lite Geodatabase* tool. This seems to prevent the issue, which appears to be related to ArcMap not letting go of settings for a previous map document and work space.

Once the GRAIP_Lite window is open, and the tables are in the correct location, click on New to open the options for a new alternative (Figure 47). If you have already created an alternative, you would also have the option to click on Open to access an existing alternative.

GRAIP Lite			\square \times
Alternative Title	Source	Target	
	Road		
		New	
		Caper	
Alternative Treatments Map Selection	Selection Type Create Naw Selection	Zoom to Selection Switch Selection Show Selected \rightarrow	
Treatment VC V1 ID.	T ₀ V ₂ T1	T ₂ LD L2 SO S1 S ₂ MLC ML1 ML2 L1	
		<u>e ge</u>	

Figure 47: GRAIP Lite window; click on New to start.

After clicking *New*, the window will provide you options for setting up the alternative (Figure 48). You can use the *Title* field to name the alternative, as well as select the *Source* road layer for the alternative. For this tutorial, the defaults will be used. The road layer in this case is the *Road* feature class from the initial model run; this is the copy of the road layer created when the *01. Initialize GRAIP Lite Database* tool was run. There is also a space to enter a description of the alternative. Click *Create* and wait while the tool creates the feature classes for the alternative.

Note that you can choose a different *Source* road layer; this is most commonly used if creating multiple alternatives where there are only a few differences between them. In such a case, one alternative can be set up, and then that layer (*Alternative1_Road*) may be used as the *Source* for the other alternative with only minor changes being made to subsequent alternatives.

Alternative							
Title	Source	Target					
Alternative 1	Road	- Alternative 1_Road					
Enter description for new alternative.		Create					
		Cancel					
Alternative Treatments							
Map Selection	Selection Type Create New Selection	\star	Zoom to Selection	Switch Selection	Show Selected		
Treatment VO V1 ID	T1 V ₂ T ₀	T2 LO	L1 L2	SO S1	S ₂	MLD ML1 ML2	

Figure 48: Create new alternative.

Once the tool has created the alternative, you will be able to see a table where you will be able to set treatment values (Figure 49). Click on Open to begin editing the treatments for the alternative. Notice that when the new alternative was created, a new map is added to ArcMap's Table of Contents. Save the map document.

Figure 49: Newly created alternative ready to open for editing.

One way to plan treatments is using a shapefile or feature class dataset to keep track of those planned treatments. For the purposes of the tutorial, the Treatments.shp shapefile contains just the planned treatments. It was created from a planarized version of the INFRA.shp shapefile used for the tutorial. Add the Treatments.shp shapefile using the Add Data dialogue (Figure 50). A separate shapefile or feature class is not necessary; if the treatments are already recorded in columns in the road layer, that information will carry over into GRAIP_Lite which means that features can be directly selected in the Alternative1_Road layer using the Select by Attributes tool. If the treatment descriptions are already in the Alternative1_Road layer, the Select by Location tool is not necessary.

Figure 50: Add Treatments layer.

Once the treatments layer has been added, double click on it to open the Layer Properties and click on the Symbology tab (Figure 51). Select the Unique values option under the Categories menu, then select Treatment in the Value Field menu. Finally, click Add All Values. The choice for color ramp isn't important; the idea is just to be able to distinguish each treatment category. Click OK when satisfied with the symbology and turn the layer on if it is not already on.

Hatches	Joins & Relates	Time		HTML Popup
General Source	Symbology Selection Display	Fields	Definition Query	Labels Routes
Show: Features Categories Unique values Unique values, many Match to symbols in a Quantities Charts Multiple Attributes 租	Draw categories using unique values of one field. Value Field Treatment Value Symbol call other values> <heading> New Road Permanent Re-Open Permanent ReConstruct ML2 ReOpen Temporary Use Decom ML2</heading>	Color Ramp Label call other values Treatment Decommission Full Recontor Decommission Full Recontor 274 New Road Permanent Re-Open Permanent ReConstruct ML2 ReOpen Temporary Lise Decom ML2	Import Count п 334 8 15 30 $\overline{6}$	
	Add All Values Add Values.,	Remove All Flamove	Advanced	

Figure 51: Treatments symbology.

It is now time to start selecting road segments and setting treatment values. In order to set the treatments, the appropriate road segments must be selected from the *Alternative1_Road* layer. There are a couple ways to do this, depending on the way the data is set up and how many road segments will receive the same treatment.

Since the *Treatments* layer in this example is derived from the *INFRA* data set used for the model, and both layers were planarized, individual road sections are geometrically identical. This means that we can select features in the *Treatments* layer using the *Select By Attributes* tool and then use the *Select By Location* tool to select the corresponding features in the *Alternative1_Road* layer. If your project has the treatments specified in your road layer, just use the *Select By Attributes* tool to directly select the features in the *Alternative1_Road* layer.

Open the *Select By Attributes* tool by clicking on *Selection* and then *Select By Attributes* in the main ArcMap menu. Make sure *Treatments* is the target *Layer*, and in the list of attribute fields, scroll to the bottom to find *"Treatment"* (Figure 52). Double click on *"Treatment"* to add it to the logic statement in the lower part of the dialogue box, then click on *Get Unique Values* to see a list of the possible values for the field *"Treatment".* Now click on the *=* button and then double click on *'Decommission Full*

Recontour'. You should now have a statement in the lower part of the dialogue box: "Treatment" = 'Decommission Full Recontour'. Click Apply to select all features in the Treatments layer where the treatment is specified to be decommissioning by fully recontouring the road.

Figure 52: Select by Attributes tool.

Open the Select By Location tool by clicking on Selection and then Select By Location in the main ArcMap menu. This tool will be used to select the corresponding features in the Alternative1_Road layer. The Select By Location tool (Figure 53) is used to select features from the Target layer or layers based on their location relative to features in the Source layer. In the Target layer(s): list, make sure Alternative1_Road is selected. Make sure Treatments is the Source layer and click the box to Use selected features. There is a list of different options for Spatial selection method for target layer feature(s):; use the option are identical to the source layer feature. This selects the features in the Alternative1_Road layer that are geometrically identical to the features selected from the Treatments layer during the previous step.

Figure 53: Select by Location tool.

The selected features will appear highlighted in both the map window and the GRAIP Lite dialogue box (Figure 54). You can check that the correct features were selected by turning on and off the *Treatments* layer and comparing the selected features on the map.

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		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native	$\overline{2}$	$\overline{2}$	$\overline{2}$	
$\overline{2}$		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native	$\overline{2}$	2	$\overline{2}$	
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		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native	$\overline{2}$	2	2	
		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native	2	$\overline{2}$	\overline{a}	
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$\overline{9}$		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native	$\overline{2}$	$\overline{2}$	$\overline{2}$	
Þ 10 11		Default Default	Default Default	Default Default	Low Low	Low Low	Low Low	50 50	50 50	50 50	Native Native	Native Native	Native Native	ø $\overline{2}$	2 ◆	$\overline{2}$ \overline{a}	

Figure 54: Selected features.

In the GRAIP Lite window, click on the box for Show Selected to show only the selected road sections (Figure 55).

	GRAIP Lite																	O x
Title	Altemative				Source		Target											
	Alternative 1			۰	Road			Alternative 1 Road										
		Enter description for new alternative					New Open											
		Alternative Treatments																
		Map Selection	Selection Type		Create New Selection		\blacktriangledown	Zoom to Selection		Switch Selection		Show Selected						
	ID	Treatment	V _D	V1	V ₂	T ₀	T1	T2	LD.	L1	$\sqrt{2}$	S ₀	S1	S ₂	ML ₀	ML1	ML2	$\hat{ }$
\mathbf{r}	5		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native	$\mathbf{2}$	2	\mathbf{Z}	
	в		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native	$\overline{2}$	$\overline{2}$	$\overline{2}$	
			Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native	$ 2\rangle$	12.	$\overline{2}$	
	ß,		Default	Detault	Default	Low	Low	Low	50	50	50	Native	Native	Native	$\overline{2}$	2.	$\overline{2}$	
	10		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native	$\overline{2}$	$\overline{2}$	$\overline{2}$	
	11		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native	$\overline{2}$	$\overline{2}$	$\overline{2}$	
	12 ²		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native	$ 2\rangle$	$\overline{2}$	$\overline{2}$	
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2-3-3-3-3-3-3-3-3	16 ₁		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native	$\overline{2}$	$\overline{2}$	$\overline{2}$	
	17		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native	$\overline{2}$	$\overline{2}$	$\overline{2}$	
	$\overline{18}$		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native	$\overline{2}$	12	$\overline{2}$	
			Alternative1, Road rows loaded, dt=1.68s															d.

Figure 55: Showing only selected features.

Click on the top row of the table in the Treatment column. It will probably take a couple clicks before the treatment menu appears, but since the tool is doing things each time you click, wait for it to be done between clicks. You will see a drop-down menu when it is ready (Figure 56). If you click too fast, it will end up selecting only the top row as a subselection to apply the chosen treatment to; if you let the program catch up between clicks, you will be able to apply the treatment to all selected features at once.

	Alternative																	
Title					Source		Target											
	Alternative 1		\bullet		Road	×	Atemative 1_Road											
		Enter description for new alternative.					New Open											
		Attemative Treatments																
	Map Selection		Selection Type		Create New Selection		\blacktriangledown	Zoom to Selection		Switch Selection		V Show Selected						
	ID	Treatment	V ₀	V1	V ₂	T ₀	T1	T ₂	L ₀	L1	L2	S ₀	S1	S ₂	MLO	ML ₁	ML2	$\overline{}$
	5		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native	2	$\overline{2}$	$\mathbf{2}$	
	в		Default	Default	Default	Low	Low	Low	-50	50	50	Native	Native	Native	$\overline{2}$	2	$\overline{2}$	
			Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native	$\overline{2}$	$\overline{2}$	$\overline{2}$	
	县		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native	Ž	12.	$\overline{2}$	
	10		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native	$\mathbf{2}$	$\overline{2}$	$\sqrt{2}$	
	11		Default	Default	Befault	Low	Low	Low	-50	50	50	Native	Native	Native	$ 2\rangle$	$\sqrt{2}$	$\overline{2}$	
	12		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native	$\overline{2}$	$\overline{2}$	$\mathbf{2}$	
	13		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native	$\overline{2}$	$\overline{2}$	$\overline{2}$ -	
	16		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native	2	$\overline{\mathbf{2}}$	$\overline{2}$	
	17		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native	$\overline{2}$	$\sqrt{2}$	$\overline{2}$	
	B		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native	$\mathbf{2}$	$\overline{2}$	$\mathbf{2}$	

Figure 56: Treatment column showing dropdown menu ready to be accessed.

When the menu opens, select "Decommission - Full Recontour" and wait while GRAIP Lite sets the treatment values in the table. When it is done, it should look like Figure 57. Notice that the many of the values in the table have now changed to reflect conditions at the disturbed and recovered conditions based on the treatment specified.

	GRAIP Lite													O x
Alternative Title		Source		Target										
Alternative 1	\cdot	Road	i.	Alternative 1 Road										
	Enter description for new alternative.			New Open										
	Alternative Treatments													
	Map Selection Selection Type	Create New Selection		\bullet	Zoom to Selection		Switch Selection		V Show Selected					
ID	Treatment	V ₀	V1	V ₂	T ₀	T1	T ₂	LD	L1	\mathbf{L}	S ₀	S1	S ₂	MLD
$\overline{5}$ ⊪	Decommission - Full Recontour	Default \blacktriangledown	Bare	Covered	Low	None:	None	150	25	25	Native	Native	Native	
۱e	Decommission - Full Recontour	Default	Bare	Covered	Low	None	None	50	25	25	Native	Native	Native	$\overline{2}$
	Decommission - Full Recontour	Default	Bare	Covered	Low	None	None	50	25	25	Native	Native	Native	\overline{c}
	Decommission - Full Recontour	Default	Bare	Covered	Low	None	None	50	25	25	Native	Native.	Native	$\overline{2}$
10	Decommission - Full Recontour	Default	Bare	Covered	Low	None:	None	50	25	25	Native	Native	Native	$\overline{2}$
	Decommission - Full Recontour	Default	Bare	Covered	Low	None	None	50	25	25	Native	Native	Native	$\overline{2}$
	Decommission - Full Recontour	Default	Bare	Covered	Low	None	None	50	25	25	Native	Native	Native	$\overline{}$
12 ₂	Decommission - Full Recontour	Default	Bare	Covered	Low	None	None	50	25	$25 -$	Native	Native	Native	
13 ²			Bare	Covered	Low	None.	None	50	25	25	Native	Native	Native	2
16	Decommission - Full Recontour	Default												
17	Decommission - Full Recontour	Default	Bare	Covered	Low	None	None	50	25	25	Native	Native	Native	

Figure 57: Treatment selected and applied to features.

When the tool has finished, uncheck the box by *Show Selected* and then *Clear Selected Features* (in the *Tools* toolbar) before making a new selection.

Repeat this process to select and set treatments for those road sections with the specified treatments of "Reconstruct ML2", "ReOpen Temporary", and "Use Decom ML2" before moving on to the next part of the tutorial.

Another tool used to select roads is the *Select Features* tool (also in the *Tools* toolbar), which allows you to select visible features by clicking on them. This tool is more useful when there are only a few road sections receiving a given treatment. Use the *Zoom In* tool to zoom in on the area of the roads listed for "Re-Open Permanent" and "New Road Permanent" so that they roughly fill the map window (Figure 58); this tool is also in the *Tools* toolbar. This will make it easier to select the road sections that will be reopened.

Figure 58: Map window zoomed in to show "Re-Open Permanent" and "New Road Permanent" features.

Making sure both the *Treatments* and *Alternative1_Road* layers are visible, use the *Select Features* tool to select the features marked "Re-Open Permanent" (Figure 59); the features will be selected in both layers.

Figure 59: Select Features tool used to select the "Re-Open Permanent" features.

In the GRAIP Lite window, click on Show Selected so that only the eight selected features are displayed (Figure 60).

Title		Source	Target										
Alternative 1	$\overline{}$	Road	÷	Atemative 1_Road									
	Enter description for new alternative.		New Open										
	Alternative Treatments												
Map Selection	Selection Type	Create New Selection	$\overline{}$	Zoom to Selection		Switch Selection		V Show Selected					
ID	Treatment	V ₀	V1	V ₂	T ₀	T1	T ₂	L ₀	L1	L2	S ₀	S1	S ₂
310		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native
419		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native
420		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native
429		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native
445		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native
484		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native
498		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native
519		Default	Default	Default	Low	Low	Low	50	50	50	Native	Native	Native

Figure 60: Showing only selected features.

Next select the treatment to be applied to these road segments (Figure 61).

Atemative Title		Source			Target										
Atemative 1	$\overline{}$	Road			Alternative 1 Road										
	Enter description for new alternative.				New Open										
	Alternative Treatments														
	Map Selection Selection Type		Create New Selection		\checkmark	Zoom to Selection		Switch Selection		Show Selected					
ID	Treatment	V ₀	V ₁	V ₂	T ₀	T1	T ₂	L ₀	L1	$\sqrt{2}$	S ₀	S1	S ₂	MLD	ML ₁
310	Re-Open - Permanent $\overline{}$	Default	Bare	Default	None	High	Low	50	50	50	Native	Crushed Rock	Crushed Rock		\bar{z}
419	Re-Open - Pernanent	Default	Bare	Default	None:	High	Low	50	50	50	Native	Crushed Rock	Crushed Rock	2	$\overline{2}$
420	Re-Open - Permanent	Default	Bare	Default	None	High	Low	50	50	50	Native	Crushed Rock	Crushed Rock	12	\overline{z}
429	Re-Open - Permanent	Default	Bare	Default	None	High	Low	50	50	50	Native	Crushed Rock	Crushed Rock	$\overline{2}$	$\overline{2}$
445	Re-Open - Permanent	Default	Bare	Default	None	High	Low	50	50	50	Native	Crushed Rock	Crushed Rock		$\overline{2}$
484	Re-Open - Permanent	Default	Bare	Default	None	High	Low	50	50	50	Native	Crushed Rock	Crushed Rock		
498	Re-Open - Permanent	Default	Bare	Default	None	High.	Low	50	50	50	Native	Crushed Rock	Crushed Rock		2
519	Re-Open - Permanent	Default	Bare	Default	None	High	Low	50	50	50	Native	Crushed Rock	Crushed Rock		$\frac{2}{2}$

Figure 61: Treatment selected and applied.

Once the treatments are set, uncheck the box by *Show Selected* and then *Clear Selected Features*.

There is one road section listed in the *Treatments* layer that does not yet exist in the *Alternatives1_Road* layer. This is the new road with the specified treatment "New Road Permanent". Use the *Create New Road* tool from the GRAIP Lite toolbar to draw in the new road following the one shown in the *Treatments* layer (Figure 62).

Figure 62: The GRAIP Lite Create New Road tool.

The tool appears in the map as a crosshair, and places vertices for the new road line at the center of the crosshair with each click. Start at one end and double click to end the feature when you get to the other end (Figure 63). Notice that the newly created feature is selected in the GRAIP Lite window.

Figure 63: Draw new road following feature in Treatments layer.

Click on Show Selected (Figure 64). Notice that the new road has all default values including a native surface type and maintenance level 2. These are the defaults for any time the surface type and maintenance level are not defined.

Figure 64: Show Selected.

Set the *Treatment* to "New Road – Permanent" and notice the change in values, specifically for *S0*, which is set to "Not a road" (Figure 65). This means the road will not produce sediment during the initial condition run because it does not yet exist.

Alternative Title	Source	Target											\square x
Alternative 1 \bullet	Road.	-	Alternative 1 Road										
Enter description for new alternative.			New Open										
Alternative Treatments Map Selection Selection Type	Create New Selection	\cdot	Zoom to Selection		Switch Selection		V Show Selected						
ID Treatment	V1 VO.	V ₂	T1 T ₀	T ₂	LD	L1	L ²	SO	S ₁	S ₂	ML0	ML1	ML2
New Road - Permanent v	Bare Covered	Default	High None	Low	50	100	100	Not a road	Crushed Rock	Crushed Rock	$\overline{2}$	$\mathbf{3}$	B.

Figure 65: Assign treatment; note that S0, the surface factor for the initial condition, is set to "Not a road".

At this point, GRAIP_Lite now knows what treatments to model and where they will be applied. Save changes to the map document, close the GRAIP Lite window, and clear all selected features.

Often, more than one alternative needs to be analyzed in order to determine which alternative represents the best course of action. In such cases, the above process of creating an alternative and specifying treatments would be repeated for each alternative to be analyzed.

The next step runs the model for each time step in the alternative. Open the *01. Run Alternative* tool in the *Alternatives* toolbox (Figure 66). The *Input Road Alternative Feature Class* should be set to
Alternative1 Road; click OK to run the tool. This tool may take a while to run, especially with larger data sets and increasing amounts of road treated.

Figure 66: The Run Alternative tool.

The last tool to run is the 02. Alternative Report tool (Figure 67), also in the Alternatives toolbox. This tool creates the same set of reports as the Basic Report tool, but it does so for each time step for the alternative. It also produces two maps showing the change in specific sediment at the disturbed and recovered conditions relative to the initial condition; these are helpful for highlighting where road related sediment impacts are expected to change, and in which direction and at what time frame, during the course of the alternative. Make sure the location of the Input GRAIP Lite Database is correct, and set the Input Alternative Name to Alternative1. Click OK to run the tool.

Figure 67: The Alternative Report tool.

If you are analyzing multiple alternatives, the 01. Run Alternative and 02. Alternative Report tools will need to be run separately for each alternative.

Post Processing

One additional tool is the Create Summary Statistics by Area tool (Figure 68) that aggregates several metrics by catchment area. This is especially useful because it calculates the specific sediment delivery to the catchment areas (Figure 69), which is effectively a measure of road-related sediment impact in each catchment. This highlights both the stream segments and the direct contributing area where the roads are having the greatest impacts on the stream network.

Figure 68: The Create Summary Statistics by Area tool, ready to be run on the Basic Run tutorial dataset.

Figure 69: Specific Sediment Delivery (Mg/yr/km²) aggregated to hydrologic catchments.

In some cases, it may be necessary to manually aggregate the data. One way to do this is to use the Intersect tool and then summarize the data based on a column in the resulting data; start by intersecting the RoadDrainPoint layer with the layer that has the features you wish to aggregate the data to. Next, summarize the table based on a unique identifier for the features you are aggregating the data. Finally, join the summary table back to the aggregate features. Summary tables can also be created using columns in the original data, for instance, road number/name/ID or jurisdiction. Most often, the data to be aggregated will be the sums of road lengths and the sediment production and delivery values.

Appendix A: Description of Existing Calibration Data Sets

Default: This calibration set consists of the other calibration sets merged into a single calibration set. It totals 77,779 observations from 5,374 km of roads and has an observed stream connection rate of 15 percent. The mean flow distance from the observed calibration points to the stream network modelled by ArcHydro for GRAIP_Lite (stream distance) was 166 m and the standard deviation was 164 m.

The baserate used with the default calibration set is 79 kg/yr/m vertical drop along the road, and comes from the data in Luce and Black, 1999. This default baserate is also the default for the more detailed, inventory-based GRAIP model.

Flow Distance to Stream (m)

Andesite – Eldorado NF: This calibration set was collected as part of a GRAIP inventory conducted as part of restoration work associated with the Power Fire on the Eldorado National Forest. It consists of 1,638 calibration points collected from 124 km of road. The observed stream connection rate was 13 percent. The mean stream distance was 210 m with a standard deviation of 162 m. Elevations range from 1,117 m to 2,431 m, with a mean of 1,718 m and standard deviation of 273 m. Mean annual precipitation is 1,287 mm with a standard deviation across the calibration set of 83 mm.

The baserate used with this calibration set is 53 kg/yr/m vertical drop along the road, and comes from three sediment monitoring plots utilizing a main settling tank, tipping bucket flow gage, and a flow splitter leading to a fines collection tank. These plots were installed in 2015.

Andesite – Plumas NF: This calibration set was collected as part of a GRAIP inventory conducted as part of restoration work associated with the Moonlight Fire on the Plumas National Forest. It consists of 1,480 calibration points collected from 111 km of road. The observed stream connection rate was 15 percent. The mean stream distance was 189 m with a standard deviation of 158 m. Elevations range from 1,483 m to 2,182 m, with a mean of 1,906 m and standard deviation of 164 m. Mean annual precipitation is 803 mm with a standard deviation across the calibration set of 56 mm.

The baserate used with this calibration set is 77.6 kg/yr/m vertical drop along the road, and comes from four sediment monitoring plots utilizing a main settling tank, a flow splitter, and a second tank to collect fines. These plots were installed in 2014.

Basalt – Payette NF: This calibration set consists of GRAIP inventories collected between 2013 and 2015 as part of watershed assessment and project planning operations on the Payette National Forest. It consists of 10,799 calibration points collected from 806 km of road. The observed stream connection rate was 17 percent. The mean stream distance was 135 m with a standard deviation of 139 m. Elevations range from 1,279 m to 2,295 m, with a mean of 1,722 m and standard deviation of 158 m. Mean annual precipitation in 1,009 mm with a standard deviation across the calibration set of 118 mm.

The base rate is 27.2 kg/yr/m and was derived from 5 sediment plots measured twice a year starting in 2013. These five sediment plots all include tipping buckets and splitters feeding into fine sediment recovery tanks.

Basalt – Umatilla NF: This calibration set consists of a GRAIP inventory collected in support of TMDL analysis in the Wall Creek watershed on the Umatilla National Forest. It consists of 6,473 calibration points collected from 725 km of road. The observed stream connection rate was 26 percent. The mean stream distance was 104 m with a standard deviation of 128 m. Elevations range from 684 m to 1,530 m, with a mean of 1,178 m and standard deviation of 142 m. Mean annual precipitation in 464 mm with a standard deviation across the calibration set of 38 mm.

The base rate is 1.5 kg/yr/m and was assumed for the Wall Creek watershed inventory based on three years of data from nine native and aggregate surfaced roads near Klamath Falls, Oregon. For more information, see the Wall Creek Watershed GRAIP Roads Assessment.

Basalt/Sandstone – Siuslaw NF: This calibration set consists of a GRAIP inventory collected during 2010 and 2011 in the North Fork Siuslaw River watershed on the Siuslaw National Forest. It consists of 5,273 calibration points collected from 261 km of road. The observed stream connection rate was 4 percent. The mean stream distance was 271 m with a standard deviation of 124 m. Elevations range from 41 m to 583 m, with a mean of 200 m and standard deviation of 74 m. Mean annual precipitation in 2,096 mm with a standard deviation across the calibration set of 79 mm.

This calibration set uses the default baserate, which was developed nearby.

Belt Super Group – Coleville NF: This calibration set consists of a small GRAIP inventory collected as a baseline for planned fuels remediation treatments in the Deer Creek watershed on the Colville National Forest. Because this watershed functions as a municipal water supply for the town of Orient, WA, all roads in the watershed are closed to motorized vehicles. It consists of 768 calibration points collected from 70 km of road. The observed stream connection rate was 8 percent. The mean stream distance was 324 m with a standard deviation of 330 m. Elevations range from 464 m to 1,625 m, with a mean of 1,151 m and standard deviation of 276 m. Mean annual precipitation in 625 mm with a standard deviation across the calibration set of 39 mm.

This calibration uses a baserate of 14 kg/yr/m derived from the plots on the Lolo National Forest for the Belt Super Group – Lolo Helena Flathead NFs calibration set.

Belt Super Group – Lolo Helena Flathead NFs: This calibration set consists of several GRAIP watershed inventories collected from 2012 through 2013 on the Lolo, Helena, and Flathead National Forests as part of work being done by the South West Crown of the Continent Collaborative Forest Landscape Restoration Program. It consists of 10,826 calibration points collected from 616 km of road. The observed stream connection rate was 6 percent. The mean stream distance was 189 m with a standard deviation of 187 m. Elevations range from 1,081 m to 2,260 m, with a mean of 1,523 m and standard deviation of 246 m. Mean annual precipitation is 680 mm with a standard deviation across the calibration set of 165 mm.

This calibration uses a baserate of 14 kg/yr/m derived from 15 separate sediment monitoring plots. Eight of these plots use a large settling tank combined with a flow splitter and sediment blanket supported by a basket for fines collection; the other seven plots use a main settling tank, tipping bucket flow gage, and a flow splitter leading to a fines collection tank. The original eight plots, with filter fabric, were installed in 2011. Four plots were then installed in 2015, and the last three plots were installed in 2016.

100

Granite – Boise NF: This calibration set consists of several GRAIP watershed inventories collected on the Boise National Forest between 2009 and 2011 in support of TMDL, 4b, and other restoration projects. It consists of 27,430 calibration points collected from 1,655 km of road. The observed stream connection rate was 17 percent. The mean stream distance was 149 m with a standard deviation of 146 m. Elevations range from 932 m to 2,581 m, with a mean of 1,707 m and standard deviation of 336 m. Mean annual precipitation is 850 mm with a standard deviation across the calibration set of 195 mm.

This calibration uses a baserate of 21.3 kg/yr/m derived from 6 separate sediment monitoring plots installed on the Lightning Creek road. Five of these plots were installed in 2009 and the sixth plot was installed in 2010 using a prototype tipping bucket. Later, a second tipping bucket was added on one additional plot. Currently, four plots consist of just a main sediment tank and two plots have a main sediment tank, tipping bucket with splitter, and fines collection tank.

Flow Distance to Stream (m)

Granite – Eldorado NF: This calibration set was collected as part of a GRAIP inventory conducted as part of restoration work associated with the Power Fire on the Eldorado National Forest. It consists of 2,106 calibration points collected from 145 km of road. The observed stream connection rate was 17 percent. The mean stream distance was 192 m with a standard deviation of 199 m. Elevations range from 1,070 m to 2,394 m, with a mean of 1,766 m and standard deviation of 311 m. Mean annual precipitation is 1,302 mm with a standard deviation across the calibration set of 86 mm.

The baserate used with this calibration set is 49.5 kg/yr/m vertical drop along the road, and comes from three sediment monitoring plots utilizing a main settling tank, tipping bucket flow gage, and a flow splitter leading to a fines collection tank. These plots were installed in 2015.

Granite – Plumas NF: This calibration set was collected as part of a GRAIP inventory conducted as part of restoration work associated with the Moonlight Fire on the Plumas National Forest. It consists of 6,946 calibration points collected from 495 km of road. The observed stream connection rate was 11 percent. The mean stream distance was 149 m with a standard deviation of 149 m. Elevations range from 1,089 m to 2,260 m, with a mean of 1,701 m and standard deviation of 189 m. Mean annual precipitation is 751 mm with a standard deviation across the calibration set of 92 mm.

The baserate used with this calibration set is 30.2 kg/yr/m vertical drop along the road, and comes from four sediment monitoring plots utilizing a main settling tank, a flow splitter, and a second tank to collect fines. These plots were installed in 2014.

North Cascades – Mount Baker Snoqualmie NF: This calibration set consists of a GRAIP inventory collected during 2013 and 2014 in the Stilliguamish River watershed on the Mount Baker – Snoqualmie National Forest. It consists of 4,040 calibration points collected from 203 km of road. The observed stream connection rate was 35 percent. The mean stream distance was 227 m with a standard deviation of 211 m. Elevations range from 308 m to 1,132 m, with a mean of 631 m and standard deviation of 165 m. Mean annual precipitation is 3,147 mm with a standard deviation across the calibration set of 232 mm.

This calibration set uses the default baserate.

Appendix B: Description of Treatments

New Road – Temporary: This treatment is used when a road will be added for use during the project and then fully recontoured at the end of the project. For the initial condition, it assumes the road does not exist. For the disturbed condition, it assumes a bare, native surfaced, ML2 or equivalent road with high traffic and drainage at 100m distances. For the recovered condition, it assumes the road has been recontoured yielding a native surfaced ML1 with no traffic covered in vegetation and having drainage every 25m.

New Road – Permanent: This treatment is used when a road will be added for long-term access. For the initial condition, it assumes the road does not exist. For the disturbed condition, it assumes a bare, crushed rock surfaced, ML3 or equivalent road with high traffic and drainage at 100m distances. For the recovered condition, it assumes the road has come to vegetative equilibrium similar to other like roads and that traffic is low.

New Road – Reroute: This treatment is used when a road will be added to make a connection during a reroute project. For the initial condition, it assumes the road does not exist. For the disturbed condition, it assumes a bare, crushed rock surfaced, ML3 or equivalent road with high traffic and drainage at 100m distances. For the recovered condition, it assumes the road has come to vegetative equilibrium similar to other like roads and that traffic is low. The maintenance level and drainage distance should be set to match the road being replaced.

Re-Open – Temporary: This treatment is used when a closed road will be temporarily opened for use during a project and then recontoured. For the initial condition, it assumes the road is closed (ML1) with normal vegetation, drainage every 50m, and a native surface. For the disturbed condition, it assumes a bare, native surfaced, ML2 or equivalent road with high traffic and drainage at 100m distances. For the recovered condition, it assumes the road has been recontoured yielding a native surfaced ML1 with no traffic covered in vegetation and having drainage every 25m.

Re-Open – Permanent: This treatment is used when a closed road will be reopened, often as part of a reroute project. For the initial condition, it assumes the road is closed (ML1) with normal vegetation, drainage every 50m, and a native surface. For the disturbed condition, it assumes a bare, crushed rock surfaced, ML2 or equivalent road with high traffic and drainage at 50m distances. For the recovered condition, it assumes the road has come to vegetative equilibrium similar to other like roads and that traffic is low. If this is part of a reroute, the maintenance level and drainage distance should be set to match the road being replaced.

Traffic Increase – ML2 with gravel: This treatment is used when gravel will be applied to a ML2 road to support temporary increased traffic loads such as during haul operations. For the initial condition, it uses the road attributes from the initial GRAIP_Lite run based on the INFRA road layer. For the disturbed condition, it assumes a bare, crushed rock surface with high traffic and a drainage distance of 100m. For the recovered condition, it assumes that the surface has degraded back to a native surface, the vegetation has recovered to an average state, the traffic level has dropped to low, and the drainage distance has returned to 50m.

Traffic Increase – ML3 with gravel: This treatment is used when gravel will be applied to a ML3 road to support temporary increased traffic loads such as during haul operations. For the initial condition, it uses the road attributes from the initial GRAIP_Lite run based on the INFRA road layer. For the

disturbed condition, it assumes a bare, crushed rock surface with high traffic and a drainage distance of 100m. For the recovered condition, it assumes that the surface is crushed rock, the vegetation has recovered to an average state, the traffic level has dropped to medium, and the drainage distance is 100m.

Traffic Increase – ML4 with gravel: This treatment is used when gravel will be applied to a ML4 road to support temporary increased traffic loads such as during haul operations. For the initial condition, it uses the road attributes from the initial GRAIP_Lite run based on the INFRA road layer. For the disturbed condition, it assumes a bare, crushed rock surface with high traffic and a drainage distance of 100m. For the recovered condition, it assumes that the surface is crushed rock, the vegetation has recovered to an average state, the traffic level has dropped to medium, and the drainage distance is 100m.

Traffic Increase – ML5 Paved: This treatment is used when gravel will be applied to a ML5 road to support temporary increased traffic loads such as during haul operations. For the initial condition, it uses the road attributes from the initial GRAIP_Lite run based on the INFRA road layer. For the disturbed condition, it assumes a bare ditch, paved surface with high traffic and a drainage distance of 200m. For the recovered condition, it assumes that the vegetation in the ditch has returned to the average condition.

SDRR Improvement – Nat-Gravel ML2: This treatment is used to simulate the effect of converting a native surfaced ML2 road to a crushed rock surface. For the initial condition, it uses the road attributes from the initial GRAIP_Lite run based on the INFRA road layer. For the disturbed condition, it assumes a bare crushed rock surface with high traffic and a 50m drainage spacing. For the recovered condition, it assumes that the traffic has decreased to low and that the vegetation has returned to average conditions.

SDRR Improvement – Nat-Gravel ML3: This treatment is used to simulate the effect of converting a native surfaced ML3 road to a crushed rock surface. For the initial condition, it uses the road attributes from the initial GRAIP_Lite run based on the INFRA road layer. For the disturbed condition, it assumes a bare crushed rock surface with high traffic and a 100m drainage spacing. For the recovered condition, it assumes that the traffic has decreased to medium and that the vegetation has returned to average conditions.

SDRR Improvement – Drainage ML3: This treatment is used to simulate the effect of increasing the drainage density of a ML3 road with a gravel surface. For the initial condition, it uses the road attributes from the initial GRAIP_Lite run based on the INFRA road layer. For the disturbed condition, it assumes a bare crushed rock surface with high traffic and a 50m drainage spacing. For the recovered condition, it assumes that the traffic has decreased to medium and that the vegetation has returned to average conditions.

Reconstruct – ML2: This treatment is used to simulate the effect of reconstructing a native surfaced ML2 road to a using a crushed rock surface. For the initial condition, it uses the road attributes from the initial GRAIP_Lite run based on the INFRA road layer. For the disturbed condition, it assumes a bare crushed rock surface with high traffic and a 100m drainage spacing. For the recovered condition, it assumes that the traffic has decreased to low and that the vegetation has returned to average conditions.

Reconstruct – ML3: This treatment is used to simulate the effect of reconstructing an ML3 road using a crushed rock surface. It is functionally the same as Traffic Increase – ML3 with gravel.

Upgrade – Pave: This treatment is used to model paving a road and establishing a 200m drainage spacing. For the initial condition, it uses the road attributes from the initial GRAIP Lite run based on the INFRA road layer. For the disturbed condition, it assumes a paved road surface with a bare ditch, high traffic, and a 200m drainage spacing. For the recovered condition, vegetation in the ditch is returned to an average condition and traffic remains high.

Storage – Close ML2: This treatment is used to model long-term closure (storage) of an ML2 road. For the initial condition, it uses the road attributes from the initial GRAIP_Lite run based on the INFRA road layer. For the disturbed condition, it assumes a native surface with no traffic and a 50m drainage spacing; vegetation is modeled as the average condition for the ML2 road. For the recovered condition, vegetation is modeled as the average condition on ML1 roads.

Storage – Drainage Removal ML2: This treatment is used to model long-term closure (storage) of an ML2 road and specifies the removal of existing drainage culverts; it is modeled the same as the Storage – Close ML2 treatment. For the initial condition, it uses the road attributes from the initial GRAIP_Lite run based on the INFRA road layer. For the disturbed condition, it assumes a native surface with no traffic and a 50m drainage spacing; vegetation is modeled as the average condition for the ML2 road. For the recovered condition, vegetation is modeled as the average condition on ML1 roads.

Decommission – Rip/Till: This treatment is used to model decommissioning a road by ripping or tilling, rather than recontouring, and then seeding or planting vegetation. For the initial condition, it uses the road attributes from the initial GRAIP_Lite run based on the INFRA road layer. For the disturbed condition, it assumes a bare native surface with no traffic and a 50m drainage spacing. For the recovered condition, vegetation is modeled as covered. If treatment will not involve seeding or planting, V2 should be set to Default rather than Covered.

Decommission – Partial Recontour: This treatment is used to model decommissioning a road by partially recontouring the road surface, and then seeding or planting vegetation. For the initial condition, it uses the road attributes from the initial GRAIP Lite run based on the INFRA road layer. For the disturbed condition, it assumes a bare native surface with no traffic and a 50m drainage spacing. For the recovered condition, vegetation is modeled as covered. If treatment will not involve seeding or planting, V2 should be set to Default rather than Covered.

Decommission – Full Recontour: This treatment is used to model decommissioning a road by fully recontouring the road surface, and then seeding or planting vegetation. For the initial condition, it uses the road attributes from the initial GRAIP_Lite run based on the INFRA road layer. For the disturbed condition, it assumes a bare native surface with no traffic and a 25m drainage spacing. For the recovered condition, vegetation is modeled as covered. If treatment will not involve seeding or planting, V2 should be set to Default rather than Covered.

Use and Decommission – ML2 Recontour: This treatment is used to model using a road for a project and then recontouring that road after project completion. For the initial condition, it uses the road attributes from the initial GRAIP_Lite run based on the INFRA road layer. For the disturbed condition, it assumes a bare native surface with high traffic and a 100m drainage spacing. For the recovered

condition, it assumes a covered native surface with no traffic and a 25m drainage spacing. If treatment will not involve seeding or planting, V2 should be set to Default rather than Covered.

Use and Decommission – ML3 Recontour: This treatment is used to model using a road for a project and then recontouring that road after project completion. For the initial condition, it uses the road attributes from the initial GRAIP_Lite run based on the INFRA road layer. For the disturbed condition, it assumes a bare crushed rock surface with high traffic and a 100m drainage spacing. For the recovered condition, it assumes a covered native surface with no traffic and a 25m drainage spacing. If treatment will not involve seeding or planting, V2 should be set to Default rather than Covered.

Appendix C: Data Structure and Requirements

Input Known Drains **Input Known Drains Input Known Drains** shapefile or feature class

Appendix D: Basic Report Example

Annual Road Surface Sediment Delivery (kg/yr)

Inventoried Roads - **Inventoried Roa ds**

0 2.5 5 7.5 10
■ ■ Kilometers

Road Surface Specific Sediment Accumulation in Streams (ton/yr/sqkm)

Inventoried Roads - Inventoried Roads

2.5 5 7.5 10
Kilometers C

Appendix E: Alternative Report Example

Annual Road Surface Sediment Delivery (kg/yr)

Alternative1 - **Current**

Alternative1 - Disturbed

Alternative1 - Long Term

Alternative1 - Current

Alternative1 - Disturbed

Alternative1 - **Long Term**

Alternative1 - Current

Alternative1 - Disturbed

Alternative1 - Long Term

Road Surface Specific Sediment Accumulation in Streams (ton/yr/sqkm) Alternative1 - Current

Road Surface Specific Sediment Accumulation in Streams (ton/yr/sqkm) Alternative1 - Disturbed

Road Surface Specific Sediment Accumulation in Streams (ton/yr/sqkm) Alternative1 - Long Term

 $- 0.000 - 1.478$ • 1.478 - 3.775 $-3.775 - 8.253$ $- 8.253 - 16.795$ - **No Sediment Delivery Inventoried Roads** - **Inventoried Roads** \mathbb{E} \mathcal{L} $z \gg k$ -7.195 $\mathbb{R} \cup$ \int_0^\star , ' . 0 0.75 1.5 2.25 3
■ ■ ■ ■ Kilometers

Change in Road Surface Specific Sediment Load in Streams (ton/yr/sqkm) Alternative1 - **Disturbed**

Change in Road Surface Specific Sediment Load in Streams (ton/yr/sqkm) Alternative1 - **Long Term**

Appendix F: Tips and Tricks

How to change traffic levels, surface types, maintenance levels, and split distances when using the Processing tools in GRAIP_Lite.

Run the Initialize GRAIP Lite Database tool. This imports the roads and creates the GL_Traffic field in the Road feature class. The process to change the traffic could also be used to update surface type or maintenance level fields as well to correct inaccuracies in the INFRA layer used as input.

- 1. Start an editing session and select the Road feature class as your editing target.
- 2. Select the roads on which you need to make the changes.
- 3. Open the attribute table for the Road feature class and scroll to the right until you find GL_Traffic.
- 4. For your selected roads, click in the GL_Traffic field, which should open a drop-down menu of the available options.
- 5. Select the option you want, which won't be <null> unless you are trying to crash GRAIP_Lite.
- 6. Save edits and exit the editing session.
- 7. Continue with the GRAIP_Lite run.

For numeric fields like GL_MaintenanceLevel or GL_SplitDistance, you don't get a drop-down menu so you need to be a little more careful about what numbers you put in there, especially for maintenance level.

How to reset the calibration zone from Default.

Run the Initialize GRAIP Lite Database tool. One of the things this does is create a CalibrationZone feature class in the geodatabase if you haven't provided one. The CalibrationZone will be the same extent as the DEM you supplied, and stores the name of the calibration data set to be used in the GL_CalibrationZone field. If you change this name to match a different existing calibration data set, it will make the model use the corresponding calibration data for baserate, vegetation factors, and delivery curves. If you change it to something that does not match the names used for existing calibration data sets, you need to provide the data for the calibration or the model will crash.

Glossary

Calibration zone: A polygon that defines which calibration, with its attendant baserate, vegetation factors, and delivery curves, is used for modelling roads within that polygon.

Connected road density: The density of stream-connected road within the contributing watershed area for a given stream reach, reported as GL_RoadDen in the DrainageLine feature class. Connected road length is determined for each road segment as the road length multiplied by the delivery probability; this is then accumulated through the stream network and normalized by contributing area. The units are km/km².

Connected road length: The length of stream-connected road within the contributing watershed area for a given stream reach, reported as GL_RoadLen in the DrainageLine feature class. Connected road length is determined for each road segment as the road length multiplied by the delivery probability; this is then accumulated through the stream network. The units are km.

Drainpoint: A point on the road network where water and sediment leave the road. In GRAIP_Lite, these are arbitrarily defined based on topography and other factors except where known drainpoints are present in the known drainpoints layer.

Sediment accumulation: The sum of all delivered sediment that is routed to a given stream reach, reported as GL_SedAccum in the DrainageLine feature class. The units are Mg/yr.

Sediment delivery: That portion of sediment produced on the road surface that makes it into the stream network, reported as GL_SedDel in the RoadSegment, RoadDrainPoint, and RoadDrainPointDiss feature classes. The units are kg/yr.

Sediment production: Surface erosion generated on the road tread and ditch, reported as GL_SedProd in the RoadSegment and RoadDrainPoint feature classes. The units are kg/yr.

Specific sediment delivery: Sediment accumulation normalized by contributing area, reported as GL_SpecSedDel in the DrainageLine feature class. The units are Mg/yr/km².

References

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Methylmercury Exposure and Health Effects in Humans: A Worldwide Concern

The paper builds on existing literature, highlighting current understanding and identifying unresolved issues about MeHg exposure, health effects, and risk assessment, and concludes with a consensus statement. Methylmercury is a potent toxin, bioaccumulated and concentrated through the aquatic food chain, placing at risk people, throughout the globe and across the socioeconomic spectrum, who consume predatory fish or for whom fish is a dietary mainstay. Methylmercury developmental neurotoxicity has constituted the basis for risk assessments and public health policies. Despite gaps in our knowledge on new bioindicators of exposure, factors that influence MeHg uptake and toxicity, toxicokinetics, neurologic and cardiovascular effects in adult populations, and the nutritional benefits and risks from the large number of marine and freshwater fish and fish-eating species, the panel concluded that to preserve human health, all efforts need to be made to reduce and eliminate sources of exposure.

INTRODUCTION

The Panel on Health Risks and Toxicological Effects of Methylmercury received the mandate to describe and synthesize current scientific knowledge on methylmercury (MeHg) exposure and its effects in humans and to identify research gaps. The present paper is not intended to be a comprehensive review and presentation of all the literature on MeHg exposure and effects in humans but builds on earlier literature, other reviews, and more recent literature in highlighting the current understanding in the field and what we consider to be remaining unresolved issues. Humans are exposed to different forms of mercury (Hg), and potential health risks from forms other than MeHg can occur, including mercury vapor from dental amalgams, as well as from occupational exposures (e.g., dental offices, chloralkali plants, fluorescent lamp factories, mercury mining) and from artesanal and small-scale gold and silver mining operations (1– 5), the present document does not cover these exposures, because the pathways of exposure and effects differ from those for MeHg. Here, we examine issues of MeHg exposure, studies on its health effects and major risk assessments, and conclude with our consensus statement.

MeHg Exposure

Sources of exposure. Methylmercury contamination poses a particular challenge to public health because this toxicant is mainly contained in fish, a highly nutritious food, with known benefits for human health. Moreover, fish are culturally vital for many communities and constitute an important global commodity. Although we often refer to "fish" in a generic way, all fish do not have similar amounts of mercury. As a result of bioaccumulation of MeHg through multiple levels of the aquatic food web, higher tropic-level pelagic fish can be contaminated with MeHg at concentrations in excess of 1 part

per million (ppm). The concentrations of total Hg vary widely across fish and shellfish species, with the mean values differing by as much as 100-fold (6). Methylmercury is bound to proteins, as well as to free amino acids, that are components of muscle tissues, and are not removed by any cooking or cleaning processes that do not destroy muscle tissues.

Although in general, MeHg accumulates in fish through the food chain, consumption of farmed fish can also lead to MeHg exposures, in part, because of the presence of MeHg in feed (7). Some studies have shown no significant difference in MeHg levels in farmed *vs.* wild salmon, although concentrations in both cases are relatively low (8, 9). Although fish and shellfish are the predominant sources of MeHg in the diets of humans and wildlife, a few reports of other sources exist. Rice cultivated in areas contaminated with mercury can contain relatively high levels of MeHg (10). Methylmercury has also been reported in organ meats of terrestrial animals (11), as well as in chicken and pork, probably as a result of the use of fish meal as livestock feed (12). Some communities also have higher MeHg exposure because of the consumption of fish-eating marine mammals (13, 14).

Profiles of exposure. Although most reports on MeHg exposure focused on specific populations generally assumed to have high levels of fish consumption, estimates of general populations exposure exist for the United States (15, 16), Germany (17), and Japan (18) [summarized in Pirrone and Mahaffey (19)]. For populations that are not selected on the basis of high fish consumption, mean hair Hg levels generally range from >0.1 μ g g⁻¹ to <1.0 μ g g⁻¹ (20–25). The mean blood Hg for such populations is generally in the range of <1.0 μ g L⁻¹ to <5.0 μ g L⁻¹, although, worldwide there are fewer data on MeHg exposure based on blood than on hair. In the United States nationally, about 5–10% of the population of women of childbearing age have hair levels exceeding 1.0 μ g g⁻¹ (16) and blood levels exceeding 5 μ g L⁻¹ (26). In Japan, where more fish is consumed, 73.7% of women of this age have hair levels above 1.0 μ g g⁻¹ and 1.7% above 5 μ g g⁻¹ (18). In Germany, the 1998 geometric mean blood level was 0.58 μ g L⁻¹ (17).

High levels of Hg exposure were identified in numerous fisheating populations throughout the world [for reviews see: Pirrone and Mahaffey (19)]. Many of these live near oceans, major lakes and rivers, or hydroelectric dams, and are often dependent on local catch, with fish an integral part of their cultural traditions. In the sea islands of the Faroes and Seychelles, median mothers' hair Hg concentrations were 4.5 μ g g⁻¹ [with 27% above 10 μ g g⁻¹ (27)] and 5.8 μ g g⁻¹ (28), respectively. In the river basins of the Amazon, where a large number of studies was carried out on populations for whom freshwater fish is a dietary mainstay, median hair Hg levels typically range between 5 μ g g⁻¹ and 15 μ g g⁻¹ (29–34).

Despite the importance of local catch, fish is also a global commodity and market fish, such as shark, tuna, and swordfish, or canned white tuna (35), consumed by persons living far away from the source can likewise have high levels of MeHg. In the United States, individuals with high blood Hg concentrations were reported among affluent urbanites who ate large quantities

of marine fish, high in the food web (36, 37). Thus, elevated MeHg exposure is present around the globe, with no geographic, social, economic, or cultural boundaries.

Biomarkers of MeHg exposure. Hair and blood Hg concentrations are both accepted as valid biomarkers of MeHg exposure, although each provides a somewhat different reflection of exposure (38). Blood gives an estimate of exposure over the most recent one to two half-lives, with the half-life of MeHg in blood being 50–70 days, whereas hair reflects the average exposure over the growth period of the segment (28). Hair Hg is predominantly MeHg, with MeHg constituting from 80% to 98% of hair total Hg (33, 39). For populations with regular and frequent fish consumption, hair total Hg and blood MeHg are consistently correlated (40). Generally, hair is 250 to 300 times more concentrated in mercury than is blood (39). However, in populations and individuals with infrequent fish consumption or where bolus doses of MeHg occur, there can be considerable inter- and intraindividual variability in the relation between hair and blood Hg levels resulting from temporal differences in the retention of Hg by each biomarker (33, 40, 41). Segmental analyses of hair Hg can provide a chronology of exposure over time (24, 28, 29, 33). However, information on short-term peaks in exposure is not well represented by such analyses (38). Another consideration is that the growth rate of hair, generally estimated at 1 cm mo^{-1} , can have both inter-and intraindividual variability (38). Recent advances in a single hairstrand analysis (42), including measurement of Hg at micron resolution by using laser ablation (43) should yield more information on the relation between Hg uptake and Hg deposition in hair.

The Hg levels in toenails and fingernails also were used as biomarkers of Hg exposure, mostly in major studies of the cardiovascular effect of MeHg (see below) (44, 45), but to what extent these reflect organic or inorganic Hg exposures remains to be clarified (46). A recent study of women, with no history of occupational exposure to Hg, showed similar correlations between Hg intake through fish consumption and both toenail and hair Hg concentrations; however, only total Hg was assessed (47). In this study, hair, toenail, and urinary total Hg were highly correlated. Urinary Hg levels largely reflect exposure to inorganic Hg (40) and are not considered useful bioindicators of MeHg exposure. There are, however, several recent reports of positive correlations between fish consumption and urinary Hg (48–50), and investigators of these studies propose that demethylation may account, at least partially, for this observation. The relation of fish consumption and inorganic Hg in different biological tissues, and its consequence for human health still need to be elucidated.

Health effects from low to moderate levels of MeHg exposure were reported in a variety of systems and domains. Each of these effects may depend on different aspects of exposure [e.g., fish-eating patterns, time of exposure (first, second, or third trimester, childhood, adulthood)]. Therefore, the different reflections of exposure provided by hair and blood Hg concentrations may provide different information about dose-response for different exposure populations and different exposure scenarios. Few studies investigated side-by-side doseresponse relations for both biomarkers. In the study in the Faroe Islands, maternal hair and fetal-cord blood predicted similar but not identical patterns of effect across various measures of neurologic performance (38).

Fish Consumption as a Predictor of MeHg Exposure

Exposure dose. Although most studies identified a clear association between the quantity and the frequency of fish consumption and Hg exposure, there is considerable interindividual and intergroup variability in the relation between the amount or the frequency of fish consumption and the levels of biomarker of MeHg exposure. Several factors mediate this relation. The MeHg concentration within and across species of dietary fish is an obvious source of variability. For example, those who eat mainly carnivorous fish and/or fish-eating mammals have relatively higher levels of Hg compared with those who eat mainly noncarnivorous fish (14, 29, 33, 51–54). Independent of the MeHg concentration, the frequency of fish consumption is also an important factor in this variability. Because biomarkers reflect the weighted average of exposure over time, short-term reporting of fish consumption may not correspond with a longer-term average of MeHg exposure. Under some circumstances, episodic exposures can result in large bolus doses of MeHg. Bolus doses can arise, for example, from the infrequent consumption of fish or fish-eating mammals with high concentrations of MeHg. Given practical limitations in sampling frequency, as well as the nature of some of the biomarkers themselves, bolus doses during putative discrete windows of sensitivity in fetal development may not be fully revealed by biomarkers of exposure.

Toxicokinetics. Although most experimental studies on the gastrointestinal absorption of MeHg indicated that nearly 100% of MeHg in fish is absorbed, recently reported animal and human data suggest that there may be substantial variability (55, 56). In animal studies, variation in absorption kinetics was related to factors such as sex and age (57). A further gap exists because human absorption studies were primarily conducted in adult male subjects.

Toxicokinetic (pharmacokinetic) models and physiologically based pharmacokinetic (PBPK) models are applied to estimate internal dose, given a known intake dose, as well as the intake dose, given a measured internal dose (38). The basic onecompartment model (39, 58, 59) is a steady-state model that is intended to predict concentration in a single compartment only (generally, blood). As such, it is less flexible than the PBPK models in predicting nonsteady state conditions and concentrations in other compartments. However, its relative simplicity has allowed it to be used with probabilistic input parameters to obtain estimates of population variability in predictions of blood concentration and intake dose (60). Estimates of concentrations in other compartments (e.g., cord blood) can be made based on empirical ratios relating mercury concentration in blood to mercury concentrations in those compartments (61).

The PBPK models have the potential to predict changes in MeHg concentration in various tissues in response to changes in MeHg intake and in response to physiological changes (e.g., pregnancy, growth). They can be used to predict short-term changes in MeHg concentrations in different compartments during intake and distribution among compartments, if the parameters are correct (62–65).

The validity of these models overall is not thoroughly established under a range of exposures to MeHg by comparison with actual human data. Although they have the theoretical advantage of making predictions under dynamic conditions, the PBPK models are computationally complex and require data on many parameters whose MeHg-specific values have not been defined. This lack of MeHg specific values is a major limitation, particularly for predicting population variability. The extent to which these models rely on coefficients derived from metabolic studies and/or physiological parameters obtained in different populations and subpopulations and studies with other metals/ elements, limits their utility. Nonetheless, both simple toxicokinetic models and PBPK models have been used with reasonable consistency for setting public health guidance.

In humans, there is increasing evidence from environmental epidemiology studies of ethnic differences in the relation

between Hg intake from fish consumption and bioindicators of exposure (56), suggesting that diet and/or metabolic differences may be influencing mercury uptake and/or excretion. As yet, such differences have not been investigated in metabolic studies. Several studies suggest that selenium (Se) may play a role in MeHg absorption or excretion (66–68), but these data are not consistent. In the Brazilian Amazon, fruit consumption was associated with lower hair Hg concentrations (69). A positive relation was reported between iron and Hg in blood samples collected from Sweden (70). Overall, little is known about the factors that may modulate Hg absorption in humans, and research is needed to better understand this complex issue.

Fetal and infant exposure. One area in which the toxicokinetic data is consistent is the finding that MeHg is actively transferred to the fetus across the placenta via neutral amino acid carriers during gestation (71, 72). Although maternal and cord blood Hg concentration is highly correlated, cord blood MeHg is consistently higher than the corresponding maternal concentration, with an average ratio of about 1.7 (24, 61, 73, 74). Consequently, biomonitoring adult women's blood MeHg as a surrogate for potential fetalexposure, the corresponding fetal level will be, on average, 70% higher than maternal blood and up to three times higher at the 95th percentile. The maternal body burden of MeHg tends to decrease during gestation consistent with hemodilution and a transfer of a portion of the maternal body burden to the fetus (24).

Neonatal and infant exposure to MeHg occurs through intake of mother's milk, which is derived from maternal plasma, has a lower level of MeHg, and is enriched in inorganic Hg relative to the whole blood (75). Thus, lactational exposure to MeHg is reduced compared with what would be expected on the basis of maternal blood MeHg. Human and animal studies showed that, after birth, there is a decline in MeHg levels, reaching 40–50% at 2–3 months of age (76–78). During this period, infant body weight increases about 1.5–2 times. Consequently, the rapid increase in body volume and the limited MeHg transfer appear to explain the dilution of MeHg in infants during breast feeding.

HEALTH EFFECTS

Neurological Endpoints

Clinical manifestations. In 1958, McAlpine and Araki (79) linked the unusual neurological disease that was associated with fish consumption from Minamata Bay to MeHg exposure. This historic recognition of the brain and nervous system as the primary target organ for MeHg poisoning, resulting in marked distal sensory disturbances, constriction of visual fields, ataxia, dysarthria, auditory disturbances, and tremor, remains unchanged (80, 81). Based on analysis of the studies of human poisoning, the World Health Organization (WHO) (39) estimated that 5% of MeHg-exposed adults would experience neurologic effects with a blood Hg level of 200 μ g L⁻ (corresponding to a hair level of approximately 50 μ g g⁻¹). This estimate, however, was called into question by a re-analysis of these studies by Kosatsky and Foran (82), who suggested that the lowest observed effect level for clinical effects is likely to be considerably lower. Indeed, anecdotal and case reports of diffuse and subjective neurologic symptoms in adults and older children with moderately elevated MeHg exposures continue to appear (36, 83). In many cases, cessation or significant curtailing of fish consumption results in improvement of symptoms in conjunction with reduction in biomarker concentrations. These suggest the possibility of clinical effects, perhaps in a sensitive subset of the general population, at levels of exposure considerably below those previously associated with clinical effects in poisoning episodes. Currently, there is no formal case description or diagnostic criteria for such effects.

Although exposures throughout the world are lower than those producing the historical epidemics of MeHg poisoning, there is growing evidence that for many populations, exposures are sufficient to alter normal functioning of several systems, constitutes an important public health problem.

Effects in neonates, infants, and children. The poisoning in Minamata brought attention to the risk from fetal exposure. Exposed to MeHg through the placenta of the exposed mother, infants showed severe cerebral palsy–like symptoms, even when their mothers had mild or no manifestation of the poisoning (84). Mental retardation, cerebellar ataxia, primitive reflexes, dysarthria, and hyperkinesias were observed. These symptoms, described over 25 years ago (80, 85), continue as the clinical hallmark of congenital MeHg poisoning. Reconstruction of maternal or fetal doses resulting in these symptoms is difficult because of a lack of concurrent sampling. An estimate of the mean maternal hair concentration, resulting in such symptoms of 41μ g g⁻¹ ppm was proposed (86); however, a large uncertainty surrounds this estimate. Health effects observed with frank poisonings should not be confused with the more subtle, populational effects observed at lower levels of exposure.

At the subclinical and the population level, several studies in different parts of the world report poorer neurologic status and slower development in newborns, infants, and/or children exposed to MeHg *in utero* and/or during early childhood (87– 98), although some studies did not observe effects (99–101). In children, MeHg exposure *in utero* is associated with lower performance on tests of language, attention, memory, and/or visuospatial and/or motor functions. Although most child studies focused on fish-eating populations with relatively high levels of MeHg exposure, in a recent study, Oken et al (90) observed an inverse relation between mercury concentration in maternal hair and infants' performance on a visual recognition memory task at levels of mercury exposure consistent with background exposure in the US population (maternal hair levels varied between 0.02–2.38 μ g g⁻¹). Interestingly, in this study, fish consumption *per se* was associated with better performance, suggesting that some positive aspects of fish consumption, perhaps n-3 (omega-3) fatty acids, are reduced or antagonized by the MeHg contained in the same fish. A similar picture is emerging among adults for the some of the cardiovascular effects of MeHg (see below).

The two major ongoing longitudinal cohort studies on children from the Faroe Islands and the Seychelles are worthy of particular mention because they have both been following children through teenage years, assessing neuropsychological performance as a function of current, childhood, and *in utero* exposure. The Faroes study consistently observed neurobehavioral deficits associated with *in utero* exposure, even when children whose mother's hair Hg levels above 10 μ g g⁻¹ were excluded (91). In the initial studies of the Seychelles cohort, no effects were observed (100–103). However, recent reports of the Seychelles 9-year-old cohort shows decreases in fine motor function associated with higher fetal exposure levels (\geq 10 μ g g⁻¹ maternal hair); the investigators suggest that adverse effects may become apparent on higher-order cognitive functions that develop with maturity (104, 105). There has been much discussion about the differences between these two wellperformed studies. Factors such as type of exposure (one of the main exposure pathways in the Faroes study is through pilot whale, while in the Seychelles, it is entirely marine fish), biomarkers of exposure (cord blood *vs.* maternal hair), differences in test batteries and age of testing; cohort size and power were considered as possible explanations for the

differences in observed outcomes. However, none of these explanations proved entirely satisfactory or clearly decisive (38, 106). Other hypotheses, such as dietary intake of nutrients that may modify Hg metabolism or toxicity, were also proposed (69). Despite whatever significant differences do, in fact, exist between the Seychelles and Faroes studies that may explain differences in results that were observed to this time, the most recent results from the Seychelles appear to indicate a convergence in findings. More work needs to be done on factors that may affect the patterns of manifestation of Hg toxicity.

Neurophysiologic studies offer strong support for nervoussystem alterations associated with MeHg exposure. These studies showed mercury-related delayed latencies for auditory and visual evoked potentials (107–110). In the Faroes longitudinal study, latency delays were observed at 7 and 14 years (107, 109). No significant dose-effect relations for evoked potentials were observed in a study of Japanese children with low mercury exposure (maternal and children hair mercury levels of 1.6 μ g $g^ ^{-1}$) (111).

Nervous system endpoints in adults. Fewer studies addressed the neurotoxic effects of Hg exposure in adults. Mercury-related deficits in motor, psychomotor, visual and/or cognitive functions have been reported for different populations within the Brazilian Amazon (112–115) and for tuna consumers from the Mediterranean (116). A recent study, in the United States, of older adults (50–70 years old) with considerably lower blood Hg levels (mean, 2.1 μ g L⁻¹) showed inconsistent evidence of effect across neurobehavioral tests (117). Studies of associations between neurobehavioral outcomes and MeHg exposure in adult populations in which frequent and lifetime fish consumption is a cultural norm, generally cannot distinguish between effects because of adult exposure and permanent developmental effects because of gestational and early childhood exposures.

Cardiovascular Endpoints

A body of evidence was developed that addresses potential associations between MeHg and a range of cardiovascular effects. These include cardiovascular disease [coronary heart disease, acute myocardial infarction (AMI), ischemic heart disease], blood pressure and hypertension effects, and alterations in heart rate variability [see Chan and Egeland (118) and Stern (119) for recent reviews]. The strongest evidence for causal associations is for cardiovascular disease, particularly AMI in adult men (44, 120–122). In general, the relative risk and the odds ratios for AMI from these studies showed a doubling in the upper range of the observed Hg exposures. Comparison of exposures in these studies to exposures in Western populations suggests that the upper percentiles of current levels of exposure in these populations may result in a significantly elevated risk of AMI. Another well-conducted study of US health professionals, however, did not find an association between Hg exposure and coronary heart disease (123). This may be because dentists with possible exposure to elemental mercury accounted for 63% of controls and had a Hg exposure more than twice that of the other groups in the cohort. It is not known whether elemental or inorganic Hg acts similarly to MeHg with respect to cardiovascular effects. In addition, two of these studies used toenail Hg as the biomarker of exposure. Because this biomarker has not been adequately compared with the more common exposure biomarkers of hair or blood Hg, it is difficult to assess the doseresponse implications of these studies in relation to current exposures.

The evidence for an association between MeHg and other cardiovascular endpoints is weaker. An association was found between increased systolic and diastolic blood pressure in Faroese children at 7 years old and gestational exposure to MeHg (124). However, the association did not persist when the cohort was re-examined at 14 years old (125). Decreased heart rate variability was also associated with MeHg exposure, and this effect persisted through 14 years of age, but the implications of this effect in children for clinically significant outcomes is not clear. There are few studies that relate adult blood pressure to MeHg exposure. A recent study in the Brazilian Amazon reported that persons with 10 μ g g⁻¹ hair Hg were three times more likely to have elevated systolic blood pressure $(\geq 130 \text{ mm})$ Hg) (126), whereas in a study of women from the United States, no clear association was observed (127).

Reproductive Outcomes

The effect of MeHg on the sex ratio of offspring at birth and stillbirth in Minamata City, Japan, in the 1950s and 1960s, including the period when MeHg pollution was most severe, showed decreases in male birth in offspring in the overall city population, among fishing families (72, 128). An increase in the proportion of male stillborn fetuses raises the possibility that increased susceptibility of male fetuses to death *in utero* could explain the altered sex ratio.

Immune System Effects

Inorganic mercury was shown to suppress immune functions and to induce autoimmunity in multiple species (129). Both MeHg and inorganic Hg were shown to produce an autoimmune response, as well as an immunosuppressive effect in several strains of genetically susceptible mice (130, 131). However, data on the immune effects of MeHg in general are sparse, and research is required in this area.

Co-contaminants

Fish tend to accumulate halogenated organics, including polychlorinated biphenyls (PCB), dioxins, and related compounds. The neurodevelopmental effects of PCBs and, to a lesser extent, dioxins, share some similarities to those observed for MeHg (132). This can potentially present difficulties in determining causality and in constructing MeHg-specific doseresponse relations. Because MeHg tends to associate more with proteins than with fats, fish species with elevated levels of MeHg are not necessarily those with elevated levels of the lipophilic halogenated organics. Thus, for fish consumption where both exposures occur, the influence of the individual contaminants can potentially be separated by statistical techniques if a variety of fish species is consumed and sufficiently precise exposure metrics are collected. In the Faroe Islands studies, both MeHg and PCBs appear to jointly affect some developmental endpoints. However, although MeHg appeared to enhance the PCB-attributable effects, the PCBs appeared to make a relatively minor contribution to the MeHg-specific effects (132, 133). Contradictory findings were observed in a study of cognitive development associated with exposures to MeHg and PCBs in the Lake Oswego area of New York State (134). In that study, elevated PCB exposure appeared to potentiate MeHg effects. However, both MeHg and PCB levels were considerably lower than in the Faroes study, and no PCB-MeHg association was observed on follow-up testing of the cohort. More work remains to be done on the joint influence of MeHg and halogenated organics, as well as other metal contaminants that may also be present in fish (135).

Elemental Hg continues to be used in dental amalgam for the treatment of dental carries. In populations with significant amalgam use, elemental Hg may account for a proportion of total Hg exposure comparable with or greater than MeHg (38).

It is known that elemental Hg vapor can cross the placenta and accumulate in fetal tissue (136–138), and animal data suggest that elemental Hg has the potential to cause adverse neurologic developmental effects (139). Both elemental Hg and MeHg are metabolized in the brain to the inorganic mercuric form (38). It is not known whether the ultimate neurodevelopmental toxicant of MeHg is MeHg itself, the inorganic mercuric ion, free radicals generated in the conversion to the inorganic species, or some combination of these. If the inorganic form is the ultimate toxicant of MeHg in the developing brain or if MeHg and inorganic Hg share common neurodevelopmental toxic mechanisms, then current estimates of risk based on MeHg exposure alone could underestimate the population risk. Additional research is clearly needed to address these questions.

Potential Benefits of Fish Consumption

Several investigators have addressed the issues surrounding the risks and benefits associated with fish consumption, in general and for remote communities that depend on fish traditionally and/or as their dietary mainstay (69, 140–142). Indeed, for many populations, fish is the primary source of protein and other nutrients. Moreover, some fish can be an important source of the omega-3 fatty acids, eicosapentaenoic acid and docosahexaenoic acid, that appear to have positive effects on at least some of the same systems adversely affected by MeHg. However, similar to MeHg, there is considerable variability in the occurrence of omega-3 fatty acids across species (143). Fatty fish have higher levels of omega-3s compared with lean fish, and freshwater fish largely have lower levels of omega-3 fatty acids compared with ocean fish (15). There is no association between MeHg concentration of the fish or shellfish species and the omega-3 fatty acid level of the species (15). Several fish and shellfish species that are low in MeHg are high in omega-3 fatty acids (e.g., anchovies, herring, salmon), whereas others that are high in MeHg can be comparatively low in omega-3 fatty acids (e.g., shark, swordfish, pike) (15).

Omega-3 fatty acids are associated with beneficial effects on neurologic development in some studies (15), as has fish consumption in general, possibly as a correlate of omega-3 intake (90). However, not all studies found such a benefit (15, 144). Omega-3 fatty acids also were linked to a reduction in the risk of cardiovascular disease (44), although such an association recently were called into question in a comprehensive review (145). For both endpoints, there is some evidence suggesting that, in addition to its intrinsic toxicity, MeHg also antagonizes the beneficial effects of the omega-3 fatty acids (44, 119, 146). Because intake of both substances arises from the same food source, this suggests that the risk-benefit analysis for either the omega-3s or MeHg will depend on an understanding of this complex interaction.

Some animal studies suggest that micronutrients that are normally found in high levels in seafood, such as Se and vitamin E, may protect against Hg toxicity without specifically modulating MeHg absorption or excretion (55). For Se,

differences across studies in the forms of Se and Hg, and the route and duration of exposure make interpretation difficult. Although there is some evidence showing protection against inorganic Hg toxicity by selenite, there is almost no evidence showing protection against MeHg toxicity by the organic Se compounds, such as selenomethione or selenocysteine, that are the forms of Se commonly found in the human diet. There is no human data that support a protective role for Se with respect to Hg neurotoxicity. For vitamin E, there is a suggestion that its antioxidant properties may protect against some of the adverse effects of MeHg (147, 148). However, there are few *in vivo* studies, and no epidemiological studies have addressed vitamin E intake.

RISK ASSESSMENT FOR MeHg

The risk assessment process for chemicals in foods is based on hazard identification, exposure assessment, dose-response evaluation, and risk characterization. The most commonly used paradigms for risk assessment are those reflecting the processes developed by the National Academy of Sciences/National Research Council (NAS/NRC) in the United States (149) and a similar process used internationally by the Joint Expert Committee on Food Additives and Contaminants (JECFA) under the Food and Agriculture Organization and the WHO (150). The NAS/NRC provided recommendations on MeHg in 2000, and JECFA continues to evaluate MeHg after their evaluation published in WHO Food Additives Series Number 52 (151).

In the risk assessment for MeHg, both NAS\NRC and JECFA used a benchmark dose approach based on a predetermined change in response rate of an adverse effect. Both used the benchmark dose lower limit (BMDL), which is the statistical lower confidence limit on the dose. Because these two major risk assessments recommend different intake levels [0.1 μ g kg-body-weight (bw)⁻¹ d⁻¹ and 0.23 μ g kgbw⁻¹ d⁻¹, respectively], here we examine the choices throughout the process that lead to these differences (Table 1):

- i. Choice of study. Currently both rely on neurodevelopment effects of MeHg as the adverse health effect used in their respective risk assessments. The NAS/NRC based their analyses on the Faroes Islands study as the primary source of epidemiological data and relied on the studies from New Zealand (87) and the Seychelles as secondary sources and derived a BMDL, based on cord blood of 58 μ g L⁻¹. The JECFA excluded the New Zealand study and, basing their BMDL calculation only on the Faroe Islands and the Seychelles studies, derived a BMDL of 12 μ g g⁻¹ in maternal hair.
- ii. Biomarker of exposure. The NAS/NRC based their analyses on cord blood, and the JECFA used maternal hair. Because some of the critical studies for these risk assessments measured only one of these biomarkers converting between cord blood and maternal hair concentration (or vice versa)

involves uncertainty. Furthermore, as the most critical period(s) of gestation for the neurodevelopmental toxicity of MeHg are not yet known, it is not clear which lengths of maternal hair are most appropriate to measure

iii. Uncertainty factor. This factor accounts for adequacy of the pivotal study, interspecies extrapolation, interindividual variability in humans, adequacy of the overall data base, and the nature of the toxicity. These are not ''safety factors'' in that they are intended to factor in quantitatively to address areas of uncertainty in the risk assessment rather than provide ''safety'' *per se*. The magnitude of the uncertainty factors is intended as an estimate of the influence of these uncertainties, rather than the application of an arbitrary layer of safety. In the assessment conducted by the NAS/NRC committee, a composite uncertainty factor of 10 was used to account for variability and uncertainty in toxicokinetics and toxicodynamic, as well as database insufficiency for endpoints possibly more sensitive than neurodevelopmental (e.g., cardiovascular endpoints). The JECFA used an overall uncertainty factor of 6.4 to address variability in both toxicokinetics and toxicodynamics. The toxicokinetic portion accounts for a factor of 3.2 based on a generalized estimate of intraspecies toxicokinetic variability (152). The toxicodynamic portion likewise accounts for a factor of 2.0 based on a generalized estimate of interindividual variability in response.

The starting points for derivation of their respective recommended intakes differ both with respect to the actual values and the approaches taken. The JECFA Committee estimated that a steady-state intake of 1.5 μ g kgbw⁻¹ d⁻¹ would be an exposure that would have no appreciable adverse effects on children, in contrast to the NAS/NRC determination of a BMDL of 1.0 μ g kgbw⁻¹ d⁻¹, which is an effect level. However, neither of these assessments reflected bioconcentration of MeHg across the placental circulation from the mother to the fetus (61). This bioconcentration and its population variability suggests that the full toxicokinetic variability is significantly larger (60, 153) than previously estimated (38, 151, 154).

The NAS/NRC used cord blood mercury for their BMDL of 58 μ g L⁻¹, as did the US Environmental Protection Agency in 2001. However, the subsequent increased recognition that cord blood mercury is, on average, 60% to 70% higher in Hg than maternal blood, coupled to the coefficient of variation around the mid-point of 1.7 described by Stern and Smith (61) as 0.56 with a 95th percentile of 3.4, supports the use of a blood mercury concentration in the mid-30 μ g L⁻¹ range to recognize this fetal-maternal blood mercury difference (152, 155). By contrast, assessments based on association of maternal hair Hg with adverse neurobehavioral outcomes in the child after *in utero* exposures to MeHg need no such adjustment for MeHg concentration.

PANEL CONSENSUS CONCLUSIONS

Methylmercury is a potent toxicant, bioaccumulated and concentrated through the aquatic food chain, placing at risk humans who consume high-end aquatic predators or for whom fish is a dietary mainstay. Elevated levels of MeHg exposure occur worldwide and are not restricted to isolated populations. Rather, exposure to MeHg at levels above those that can be considered clearly safe and without risk of adverse effect occur throughout the globe and across the socioeconomic spectrum.

Hair and blood Hg concentrations (including cord blood Hg concentrations) are valid biomarkers of MeHg exposure. Each conveys somewhat different information on exposure. The most useful picture of exposure is likely to be obtained by data from both biomarkers, along with specific dietary information on fish

consumption and other dietary data. Urinary Hg concentration is a biomarker of inorganic Hg. More research characterizing the relations between toenail Hg, hair Hg, blood Hg, and urinary Hg, and the relations between MeHg and inorganic Hg should be considered a priority. Single-strand and, particularly, continuous single-strand hair analysis of Hg concentration should be pursued as the best method for elucidating dynamic changes in MeHg exposure. This is particularly relevant for studies of the effect of *in utero* exposure to MeHg to assess the significance of bolus doses.

Total fish consumption without differentiating fish species is not necessarily a dependable metric for estimating MeHg exposure. To be useful for such purposes, valid data on the MeHg concentration of each species, as well as the frequency and the amount of consumption for each species must be included.

There is sufficient evidence to state that MeHg is a developmental neurotoxin, and developmental or fetal neurotoxicity has constituted the basis for risk assessments and public health policies. Although uncertainties in the risk assessment for the neurodevelopmental effects of MeHg remain, there is sufficient evidence to warrant a public health response based on prudent selection of fish species in the diet. Development of a formal case description and diagnostic criteria for the clinical effects of MeHg observed in some adults and older children with moderately elevated MeHg exposure should be a priority for clinicians involved with MeHg research.

Current studies suggest that present levels of exposure to MeHg have the potential to result in an elevated risk of cardiovascular disease to a significant fraction of the population. However, additional studies in other populations would clarify this picture. Quantitative dose-response assessment of existing studies should be undertaken. The potential effect of MeHg on the immune system should be investigated with respect to adverse effects on immune response, as well as with respect to individual sensitivities to MeHg, potentially including autoimmune responses.

To date, it has been possible to statistically separate the neurodevelopmental effects of MeHg and PCBs in key studies where both exposures occur in the fish-consuming population. However, knowledge of the mechanisms and interactions of PCBs and other halogenated organics with MeHg is an important missing piece in understanding the overall risk for fish consumption. Research into the potential interactions of inorganic Hg and MeHg should be considered a priority. Although the possible interactions between Se and MeHg are a fruitful area for further research, there is currently no clear evidence that dietary Se can modulate the toxicity of MeHg.

Because the intake of both omega-3 fatty acids and MeHg occurs from fish consumption and because MeHg appears to antagonize the beneficial effects of the omega-3s as well as exerting its own intrinsic toxicity, a proper assessment of risks and benefits for the combination of the two must address their complex interaction. Currently, there are insufficient data on this interaction to describe a coherent picture. Despite the lack of a clear picture of the interaction of the omega-3 fatty acids and MeHg, there are fish with high levels of omega-3s and relatively low levels of MeHg. Consumption of fish with low levels of MeHg and organic contaminants constitute a ''winwin'' situation and should be encouraged regardless of the underlying nature of the omega-3-MeHg interaction.

To preserve human health, all efforts need to be made to reduce and eliminate sources of exposure, through regulation and dissemination of information. In addition to documenting the multiple health hazards associated with exposure to MeHg throughout the lifespan, research needs to focus on identifying factors that influence the uptake and the toxicity of MeHg and

on examining the potential benefits of different fish species. These studies will provide information on maximizing nutritional intake from consumption and minimizing risk from exposure to MeHg.

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Aquatic Resources 2016 Baseline Study

Stibnite Gold Project Midas Gold Idaho, Inc.

April 2017

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SECTION 1 INTRODUCTION

1.1 Purpose of Study

The aquatic resources baseline study is conducted to characterize existing aquatic habitat and biological conditions prior to the start of proposed mining operations at the Stibnite Gold Project in central Idaho. The study describes the existing aquatic resources in the project study area, and it will be used to support the United States Forest Service (USFS) Environmental Impact Statement (EIS) for the Stibnite Gold Project.

1.2 Background

Figure 1-1 shows the location of the Stibnite Gold Project. The project is in the Stibnite-Yellow Pine Mining District in central Idaho, near the village of Yellow Pine. Located in Valley County, the District is characterized by historic mining activities and unpatented (federal land) and patented (private land) deposits of gold, silver, tungsten, and antimony. The Stibnite-Yellow Pine Mining District is in the Boise National Forest (BOI), but is administered by the Krassel Ranger District of the Payette National Forest (PAF).

Mining began in the District in the late 1800s, and continued on and off through 1997. Beginning in 2009, Midas Gold Idaho, Inc. (Midas Gold), began to acquire mining claims throughout the District from prior owners or by staking claims on its own behalf. With federal and the State of Idaho (State) approval, Midas Gold initiated mineral exploration activities in 2009 to better define the mineral deposit potential for the area. This effort included using the existing road network and the construction of several temporary roads to access drill sites, build drill pads, drill on both National Forest System (NFS) and private lands, and to access disturbed areas for reclamation when exploration work concludes.

The PAF Krassel Ranger District has jurisdictional authority over surface disturbance associated with mining and exploration activities on NFS land in the Stibnite-Yellow Pine Mining District. The Payette Lakes Supervisory Area of the Idaho Department of Lands has jurisdictional authority over exploration and mining-related activities on private lands within its administrative area (Idaho Administrative Procedure Act [IDAPA] 20.03.02).

In addition to the mining activities occurring in the Stibnite-Yellow Pine Mining District, future mine plans may include proposed access roads that provide transportation routes to and from the project. Proposed access roads would be on land located in the BOI, administered by the Cascade Ranger District, as well as PAF land administered by the Krassel Ranger District.

1.2.1 Project Area Description

Figure 1-2 shows the project area. The terrain within the project area consists of narrow valleys surrounded by steep mountains. Elevations along valley floors range from 6,000 to 6,600 feet above mean sea level (msl). The surrounding mountains reach elevations over 8,500 feet above msl. The main drainage basin in the project area is the East Fork of the South Fork of the Salmon River (EFSFSR).

The EFSFSR flows for 16 miles before joining Johnson Creek near the town of Yellow Pine. The project area encompasses the watersheds of tributaries of the EFSFSR, including Sugar Creek, Meadow Creek, Johnson Creek, Riordan Creek, Burntlog Creek, and Trout Creek. The project area includes Cabin Creek and Warm Lake Creek, which are tributary streams to the South Fork Salmon River (SFSR). The primary uses or activities in the area have been mineral exploration, mining, logging, and dispersed recreation.

During non-winter conditions (when roads are clear of snow), the project site can be accessed from the City of Cascade by traveling northeast on Warm Lake Road/Forest Service Road (FS) 579 (FS 579) for about 37 miles to Landmark, then north on Johnson Creek Road (FS 413) for 28 miles to Yellow Pine, and 14 miles east on Stibnite Road (FS 412) (**Figure 1-2**). The site can also be accessed from the City of McCall during non-winter conditions by traveling east on Lick Creek Road (FS 412) for 33 miles to East Fork Road (FS 412), then 16 miles to Yellow Pine, and 14 miles on Stibnite Road (FS 412).

During winter, the site can be accessed only from Cascade by traveling 24 miles northeast on Warm Lake Road (FS 579) to its intersection with South Fork Road (FS 474/674), then north on South Fork Road for 32 miles to East Fork Road (FS 412), then 16 miles east on East Fork Road to Yellow Pine, and 14 miles on Stibnite Road (FS 412).

Figure 1-1. Vicinity Map

1.3 Organization of Report

- Section 1, the introduction, explains the purpose of the baseline study and provides background information on the aquatic resources study area, the project site, and surrounding areas
- Section 2 provides an overview of the aquatic resources study area
- Section 3 summarizes the methodology used to characterize the existing aquatic resources
- Section 4 details the affected environment as it relates to aquatic resources.
- Section 5 provides the results and summary of the annual monitoring efforts
- Section 6 contains project references, a glossary, and a list of abbreviations and acronyms
- Section 7 includes the list of preparers
- Six Appendices provide the survey results, including the stream habitat survey statistical results; macroinvertebrate survey results; water temperature monitoring results and comparison to the Watershed Condition Indicators; laboratory reports for the metals testing in soils, macroinvertebrates and fish tissues; genetic sample results; and snorkel survey results

Figure 1-2. Project Area

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SECTION 2 AQUATIC RESOURCES STUDY AREA

2.1 Description of Study Area

The aquatic resources study area includes all identified aquatic resources that could potentially be affected by the Stibnite Gold Project, including locations where Project access roads could be modified or constructed. The exact extent of potential impacts on aquatic resources will depend on the selected alternative and associated features.

Figure 2-1 shows the aquatic resources study area, including the EFSFSR and its tributaries, which include Meadow Creek, East Fork Meadow Creek (EFMC), Fern Creek, Garnet Creek, Fiddle Creek, Midnight Creek, Hennessy Creek, Cane Creek, Cinnabar Creek, Sugar Creek, Tamarack Creek, and Profile Creek. It also covers tributaries to Johnson Creek, including Burntlog Creek, Trapper Creek, and Riordan Creek. Two additional control sites, tributaries to the SFSR, include Goat Creek and Fourmile Creek. Control sites are considered stream sites in which there are fewer human impacts, and sites in which the Project activities would not have any effect.

A detailed description of each stream-reach survey is provided in **Section 4**.

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Figure 2-1. Aquatic Resources Study Area

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SECTION 3 METHODOLOGY

This methodology section details the literature used to develop the aquatic resources baseline study, highlights federal regulations that will be required to advance the Stibnite Gold project, and presents the methods employed to monitor aquatic conditions (stream habitat, macroinvertebrates, metals, and fish) in the aquatic resources study area.

3.1 Literature Review

Over the years, various habitat, fish, and invertebrate studies have been conducted and multiple management, guidance, and methods documents have been prepared, contributing to the understanding of the aquatic resource study area. A bibliography of those studies and documents is provided in the 2016 Aquatic Resources Baseline Study Work Plan (MWH Americas, Inc. [MWH] 2016). The following list identifies the key documents used in the development of baseline report analyses, including field protocols and background information.

- *Streams of Idaho (303(d) Impaired 1998* (Idaho Division of Environmental Quality [IDEQ] 2002) contains the water-quality constraints on each water body in Idaho.
- *Payette National Forest Land and Resource Management Plan (LRMP)* (PAF 2003) describes management goals and objectives, aquatic resource protection methods and metrics for desired aquatic resource conditions, as well as the availability and suitability of land for resource management.
- *PACFISH/INFISH Biological Opinion (PIBO): Effectiveness Monitoring Program Seven-Year Status Report, 1998 through 2004* (Henderson et al. 2005). This report provides a protocol-level survey to evaluate the effect of land management activities on aquatic and riparian communities at multiple scales, providing consistency and repeatability with field surveys in the PAF. It describes the framework for monitoring aquatic and riparian resources throughout the Upper Columbia River Basin, and was used for the aquatic resources baseline field investigations conducted by MWH.
- *PIBO Field Protocol* (PAF 2013, 2014, 2015) provides the specific protocols for conducting the PIBO surveys in the PAF.
- *Cobble Embeddedness Field Protocol* (PAF 2007a) provides the protocol for conducting the cobble embeddedness surveys.
- *Free Matrix Field Protocol* (PAF 2007b) provides the protocol for conducting the free matrix and surface-fines surveys.
- *Core Sampling Procedure* (PAF 2007c) provides the protocol for conducting the modified McNeil core substrate surveys.
- *Stream Temperature Technical Report* (USFS 2013) provides the protocol for the underwater installation of thermographs and the monitoring of annual temperatures in rivers and streams.
- *Idaho Small Stream Ecological Assessment Framework: An Integrated Approach* (Grafe 2002) describes methods and evaluations for macroinvertebrate sampling.
- *PIBO-EMP Effectiveness monitoring for streams and riparian areas: sampling protocol for stream channel attributes* (Heitke et al 2011) provides field protocols for stream habitat, benthic macroinvertebrates, and fish surveys under the Pacific Anadromous Fish Strategy (PACFISH) Inland Fish Strategy (INFISH) Biological Opinion Effectiveness Monitoring Program (PIBO) protocols.
- *Geography and Timing of Salmonid Spawning in Idaho* (Miller et al. 2014) describes the reproductive timing for Chinook salmon, steelhead, bull trout, westslope cutthroat trout, and brook trout within the State of Idaho.
- *Underwater Methods for Study of Salmonids in the Intermountain West* (Thurow 1994) describes methods for fish counts during snorkel surveys.

3.2 Field Investigations

MWH conducted field investigations from 2012 to 2016, which included stream habitat and biological surveys at multiple sites in the aquatic resources study area, and included at least one control site (the number of control sites depended upon the survey type). Surveys were conducted in the EFSFSR, Profile Creek, Tamarack Creek, Hennessy Creek, Sugar Creek, Cane Creek, Cinnabar Creek, Midnight Creek, Fiddle Creek, Garnet Creek, Meadow Creek, EFMC, Fern Creek, and several unnamed tributaries.

Additional stream habitat surveys were conducted in Riordan Creek, Trapper Creek, and Burntlog Creek, to assess watershed conditions and overall basin health, factors that could potentially be impacted by construction of the proposed Burntlog access route alternative (HDR 2013). The stream habitat assessment also included control sites on Goat Creek and Fourmile Creek. All aquatic resources baseline survey sites were permanently marked with safety-capped rebar to support accurate location for long-term monitoring of the sites.

Methods employed for the field surveys primarily followed the PAF-modified protocol for PIBO (Henderson et al. 2005). PIBO was developed in response to monitoring needs addressed in Biological Opinions from the U.S. Fish and Wildlife Service (USFWS), for bull trout, and National Marine Fisheries Service (NMFS) for steelhead. It provides a consistent framework for monitoring aquatic and riparian resources within the range of the coverage for the Biological Opinion through PIBO. The intent of the PIBO is to help determine whether land management practices are maintaining or improving riparian and aquatic conditions, at both the landscape and watershed scales, on federal land throughout the Upper Columbia River Basin.

The following sections describe the methods and analyses for each of the aquatic surveys.

3.2.1 Stream Habitat Surveys

Stream habitat surveys provide data on the stream bed and bank conditions over time, so that changes in stream conditions can be identified. Surveys for stream habitat characteristics have been conducted annually since 2012, with data collected in early August and mid-September. Stream habitat surveys were conducted in the EFSFSR, Sugar Creek, Meadow Creek, EFMC, Cane Creek, Burntlog Creek, Trapper Creek, and Riordan Creek, with control sites at Tamarack Creek, Goat Creek, Fourmile Creek, and Profile Creek. **Figure 3-1, Table 3-1** and **Table 3-2** show the stream habitat survey locations and survey activity for 2012 through 2016.

The stream habitat survey locations were selected based on recommendations by the PAF, including several survey sites previously established by the PAF. Survey sites were added in 2014 to include streams in watersheds that may be affected by the proposed Burntlog access route alternative, and to provide additional control sites for the baseline study. One site, on Fiddle Creek, was added in 2015 to encompass more of the project area that may be directly affected by project activities.

Stream habitat surveys are subject to specific seasonal time constraints. Core sediment samples (first collected in 2013) must be completed before August 15, to avoid spawning Chinook salmon. Free matrix and cobble embeddedness surveys occur during low-flow conditions (in September or October) with an exception for surveys on lower Meadow Creek. In 2015, following the release of Chinook salmon into Meadow Creek, these surveys were cancelled because of the presence of Chinook salmon redds. Therefore, free matrix and cobble embeddedness surveys in lower Meadow Creek must occur prior to any Chinook salmon release (typically in late August). This modification was implemented in 2016.

Protocols for the stream habitat surveys have been relatively consistent each year; however, modifications were made to the protocols based on revisions by the PAF. For example: bankangle measurements and modifications to bank-stability criteria were added to the 2014 protocol; stream shade was measured per the protocol in 2012, but the PAF removed this measurement from the protocols in 2013; and large woody debris (LWD) sampling protocols changed in 2014, and then again in 2015.

The following list is a summary of specific protocols and parameters used to sample stream channel conditions, the form of organisms, and relationships between their structures (morphology) and streambed materials (substrate) at the survey sites.

• Cobble embeddedness surveys (methods based on PAF 2007a): Surveys were conducted at five sites in the historic mining area; two control sites on Tamarack and Profile Creeks, and three sites for the Burntlog access route (**Figure 3-1**). Embeddedness measures the degree to which finer particles cover the larger particles.

Parameters measured:

- o Substrate size (diameter)
- o Depth of embeddedness in the streambed
- Free matrix surveys (methods based on PAF 2007b): Surveys were conducted at 17 existing sites in the Stibnite Gold project area: two control sites on Tamarack and Profile Creeks, and eight sites for the proposed Burntlog access route, including two control sites on Goat and Fourmile Creeks (**Figure 3-1**).

Parameters measured:

- o Fine sediment particles (surface fines) which measures the amount of fine sediment, less than 6.33 millimeters (mm) on the substrate surface
- o Free cobbles (non-embedded material) and embedded cobbles, which measures substrate size and count of loose cobbles compared with substrate size and count of embedded cobbles
- Modified McNeil core and suspended sediment sampling (methods based on PAF 2007c): McNeil core sampling was conducted at one location in the EFSFSR (**Figure 3-1**).

Parameters measured:

- o Spawning sediment depth fines (proportion of the subsurface sediments made up of fine materials) and overall sample particle size distribution based on the water displacement method
- o Suspended sediment concentration
- Channel morphology and condition evaluation assessed via PIBO surveys (PAF 2013, 2014, 2015): a total of 22 PIBO locations have been established, 15 of which are located within the historic mining area (excluding sites for the Burntlog access route) (**Figure 3- 1**). Because there were extensive avalanches in 2014, PIBO surveys were conducted at two sites in the EFSFSR to determine whether these avalanches changed channel morphology. PIBO surveys were also conducted at two control sites on Tamarack and Profile Creeks, and five sites for the Burntlog access route (including two control sites on Goat and Fourmile Creeks) (**Figure 3-1**).

Parameters measured:

- o Bankfull width
- o Wetted width
- o Wetted depth
- o Bankfull depth
- o Bank stability
- o Bank angle
- o Bed sediment size
- o Reach gradient
- o Pool dimensions
- o Large woody debris
- o Detailed survey site drawing (or verification of the previous year's drawing for accuracy)

• Water temperature monitoring (via Tidbit v2 Water Temperature Data Loggers, model UTBI-001, with methods based on USFS 2013): a total of nine data loggers have been installed in the historically mined Stibnite area, and in creeks that may be affected by the Burntlog access route (**Figure 3-1**). Seven sites were originally installed in 2013, eight additional installations occurred in 2014, and one additional site was added in 2015.

Parameters measured:

- o Thermographs were installed and the temperature was monitored at 15-minute intervals, in accordance with the USFS Rocky Mountain Research Station (RMRS) protocol for continuous stream water temperature monitoring (USFS 2013).[1](#page-174-0) Several data losses or gaps have occurred over the years due to malfunctioning field download equipment or high spring-water runoff flows. Specific download events resulting in loss of data include:
	- MWH-001 (EFSFSR) Installed in 2013, but could not download site due to frozen stream during the 2014 download event. Data loss spans from install date to September 2014.
	- MWH-005 (EFSFSR) No data has been successfully downloaded due to multiple high spring flows, resulting in the loss of the data logger and a field data shuttle download malfunction. The last reinstallation was completed in August 2016, and planned download is set for the first quarter of 2017, prior to peak spring runoff.
	- MWH-008 (Sugar Creek) Installed in September 2013, and lost in the spring of 2016 due to high runoff flows. Data gap spans from November 2015 to August 2016, and it was reinstalled in September 2016.
	- MWH-051 (Burntlog Creek) Installed in September 2014 and lost in the spring of 2016 due to high runoff flows. Data gap spans from November 2015 to August 2016, and it was reinstalled in September 2016.
	- MWH-055 (Riordan Creek) Installed in September 2015, but no data has been successfully collected to date. This site is only downloaded on an annual basis due to its remoteness. The sampling reach is characterized predominately by a silt / sand, and contains no sizable boulders for permanent epoxy installation. The data logger was attached to one of the few trees located on the bank, using parachute cord and weighted down by fishing weights. The data logger did not remain attached and it was lost. Reinstallation of the temperature data logger will be completed in 2017, and will include stainless steel cabling.
	- MWH-057 (Goat Creek) Installed in September 2014 and lost in the spring of 2016 due to high runoff flows. Data gap spans from November 2015 to August 2016, and it was reinstalled in September 2016.

 \overline{a}

¹ Water temperature data were originally downloaded annually, but the download schedule is now adjusted to two to three times a year to minimize the potential for data losses. Several data losses or gaps occurred over the years, due to malfunctioning downloads from the field data shuttle or loss of the data logger from high spring runoff flows.

Because the water temperature monitoring will continue for multiple years, the previous loss of data will not affect the ability to acquire an ample period of record to develop a stream temperature baseline. To supplement temperature data from the project installed thermographs, stream temperature data was downloaded and summarized from five U.S. Geological Survey (USGS) gauge sites within the aquatic resource study area (**Figure 3-1**).

Figure 3-1. Habitat Survey Locations and Survey Activity, 2012 – 2016

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Table 3-1. Cobble Embeddedness, Free Matrix, and PIBO Surveys Conducted from 2012-2016

Table 3-1. Cobble Embeddedness, Free Matrix, and PIBO Surveys Conducted from 2012-2016 *(continued)*

Note: Cobble Embeddedness and Free Matrix surveys were conducted by PAF in 2012.

Key:

EFMC = East Fork Meadow Creek

EFSFSR = East Fork of the South Fork of the Salmon River

PIBO = Pacific Anadromous Fish Strategy/Inland Fish Strategy Biological Opinion

SFSR = South Fork of the Salmon River

Table 3-2. McNeil Core Sediment and Water Temperature Surveys Conducted from 2013-2016

Key:

EFMC = East Fork Meadow Creek

EFSFSR = East Fork of the South Fork of the Salmon River

SFSR = South Fork of the Salmon River

Stream Substrate Quality Analysis 3.2.1.1

Three distinct sampling techniques were used for sediment monitoring. McNeil core sampling is the oldest of the three sampling methods used for evaluating streambed sediments. Core sampling also provides the most comprehensive assessment of substrate composition, but is also the most labor-intensive and is difficult to implement in smaller streams. Cobble embeddedness measurements provide a method to assess interstitial space in streambed cobbles available to small fish and macrointervertebrates. Free matrix is the most recently developed of the three methodologies and is considered a surrogate for cobble embeddedness. Free matrix protocols also provide measurements of surface fines whereas cobble embeddedness does not. Both cobble embeddedness and free matrix are better suited for small streams and are less labor-intensive compared to core sampling making them easier to establish across the study area (Nelson and Burns 2005).

MWH measured several parameters, including cobble embeddedness, surface fines, free matrix, and depth fines (core sampling) to evaluate substrate quality. The following statistical calculations were performed using Microsoft Excel:

- **Mean (µ)**: The sum of a collection of numbers divided by the total amount of numbers in the collection. While the arithmetic mean is often used to report [central tendencies,](http://en.wikipedia.org/wiki/Central_tendency) it is not a [robust statistic,](http://en.wikipedia.org/wiki/Robust_statistic) meaning that it is greatly influenced by [outliers](http://en.wikipedia.org/wiki/Outlier) (values that are much larger or smaller than most of the values).
- **Standard Deviation (σ)**: Standard deviation quantifies scatter within a dataset; how much the values vary from one another. It is expressed in the same units as the data.
- **Standard Error of the Mean (SEM)**: The standard error of the mean (SEM) quantifies precision of the mean. It is a measure of how far along a distribution the sample mean is likely to be from the true population mean. It is expressed in the same units as the data. A smaller value of SEM indicates that the sample mean is close to the population mean. A smaller SEM indicates more accurate data.
- **Confidence Interval (CI)**: Confidence intervals consist of a range of values (interval) that act as good estimates of the unknown [population parameter.](http://en.wikipedia.org/wiki/Population_parameter) Confidence intervals are typically stated at the 95 percent confidence level and are shown as a value range (i.e., \pm 5 percent of the mean).
- **Coefficient of Variation (CV)**: Coefficient of variation (CV) is a measure of the dispersion of data points in a data series around the mean. CV is the ratio of the standard deviation to the mean, and it is a useful statistic for comparing the degree of variation from one monitoring site to another or between different years, even if the means are drastically different from each other.
- **Analysis of Variance (ANOVA)**: Comparison of means among sites and years, for appropriate variables.

For all habitat data box and whisker plots were developed by year for each site and subwatershed (see **Appendix 1**). Box and whisker plots display variation in samples of a statistical population without making any assumptions of the underlying statistical distribution, and they are used for quality control. A box and whisker plot is a quick way of graphically examining one or more sets of data, and is useful for comparing distributions among multiple datasets.

Cobble Embeddedness Evaluation 3.2.1.Z

Cobble embeddedness provides an indication of the amount of interstitial space in streambed cobbles that is available to small fish and macroinvertebrates. Studies have shown that embedded substrate material affects both fish and the aquatic invertebrates that are commonly eaten by fish. For example, *Epeorus albertae,* a species of mayfly, is intolerant to the addition of fine sediment in streambeds. Juvenile steelhead have been shown to have impaired growth as embeddedness increased because of decreased prey availability and increased energy-wasting activity (Bjornn et al. 1977, Suttle et al. 2004).

Cobble embeddedness surveys are conducted annually at 10 sites within the aquatic resources study area, including sites historically monitored by the PAF. Several new sites within the Burntlog and SFSR watersheds were included, to provide data relevant to possible future access road alternatives (**Table 3-3**). Cobble embeddedness is the specific measurement of both the depth of embeddedness and total diameter for substrate for which the greatest diameter is

between 45 mm and 300 mm. This measurement provides an estimate of overall percent embeddedness of the sampled reach.

Climatic and hydrologic variability occurs in natural systems, so collecting a robust dataset is not always practical to assess baseline conditions. Since year-to-year substrate conditions can be highly variable, tiered WCIs have been developed by the PAF to support management decisions regarding substrate information for granitic and non-granitic basins (Nelson and Burns 2005, Nelson and Burns 2007, Zurstadt et al. 2016a). The tiered WCI system allows for a single mean assessment to be applied when less than five years of data are available. Currently, no more than four data points exist for a single monitoring site, so WCIs based on a single year are currently being applied in this assessment. WCIs are included in tables and figures in this baseline report and are discussed for reference (Nelson and Burns 2005, Nelson and Burns 2007, Zurstadt et al. 2016a).

The PAF uses different WCIs to assess habitat health in granitic versus non-granitic watersheds (Nelson et al. 2006, Zurstadt et al. 2016a). Each sample site was identified as either a granitic or a non-granitic geologic type based on the spatial analysis of a PAF geology Geographic Information System (GIS) file. Areas designated with intrusive lithology were considered granitic, and areas with extrusive or metamorphic lithology were considered non-granitic. Nongranitic thresholds were adopted from the Big Creek Restoration and Access Management Plan Environmental Assessment (USFS 2016) and applied to this study (**Figure 3-2**).

Table 3-3 shows the substrate watershed condition indicators for granitic and non-granitic watersheds of the SFSR subbasin, for assessing cobble embeddedness.

 \mathbb{Z}^* $\left($ $\left\langle \right\rangle$ \sum Sugar Creek subwatershed **MWH-009 MWH-010 MWH-008 Hest End Creek** Midnight Creek \leftarrow MWH-059 \sim MWH-062 arnel **Headwaters East Fork** South Fork Salmon River subwatershed **MWH-049 MWH-014** /MWH-013 **MWH-006 MWH-060** $MWH-044$ **MWH-047** East Fork of the South
Fork of the Salmon Rive Upper Indian Creek subwatershed Meadow Creek Lookout Ra

Figure 3-2. Geologic Types in the Project Area

Table 3-3. Substrate Watershed Condition Indicators for Cobble Embeddedness in Granitic and Non-Granitic Watersheds of the South Fork Salmon River Subbasin

Source: Nelson and Burns, 2005, Nelson et al. 2006, Zurstadt et al. 2016a

Key:

 $\frac{1}{2}$ = greater than

≤ = less than or equal to

% = percent

Free Matrix and Surface Fines Evaluation 3.2.1.3

Free matrix measurements monitor streambed interstitial conditions and can be used as a surrogate for cobble embeddedness. Contrary to cobble embeddedness, free matrix samples at the reach-level follows a simpler sampling protocol, likely resulting in less chance of measurement error by the field observer (Nelson and Burns 2005). Additionally, the free matrix protocol includes an estimate of surface fines. Free matrix and surface fines data have been collected annually at 27 sites within the aquatic resources study area, from 2013 through 2016 (**Table 3-1**).

When measured correctly, free matrix measurements are correlated to, but different than the cobble embeddedness measurements. Free matrix measurements are simply counts of the free and embedded substrate material for which the greatest diameter is between 45 mm and 300 mm. Surface fines are defined as substrate material less than 6.33 mm in greatest dimension (described as diameter by the PAF). Pebble and gravel-size material—which are greater than or equal to 6.33 mm, and less than 45 mm in greatest diameter—are not measured within the free matrix protocol. Nelson and Burns (2007), state that surface fines represent a poor indicator of salmonid habitat condition, and they "caution that visually determining the frequency of fine particles is problematic."

Cobble embeddedness sample events are limited to very specific habitat strata, they require detailed measurements, and are subject to higher probability of sampling error (Nelson and Burns 2005). By contrast, free matrix measurements cover the stream reach at a variable length, and is a simpler sampling protocol. Free matrix and cobble embeddedness data are collected in a double sampling protocol, to be able to estimate cobble embeddedness for reaches where it is not measured. In addition, free matrix data can serve as a quality-control mechanism for cobble embeddedness measurements, since measurement error is less likely when using free matrix. Cobble embeddedness and free matrix data should correlate for the same site; however, there is no correlation expected with surface fines.

Since year-to-year substrate conditions can be highly variable, tiered WCIs have been developed by the PAF to support management decisions regarding substrate information in both granitic and non-granitic basins (Nelson and Burns 2005, Nelson and Burns 2007, Zurstadt et al. 2016a). Therefore, WCIs are shown for both single sampling events and 5-year mean values for each basin type. **Table 3-4** lists substrate WCIs for granitic and non-granitic areas of the SFSR subbasin (free matrix and surface fines). These WCIs address the proportion of substrate samples

with particles featuring the greatest diameters between 45 mm and 300 mm that are free (i.e., not embedded) in the streambed (free matrix), and the proportion of sampled areas covered by sediment less than 6.33 mm in greatest diameter (surface fines).

Source: Nelson and Burns 2005, Nelson et al. 2006, Zurstadt et al. 2016a

Key:

> = greater than ≥ = greater than or equal to

 \leq = less than

≤ = less than or equal to

 $% =$ percent

Modified McNeil Core Sampling (Depth Fines) Evaluation 3.2.1.4

The modified McNeil core substrate sampling (core sampling) measures the proportion of the subsurface stream sediment consisting of fine materials, and is also referred to as depth fines. Percentage of depth fines is an indicator of spawning substrate quality. The core sampling provides an established method for evaluating sediment conditions and trends in streambed sediments, and offers a complete assessment of substrate composition.

Core samples from the EFSFSR at site MWH-033 have been collected annually since 2013 (**Table 3-2)**. The core sampling site is approximately one-third of the way from the bottom of PIBO survey delineated reach. It is also upstream from the free matrix/surface fine survey location for MWH-033, since free matrix/surface fines are conducted starting at the downstream end of PIBO surveys. Core sampling occurs in five, 8-by-8 foot grid subsites following the PAF protocol (PAF 2007c).

Core sampling will continue until an adequate sample size is attained. As described above, since year-to-year substrate conditions can be highly variable, tiered WCIs have been developed by the PAF to support management decisions regarding substrate information in granitic and nongranitic basins (Nelson and Burns 2005, Nelson and Burns 2007, Zurstadt et al. 2016a). At this point in the baseline study, data has been collected for four years at the core sample site. All WCI indicators are included in tables, charts, and in discussion for reference.

Core sampling provides the percentage of depth fines of less than 6.33 mm, and the percentage of depth fines less than 0.85 mm. Core samples are only taken in the EFSFSR at site MWH-033, therefore a comparison of means of depth fines is not applicable. Different WCIs are used to assess habitat condition in granitic and non-granitic watersheds (**Table 3-5**) for depth fines less

than 6.33 mm, but the WCIs are based on a 5-year mean, with no values provided for single-year samples. For depth fines less than 0.85 mm, Jensen et al. (2009) recommends a 10 percent threshold regardless of geologic substrate. In general, a lower proportion of depth fines indicates higher-quality spawning substrate (Jensen et al. 2009).

Table 3-5. Substrate Watershed Condition Indicators and Thresholds for Depth Fines (Core Sampling) in Granitic and Non-Granitic Areas of the South Fork of the Salmon River Subbasin

Source: Nelson and Burns 2005, Zurstadt et al. 2016a

Note: WCIs for Depth Fines are based on a 5-year mean only.

*10% thresholds are not part of WCI, rather they are a threshold suggested for long-term trend data for fines <0.85 mm at core sites, based on Jensen et al. 2009 research. Key:

Key:

> = greater than

 \leq = less than

≤ = less than or equal to

% = percent

mm = millimeters

Channel Morphology and Condition Evaluation (PIBO Surveys) 3.2.1.5

Stream habitat parameters are directly related to fish habitat quality. As shown in **Table 3-1**, PIBO habitat data have been collected at multiple sites since 2012, most of which have been surveyed twice to date. The PIBO surveys are being conducted on a rolling 5-year monitoring schedule (each site is surveyed once every five years), unless a major event such as a landslide, avalanche, or flood occurs. Currently, each PIBO site has data for two years, except for MWH-059 (EFSFSR), MWH-060 (Meadow Creek), MWH-061 (Profile Creek), and MWH-062 (Fiddle Creek). MWH-059, 060 and 061 were established in 2014, and MWH-062 was established in 2015. The PAF advised not using WCI indicators yet for PIBO data comparisons (Zurstadt 2014).

Several parameters are directly influenced by flows or by sampling methodology. Therefore, differences in the data between years does not necessarily indicate that habitat has changed at a site. Parameters for which protocols have changed since 2012, or parameters that are directly affected by field measurements being recorded at different flows include:

- Reach Average Wetted Width-to-Wetted Depth (WW:WD) Ratio and Pool Wetted Width-to-Wetted Max Depth Ratio: These parameters are directly related to the stage (water level) of the stream at the time of sampling. Observed stream stage is dependent on flow, and thus on annual and daily precipitation, surface/groundwater interactions, and the time of year a stream is sampled (i.e., baseflow vs. early summer/spring runoff conditions). As stream stage increases, stream depth as well as wetted width increases; however, depth and wetted width may not change at the same rate with increasing flow, leading to variability in their ratio
- Bank angle measurements were included in the PIBO protocol in 2014
- Bank stability criteria were altered in the 2014 surveys, but the data are comparable
- LWD protocols were modified by the PAF in 2014 and 2015 (PAF 2014 and 2015). **Table 3-6** details the modifications to the LWD protocol between survey years, and Zurstadt et al. (2016b) provides the justification for these changes. While there is an established WCI for LWD, there is flexibility based on local conditions, such as riparian vegetation type/community type, stream channel width and type (PAF 2003). WCI thresholds and descriptions are detailed in **Table 3-7**
- Pools criteria in the PIBO protocol is dependent on flows and water level during the sampling season, so the same unit may not be counted as a pool each year. For example, wetted channel width is directly related to the stream stage at the time of sampling. Therefore, during lower flows, the same "pool" may not meet the wetted width criteria described above. Pools are defined based on the following criteria:
	- o Depressions in the streambed that are concave in profile, both laterally and longitudinally
	- o Bounded by a head crest (upstream break in streambed slope) and a tail crest (downstream break in streambed slope)
	- o Only consider main channel pools where the thalweg (i.e., middle of the deepest part of the channel of a river or other stream) runs through the pool, and not backwater pools
	- o Span at least half of the wetted channel width at any location within the pool
	- o Maximum depth is at least 1.5 times the pool tail crest depth

Note:

1The stem of all large woody pieces must extend below bankfull elevation to be counted.

Key:

LWD = large woody debris m = meter

Table 3-7. Watershed Condition Indicators for Large Woody Debris (LWD)

Source: PAF 2003

Water Temperature Analysis 3.Z.1.6

Thermograph installation sites were chosen by the USFS to record water temperatures (**Table 3-2**). Some locations coincide with past PAF water temperature monitoring sites which provides a longer-term record for comparison and analyses. Stream temperatures are recorded at 15 minute intervals by the data loggers, which were attached to instream boulders using epoxy and PVC canisters, as detailed in the USFS Rocky Mountain Research Station protocol (USFS 2013). One exception to this installation method is MWH-055 which was attached to a submerged log due to the lack of boulders or large rocks in the streambed.

Data from the thermograph data loggers were originally downloaded on an annual basis, but more frequent (two to three times per year) downloads were completed to limit data loss due to data shuttle download malfunctions or high spring runoff events which has resulted in the loss of data loggers.

To supplement the aquatic resources study area thermograph data, data from five USGS stream gauges have been incorporated into the baseline study analysis.

3.2.2 Macroinvertebrate Surveys

As described in detail in **Appendix 2**, macroinvertebrate samples were collected at 11 sites (located in the EFSFSR, Meadow Creek, Sugar Creek and Tamarack Creek) during the summers of 2012 through 2014, and in 2016, following the PIBO protocol (Heitke et al. 2011). The macroinvertebrate survey sites are shown in **Figure 3-3**.

Field Methods 3.2.2.1

Macroinvertebrates were collected with D-nets using a targeted composite method. As detailed in the PIBO protocols, two individual samples were taken at random locations in four consecutive riffle/run habitat units at each sampling site, for a total of eight individual samples per site. These eight samples were then composited in the field into a single benthic macroinvertebrate sample per site and preserved in 95 percent ethanol.

Laboratory Methods 3.2.2.2

Composite macroinvertebrate samples were transported to the EcoAnalysts, Inc. laboratory in Moscow, Idaho for sorting, subsampling, and taxonomic (e.g., family, genus, species) identifications. A subsample of 500 organisms was randomly sorted out of each composite sample. All organisms in the subsample were then identified to the lowest practical level, generally genus or species, and enumerated. If, for some reason, an individual could not be identified to the lowest practical target resolution (e.g., due to poor condition or damage, being immature, or for some other reason being indeterminate), then it was identified to the lowest taxonomic resolution possible for that individual specimen.

All individuals not identified to target resolution were noted as being unique or non-unique taxa, relative to the other taxa in the sample, so that non-unique taxa could be excluded from taxa richness (total number of identifiably distinct organisms in a sample) calculations. Excluding non-unique taxa from richness metrics avoids erroneously inflating taxa richness values.

The Idaho Stream Macroinvertebrate Index (SMI) was calculated for all sites and sampling years (Grafe 2002). The SMI was developed by the IDEQ and is extensively used to evaluate stream conditions throughout Idaho. The SMI metrics were selected from a larger list of candidate

metrics based on how well they predicted macroinvertebrate and habitat/water quality relationships. Candidate metrics fall into one of several categories that include:

- Richness
- Evenness/diversity
- Relative abundances
- Functional feeding groups
- Habit/behavior
- Pollution tolerance

Because taxa composition and natural habitat conditions vary among ecoregion types, the Idaho SMI has been calibrated for different Idaho regions based on the reference conditions used to develop the regional metric scoring formulas. The Stibnite Gold Project is located within the Idaho SMI Central and Southern Mountains region and the scoring formulas for that region were used in this study. It should also be noted that non-unique taxa, as previously defined, were excluded from all taxa richness metric calculations.

The following SMI metrics were evaluated in this study and are further described in **Appendix 2**:

- Total Taxa Richness Total number of identifiably distinct taxa in a sample
- Ephemeroptera Taxa Richness Total number of identifiably distinct taxa in the insect order Ephemeroptera (mayflies)
- Plecoptera Taxa Richness Total number of identifiably distinct taxa in the insect order Plecoptera (stoneflies)
- Percent Plecoptera Relative abundance of stoneflies, expressed as a percent of the total number of individuals in the sample
- Trichoptera Taxa Richness Total number of identifiably distinct taxa in the insect order Trichoptera (caddisflies)
- Hilsenhoff Biotic Index (HBI) Index of community tolerance to organic pollution
- Percent 5 Dominant Taxa Percentage of individuals in a sample comprised of the five most abundant taxa
- Scraper Taxa Richness Total number of distinctly identifiable taxa in a sample whose primary feeding strategy is to scrape attached periphyton (aquatic organisms, such as algae, that live attached to rocks or other surfaces) and other particulates
- Clinger Taxa Richness Total number of clinger taxa in a sample

The following five additional informative metrics were calculated and presented for all sites and sampling years (and are further described in **Appendix 2)**:

- Long-lived Taxa Richness The total number of taxa in a sample that require more than one year to complete their life cycle
- Metals Tolerance Index (MTI) McGuire's Metal Tolerance Index (MTI) is an index of community tolerance to metal contamination that ranks taxa by their sensitivity to metals
- Intolerant Taxa Richness The number of taxa in a sample with HBI values of 0 to 2
- Percent Tolerant Individuals The relative abundance of all individuals in a sample having HBI values of 8 to 10, expressed as a percentage of the total number of individuals in the sample
- Shannon-Weaver H' (log e) $-$ A community diversity index

Figure 3-3. Macroinvertebrate Survey Locations for all Survey Years (2012 – 2014 and 2016)

The functional feeding groups (FFG) represented in the insect community are a direct function of available food and habitat resources, and are often good indicators of habitat quality and stability. The following FFGs were evaluated in this study, and are further described in **Appendix 2**:

- Percent Filterers The relative abundance of all individuals in a sample whose primary feeding mechanism is to filter suspended fine particulates
- Percent Gatherers The relative abundance of all individuals in a sample whose primary feeding mechanism is to gather deposited fine particulates
- Percent Predators The relative abundance of all individuals in a sample whose primary feeding mechanism is to pierce or engulf other invertebrates
- Percent Scrapers The relative abundance of all individuals in a sample whose primary feeding strategy is to scrape attached periphyton and other particulates
- Percent Shredders The relative abundance of all individuals in a sample whose primary feeding mechanism is to shred coarse particulate organic matter (CPOM)
- Percent Unclassified The relative abundance of all individuals in a sample whose primary feeding mechanism is unknown or unclassified
- Filterer Richness The total number of distinctly identifiable taxa in a sample in which the primary feeding mechanism is to filter suspended fine particulates
- Gatherer Richness The total number of distinctly identifiable taxa in a sample in which the primary feeding mechanism is to gather deposited fine particulates
- Predator Richness: The total number of distinctly identifiable taxa in a sample in which the primary feeding mechanism is to pierce or engulf other invertebrates
- Scraper Richness The total number of distinctly identifiable taxa in a sample in which the primary feeding strategy is to scrape attached periphyton and other particulates
- Shredder Richness The total number of distinctly identifiable taxa in a sample in which the primary feeding mechanism is to shred CPOM. This is an indicator of terrestrial vegetation input
- Unclassified Richness The total number of distinctly identifiable taxa in a sample in which the primary feeding mechanism is unknown or unclassified

PIBO Observed/Expected Index 3.2.2.3

The PIBO observed/expected (O/E) index, derived from the statistical model RIVPACS (River Invertebrate Prediction and Classification System), was also used to assess biological condition of sampled sites in all surveyed years. O/E models compare the macroinvertebrate taxa observed at sites of unknown biological condition (i.e., test sites) to the assemblages expected to be found in the absence of anthropogenic stressors (Hawkins et al. 2000). O/E scores were calculated for taxa having a probability of capture greater than or equal to 0.5, to increase the precision of O/E estimates and subsequent model sensitivity to stressors.

Biological condition was subsequently assessed based on the precision of the reference site data used to develop the PIBO O/E model (mean $= 0.95$, standard deviation (SD) $= 0.16$). In general, departures from a ratio of 1.0 indicate that the taxonomic composition in a stream sample differs from that expected under less-disturbed conditions. The biological condition rating for each sampling site is a qualitative rating compared to reference sites ranging from:

- Good (i.e., comparable to reference conditions) Less than one standard deviation below the mean of reference sites
- Fair Between 1 and 2 standard deviations below the mean of reference sites
- Poor More than 2 standard deviations below the mean of reference sites

Additional details are provided in **Appendix 2**.

3.2.3 Metals Surveys

Metals within the watersheds were sampled by collecting stream sediment, macroinvertebrates and fish.

Sediment 3.2.3.1

Sediment samples were collected at 16 survey sites (**Figure 3-4**). The sites selected for testing are at, or near, those established for the Stibnite Area Site Characterization Report (URS 2000), and most overlap with the existing macroinvertebrate sites established for the baseline study. Sediment samples were collected using either a scoop or Ponar grab sampler (depending upon habitat type) and were sifted through a 250-micron sieve until a 500 milliliters (ml) container (provided by SVL Laboratories, Inc.) was approximately two-thirds full, or contained approximately 350 ml. Samples were kept refrigerated until they were shipped to the SVL Laboratory and then analyzed for a full suite of metals and metal-like anions.

The laboratory results provide data that can be used in a limited way to assess the nature and extent of metals in sampled streams' sediment, and their potential exposure to aquatic life. The data were assessed using selected ecological screening-level benchmarks for sediment as compiled by the Environmental Protection Agency (EPA) (EPA 2006). The EPA benchmarks use a synthesis of multiple screening values, which may variously include Threshold Effects Levels (TEL), Probable Effects Levels (PEL), and/or other criteria. When performing a complete risk assessment or full analysis of biological effects due to metals in the environment, it would be essential to consider the nature of these levels. The limited assessment provided in this baseline report is appropriate for the data collected. With limited data on metal content in sediment available, as well as the inherent variability in sediment substrate, it is not prudent or meaningful to provide any conjectures on 2016 temporal trends' results.

Figure 3-4. Sediment and Macroinvertebrate Survey Locations for Metals Testing in 2016

Macroinvertebrates 3.2.3.2

Macroinvertebrate samples for metals analyses were collected using a D-net at the same survey sites as the sediment metals collections (**Figure 3-4**). The samples were stored in containers provided by ALS Environmental, Inc., the laboratory conducting the analysis. Approximately 10 grams of macroinvertebrates were collected and stored in a freezer until the samples were shipped to the ALS laboratory located in Kelso, Washington.

Many metals can be highly toxic to freshwater invertebrates; however, toxicity varies across taxa, with some taxa more tolerant of metals pollution than others. Therefore, there are no standard benchmarks for metals concentration in invertebrate tissue. The most common effect of metals pollution is a change in macroinvertebrate community composition (i.e., to pollutiontolerant taxa). Due to the limitations described above, the laboratory results are primarily compared spatially between sample locations, as well as temporally with past analyses that were conducted in 1995, 1996, and 1997 (URS 2000). However, tissue concentrations for arsenic, mercury, and selenium are compared to values that are shown to have dietary toxicity for fish (i.e., toxicity to fish from eating invertebrates with high tissue concentrations of these metals). Dietary toxicity data comes from a recent review of toxicological data by NMFS (NMFS 2014). Other metals, including antimony, either did not have enough data to determine dietary toxicity or there was no evidence of dietary toxicity.

Fish Tissue 3.2.3.3

Westslope cutthroat trout and sculpin were collected during the electrofishing fish population study in 2015 from 14 locations, to test for heavy metal bioaccumulation in fish tissue. The fish kept for analyses were collected after fish were marked and after the snorkel surveys were completed. The fish were kept cool while in the field, stored in the freezer at the end of the day, and kept frozen until they were sent to the Jupiter Environmental Laboratories, Inc. for analysis.

3.2.4 Fish Surveys

Multiple fisheries' survey methods have been employed since 2012. These methods include snorkeling, videography, electrofishing, environmental deoxyribonucleic acid (eDNA) collections, and tissue collections for genetic testing. **Table 3-8** identifies all survey sites and the methods employed at these sites except for most the eDNA sites established through the Range-Wide Bull Trout eDNA Project.

Creek	MWH Site ID	2012	2013	2014	2015	2016
EFSFSR	009	D		D/N		
Sugar Creek	010	D	D	D		
EFSFSR	011	D	D	D/N	D/E	
EFSFSR	013	D	D	D		
Meadow Creek	014	D	D	D	D/E	
Meadow Creek	015	D	D	D/N		
Meadow Creek	016	D		D	D/E	
Tamarack Creek	017	D	D	D/N		
Sugar Creek	018	D	D	$\mathsf D$	D/E	
Cinnabar Creek	019	D	D	D	D/E	
Sugar Creek	020	D	D			
Cane Creek	021	D	D			
EFSRSR	022	D	D	$\mathsf D$		
Fiddle Creek	023	D	D	D		
Fiddle Creek	024	D				\vee
EFSFSR	025	D	D		D/E	
EFSFSR	026	D	D	D/N	D/E	
EFMC	027	$\mathsf D$	D	$\mathsf D$		
EFMC	028	D	D	D/N		
Sugar Creek	029	D	D/N	D		
EFSFSR	030	D	D	D/N		
EFSFSR	032		D	D/N		
EFSFSR	033		D			
Meadow Creek	034		D		D/E	
EFSFSR	044		D			
Meadow Creek	047		D/N	D	D/E	
Midnight Creek	063					D
Midnight Creek	064					\bigvee
Hennessy Creek	076					
Hennessy Creek	077					V
Fiddle Creek	$425 - 1$					D

Table 3-8. Fisheries Survey Site Locations for All Survey Years

Note:

Site MWH-014 includes the 2012 survey at site MWH-031. Site was moved a short distance upstream in 2013 to provide a survey location consistent with the stream habitat and macroinvertebrate surveys. Site MWH-076 was too shallow to either snorkel or record via a video

Site 425-1 is a designated e-DNA site in the U.S. Forest Service's Range-Wide Bull Trout e-DNA Project, and sites MWH-024, MWH-063, MWH-064, MWH-076, and MWH-077 also include e-DNA samples. Other sites defined in the Range-Wide Bull Trout eDNA Project are not included in this table.

Key:

 $D = day$ dive

 $E =$ electrofishing

EFMC = East Fork Meadow Creek

EFSFSR = East Fork of the South Fork of the Salmon River

N = night dive

V = videography

Genetics Studies 3.2.4.1

During the electrofishing surveys in 2015, described below, fin clips were taken from fish identified as either a rainbow trout or as a westslope cutthroat trout. The purpose of these fin-clip collections was for genetic sampling, to help determine morphological characteristics separating rainbow trout and cutthroat trout, and to get a better understanding of their potential hybridization and distribution.

Each genetic sample was collected by attaching a portion of a fish fin to chromatograph paper, and then allowing the sample to dry. Fork length was measured and a photograph was taken of each fish from which a genetic sample was collected. Fish collected for the heavy metal bioaccumulation study were kept whole and put into a freezer as quickly as possible. All fish were kept frozen until they reached the lab for processing.

In 2016, eDNA was collected at 22 established Range-Wide Bull Trout eDNA Project survey sites, established by the Rocky Mountain Research Station of the USFS (**Figure 3-5**). Eight additional sites were sampled, covering smaller streams not included in the larger USFS program. Samples were collected under as sterile of conditions as possible. Nitrile gloves were worn while holding the base of the filter cup. The filter cup was faced upstream, into the current, and a peristaltic pump was used to pull five liters of water through the filter. Disposable tweezers were used to remove the filter which was placed in a desiccant, and labeled with the site number, creek name, date, and global positioning system (GPS) location. The sample was then placed on ice until it was stored in a freezer. All samples were provided to the PAF, which was responsible for submitting the samples to the laboratory. Species for which the laboratory tested were Chinook salmon, westslope cutthroat trout, rainbow trout (or more appropriately, *O. mykiss* as eDNA testing cannot distinguish between subspecies at this time), bull trout, brook trout, and Pacific lamprey.

Also in 2016, fin tissues were collected from 11 trout (targeted rainbow and/or golden trout) in Meadow Creek Lake for genetic testing. The tissues were placed on chromatographic tissue sample sheets that provided individual identification labels. Photographs were taken of each fish, and the photograph number was included on the sample sheet. Once the chromatographic sheet was air dry, a single piece of white copy paper was placed on top of the dry chromatographic sheet, and then placed between two pieces of card stock. These were then stored in a single manila envelope and delivered to the PAF, which was responsible for submitting the samples to the laboratory.

Population Abundance Study 3.2.4.2

Fisheries' surveys in 2015 were conducted using a combination of electroshocking and snorkel surveys to calibrate previous and future snorkel surveys, to provide more accurate population estimates, biomass estimates, and length frequencies. Although a total of 11 sites were selected for sampling, only nine sites were surveyed due to inclement weather. These sites were selected based on the separation of all surveyed units into different strata—as defined by habitat characteristics—that would most likely influence sampling efficiency (Peterson et al. 2004, Meyer and High 2011). Previous PIBO stream habitat survey results, as well as professional judgment, were used to delineate the strata and to categorize existing snorkel sites as most representative of each. Habitat variables used for strata delineation included channel complexity, amount of LWD, stream channel gradient, substrate size, channel width, and stream bank complexity. Based on professional judgment, five strata were considered representative within the aquatic resources study area. To minimize take of Endangered Species Act (ESA)-listed fish, the survey sites selected were those that had a reduced likelihood of encountering adult and juvenile Chinook salmon and large-bodied bull trout.

The following sites by strata, each 100 meters (m) in length, were surveyed and are also shown in **Figure 3-6** and **Table 3-7:**

- Strata 1: Sites MWH-034 (Meadow Creek), MWH-026 (EFSFSR), MWH-016 (Meadow Creek), and MWH-019 (Cinnabar Creek) – complex channel with moderate LWD, moderate gradient, medium sized substrate, and small channel width. PIBO survey completed in strata at MWH-034, and MWH-016.
- Strata 2: Sites MWH-018 (Sugar Creek) and MWH-014 (Meadow Creek) simple channel, low amounts of LWD, low gradient, small substrate, and medium channel width. PIBO survey completed in strata at MWH-014.
- Strata 3: Site MWH-025 (EFSFSR) complex channel, high amounts of LWD, moderate gradient, large substrate, and moderate channel size. MWH-013 was intended for inclusion, but time constraints resulted in this site being eliminated from the survey. PIBO surveys were conducted at MWH-013 for comparable habitat information.
- Strata 4: Sites MWH-011 (EFSFSR) simple channel with low amounts of LWD, large substrate, and large channel width. No PIBO surveys are associated with this strata, although a PIBO site is located approximately 500 m downstream.
- Strata 5: Site MWH-047 (Meadow Creek) meadow channel, low amounts of LWD, low gradient, small substrate, and small channel width. MWH-027 was intended for inclusion, however time constraints resulted in this site being eliminated from the survey. PIBO survey completed in strata.

Fisheries' survey methods included mark-recapture, four-pass depletion electrofishing surveys and daytime snorkel surveys within a 100 m stretch of river confined with blocknets. Habitat was characterized for each surveyed site Surveys were conducted by two separate crews, each consisting of two USFS and three MWH biologists and technicians. Additionally, Russ Thurow and John Guzevich, both with the USFS Rocky Mountain Research Station, trained both crews because of their extensive experience in snorkel surveys, electrofishing surveys and calibration studies.

Habitat data were collected at each fish survey site to provide the overall site-specific conditions that could affect the ability to either physically collect the fish through electrofishing, or affect the ability to visually observe the fish underwater via snorkeling. Habitat parameters measured in the 100-m unit included:

- Average stream gradient
- Stream conductivity
- LWD count
- Length of stream with undercut bank
- Length of stream with overhanging vegetation
- Average width of overhanging vegetation
- Percent of unit with submerged vegetation (including tree roots) and turbulent cover
- Dominant pool-forming feature (scour, LWD, boulder)
- At 10 transects roughly equidistant throughout the unit:
	- o Wetted width
	- o Channel depth at $\frac{1}{4}$, $\frac{1}{2}$, and $\frac{3}{4}$ across the channel width
	- o Percent of substrate less than 6 mm, between 6 mm and 75 mm, between 75 mm and 150 mm, and greater than 150 mm
- Five to 10 substrate measurements at equal distances across the width

Water temperatures were recorded in 15-minute increments from the time the block nets were first installed to the time they were removed. The intent was to record water temperatures during the snorkel surveys.

Data collected in 2015 provide a calibration for the past and future snorkel surveys in order to estimate fish populations and densities, as well as to determine the best method of collection/fish observation at specific sites. To date, these data have not been fully processed, requiring further calculations to provide snorkel survey efficiencies. Results of the calibration will be presented in an addendum report, which will also include the population estimates and fish densities for the 2012 through 2015 snorkel surveys.

Snorkel Surveys 3.2.4.3

From 2012 through 2014 fish surveys were conducted via snorkeling, following the modified protocol established by the PAF, and techniques described in Thurow (1994) and O'Neil (2007). Since 2012, 31 sites, each approximately 100 m long, have been surveyed at a protocol level (**Figure 3-6** and **Table 3-8**). Both day and night dives were conducted at two sites in 2013, and eight sites in 2014 (**Table 3-8**). Night dives were conducted to observe younger age classes, which often move or use habitat differently at night. Bull trout may also be more observable at night because juveniles, which are bottom oriented, are often found in between cobbles and boulders or among woody complexes, which make them difficult to detect during daylight hours. Additionally, bull trout may increase their nocturnal movements to forage in the water column (Berge and Mavros 2001). Specific night-dive site locations were selected based on access and safety.

Fish species and size class were recorded when feasible, for each survey site, to determine fish distribution. At times, fish moved too quickly to be able to confidently identify the species, thus, notes were recorded indicating fish that were observed but not included in the overall count, which was based on species.

Additional data was recorded, including: water and air temperature, weather conditions, stream gradient, estimated average and maximum depth, and channel widths, which were recorded at 10-m intervals beginning in 2013 (only general estimates of stream widths were made in 2012). Based on protocol recommendations, water temperatures below 9 degrees Celsius (°C) (in 2012) and 10°C (2013 and later) were avoided during day dives, however there was no required temperature minimum for night dives.

The survey team photographed the upstream and downstream end of each survey site at least one time. GPS locations were recorded the first time a site was surveyed, and permanent markers were placed to document site locations.

During these survey years, some bull trout were misidentified as brook trout or as potential brook trout/bull trout hybrids, based on the color and spotting patterns of the fish observed. However, in the 2015 electrofishing survey conducted jointly with the USFS, these fish were identified as bull trout. All brook trout data from 2012 through 2014 were converted to bull trout. Additionally, following the genetics sampling, it appears that rainbow trout do not occur upstream from the Yellow Pine pit. Therefore, fish formerly identified as rainbow trout were changed to westslope cutthroat trout.

At the request of Midas and the USFS, fish observational surveys were conducted via snorkeling or videography at five eDNA survey locations (**Table 3-8**). The original intent was to snorkel six locations; however, four locations were too shallow and/or narrow to snorkel, so underwater videos were recorded, and one site was too shallow to snorkel or record by video. When snorkeling was not feasible, but videography was possible, Panasonic and GoPro digital cameras were used to film as much of the stream channel as possible, filming from downstream and moving upstream. Neither the snorkel surveys nor video surveys were protocol level, as substantial portions of the channels had to be walked around due to shallow depths, LWD aggregates, or cascades. Only the lower Fiddle Creek survey site (425-1) was approximately 100 m long, the remaining units were substantially shorter because of gradient, overgrown vegetation, instream LWD, cascades, or other conditions precluding accessibility.

Figure 3-5. Environmental DNA Survey Locations in 2016

Figure 3-6. Snorkel Survey Sites Established During All Survey Years

SECTION 4 AFFECTED ENVIRONMENT

This section summarizes the affected environment for the aquatic resources of the Stibnite Gold Project. It includes a description of the surveyed streams and the fish species found in the aquatic resources study area.

4.1 Description of Surveyed Streams

The aquatic resources study area is characterized by cool, dry summers. Most precipitation occurs as winter snow, with peak stream discharge generally occurring in late spring (May through June). The following are descriptions of important aquatic features in the aquatic resources study area:

East Fork of the South Fork of the Salmon River – Habitat type varies in the EFSFSR. In the section upstream from its Meadow Creek confluence, the EFSFSR is characterized by narrower channels with mostly moderate to high gradients (around 2 to over 6 percent), and overall mostly smaller substrate size, with sand, gravel and smaller cobble often dominating the unit with occasional boulders. Between Meadow Creek and the Yellow Pine pit, the EFSFSR widens, has moderate to high gradients (approximately 2 to 8 percent), and has larger streambed material including abundant cobble and boulders. A large cascade (with an approximately 22 percent gradient) immediately upstream from the Yellow Pine pit likely precludes fish passage. Between the Yellow Pine pit and the confluence with Sugar Creek, the EFSFSR is similar in width, gradient and substrate material as immediately upstream, but many of the larger boulders and cobble are sharp and more angular, and pools are frequently deeper. Downstream from the Sugar Creek confluence, the EFSFSR gradients vary, ranging from low gradient pools and runs to large gradient cascades. The substrate is also variable, often dependent on the channel gradient. The lower-gradient sections are often dominated by gravel and cobble, with higher-gradient units dominated by large cobble and boulders. Throughout most of the aquatic resources study area, the EFSFSR has relatively abundant riparian vegetation; the main exception is in the vicinity, both upstream and downstream, of the Yellow Pine pit. Upstream from the Meadow Creek confluence, the stream channels often have very high amounts of LWD. The 2014 avalanches, all downstream from Sugar Creek, resulted in increased LWD downstream in that section of the EFSFSR.

Tamarack Creek – This EFSFSR tributary was surveyed as a control site because, located in the boundary of the wilderness, it is considered to have minimal impacts other than trails and campsites. Only the lower section of the creek was observed and documented. The lower section of Tamarack Creek has moderate gradients with some bank sloughing. Dominant substrate is comprised of gravel and cobble, with some bedrock and boulders. Some LWD is dispersed throughout the creek.

Profile Creek – This EFSFSR tributary was surveyed as a control site, and is a moderategradient stream that runs alongside the Warren-Profile Gap Road for several miles. Only the lower section of the stream was observed and documented. Riprap from the Warren-Profile Gap Road embankment forms the western bank of Profile Creek. The stream has very little LWD and has a riparian zone (comprised mostly of willows and berries) that is narrow along the bank formed by Warren-Profile Gap Road. Due to the steep surrounding terrain, the stream is shaded throughout much of the day. The dominant substrates are gravel, cobble and boulder.

Yellow Pine Pit – During mining activities in the 1930s through the 1950s, the nearly 5-acre Yellow Pine pit was created by open pit mining, and the EFSFSR was diverted through the Bradley Tunnel to Sugar Creek (Hogen 2002). After mining ceased in 1952, the EFSFSR was allowed to flow through the abandoned mine pit. The pit currently has a maximum depth of approximately 35 feet. Diverting the EFSFSR back into the stream channel and pit created a long riffle with a high gradient that precluded fish passage into the upper watershed. This lake-like water body is inhabited at times by both fish and mammals, including large Chinook salmon, bull trout and river otters.

Sugar Creek – This EFSFSR tributary, downstream from the Yellow Pine pit, is a relatively low-gradient stream. A closed, but still locally used Forest Service road closely parallels Sugar Creek for nearly 2 miles before crossing the creek. This road may confine the movement of Sugar Creek, specifically in areas where it bounds the banks with riprap rock material. Much of Sugar Creek has large aggregates of LWD. The dominant substrates are sand, gravel and cobble.

Cane Creek – This tributary to Sugar Creek has its headwaters in the village of Yellow Pine. No roads access this small creek, and only its confluence with Sugar Creek was observed and documented. The stream is characterized by a dominant substrate of cobble and gravel with some boulders, and a low-to-moderate gradient. Cane Creek was surveyed in 2012 and 2013, but not in 2014.

Cinnabar Creek – This small tributary flows into Sugar Creek just downstream from Cane Creek. It is relatively high gradient, with abundant boulders often creating small cascades, but the dominant substrate is made up of gravel, sand and cobble.

Fiddle Creek – This is a small tributary of the EFSFSR just upstream from the Yellow Pine pit. The lower section has a very steep gradient where it flows into the EFSFSR, making it unsuitable for fish passage. Upstream of this barrier, Fiddle Creek retains a relatively high gradient in a relatively narrow channel, with an increasing number of side channels. The creek itself can be difficult to access due to thick tall-shrub overstory dominated by alder. The upper-most section of Fiddle Creek flattens in gradient, becoming a slower meandering stream where a reservoir formerly existed. Large amounts of LWD occur throughout the creek, and the dominant streambed substrate consists of boulders, large cobble and gravel.

Meadow Creek – This EFSFSR tributary is in a flat-bottomed valley surrounded by steep mountains. Elevations range from 6,300 feet above sea level to over 7,500 feet. The downstream end of the valley shows remnant effects from early mining activities, along with a large outwash feature created by a dam failure in the EFMC drainage south of Meadow Creek Mine. Meadow Creek has been modified over the years to accommodate and overcome conditions created by past mining operations, including regrading and revegetating the lower section of the creek in 2004 and 2005.

East Fork Meadow Creek – The stream characteristics of this tributary to Meadow Creek was, and is still, substantially affected by a dam that was constructed to supply hydroelectric power for milling operations. The dam was constructed in 1929 and enlarged in 1931. After an inspection by Forest Service engineers, the dam was breached in 1958 to reduce the threat of catastrophic failure. However, the dam failed in 1965 due to record snow melt and runoff rates, depositing large volumes of sediment into Meadow Creek, the EFSFSR, and the Yellow Pine pit (URS 2000). The reach of the EFMC that eroded during the dam failure is still considered unstable, and continues to deposit sediments into Meadow Creek and the EFSFSR during snow melt and storm events. Upstream from this unstable section, the EFMC has reestablished and

stabilized into a low-gradient stream flowing through a large meadow. There are few trees, but the banks have abundant grasses. The channel is very narrow, and has a very low gradient (less than 2 percent). The dominant streambed material is sand and gravel.

Garnet Creek – This EFSFSR tributary is a narrow, shallow stream not far downstream from the Meadow Creek confluence. During the summer surveys, the creek near the confluence only several inches deep. Garnet Creek cuts through a formerly burned hillside. Most of the vegetative cover is grasses, however, shrubs and trees grow alongside the banks, and woody vegetation is found within the channel.

Midnight Creek – This EFSFSR tributary has dense overhanging vegetation in the lower portion of the narrow channel, which completely restricted survey access. Midnight Creek overall is a narrow, shallow, high-gradient system.

Hennessy Creek – This EFSFSR tributary is a narrow, low-flow stream downstream from Fiddle Creek. Hennessy Creek flows in a constructed ditch alongside Stibnite Road, and it flows under the adjacent waste rock dump before dropping down a very high gradient into the EFSFSR. This high-gradient segment is a barrier to fish passage. Hennessy Creek is densely vegetated and shallow.

Riordan Creek – This tributary to Johnson Creek is overall a relatively low-gradient stream. Roughly halfway down the length of the stream is Riordan Lake. Downstream from the lake, Riordan Creek has a slightly higher gradient, particularly just before it enters Johnson Creek. A trail with bridges that are open to small off-road vehicles crosses the creek several times above and to the north side of the lake. The dominant substrates are sand and gravel.

Trapper Creek – This tributary to Johnson Creek has a high gradient near its confluence with Johnson Creek. The downstream portion consist of large boulders and cascades. The upper reaches of Trapper Creek contain abundant LWD pieces. The stream's dominant substrates are gravel and cobble.

Burntlog Creek – This tributary to Johnson Creek is a moderate-gradient stream that parallels Johnson Creek in the lowest reaches, and it occupies a steep valley floor in the upper reaches of the drainage. The upper reaches have moderate amounts of LWD from extensively burned areas, and minimal overhead canopy. The dominant substrates are sand, gravel and cobble.

Goat Creek – This tributary to the SFSR was surveyed as a control site for the Burntlog access route alternative-study sites because it is considered to have minimal human impacts other than trails and campsites. Only the lower section of the stream was observed and documented. Goat Creek is a moderate-gradient stream that occupies the valley bottom of an extensively burned area, above the South Fork Road crossing. The drainage has high amounts of LWD, both in single pieces and contained within several channel-spanning masses. The stream near the survey site was extensively burned, which affected the riparian area. The dominant substrates are sand, gravel and cobble, with some exposed bedrock and boulders.

Fourmile Creek – This tributary to the SFSR was surveyed as a control site for the Burntlog access route alternative because it is considered to have minimal human impacts other than trails and campsites. Only the lower section of the stream (below the South Fork Road crossing, and Fourmile campground) was observed and documented. Fourmile Creek is a moderate-gradient stream. The drainage has moderate amounts of LWD, both in single pieces and contained within aggregates. A large aggregate complex contributes to a split channel in the upper portions of the survey reach. The stream near the survey site has a dense riparian zone that is shaded during much of the day. The dominant substrates are gravel, cobble and boulders.

4.2 Fish Species

This section describes the status, distribution, and basic life histories for Snake River spring/summer Chinook salmon, Snake River Basin steelhead, Columbia River bull trout, and westslope cutthroat trout.

4.2.1 Snake River Spring/Summer Chinook Salmon

Snake River spring/summer Chinook salmon (*Oncorhynchus tshawytscha*) were listed as threatened in 1992 (NMFS 1992). The spring/summer Chinook salmon in the aquatic resources study area are part of the EFSFSR population. Chinook salmon occur naturally in the EFSFSR up to the Yellow Pine pit at Stibnite, but are introduced upstream from the Yellow Pine pit by the Idaho Department of Fish and Game (IDFG) with the Nez Perce Tribe. Critical habitat for Chinook salmon was designated in 1993 (58 FR 68543) and specifically defines geographic areas and essential habitat elements. Critical Habitat is designated in the EFSFSR up to approximately 1,200 feet upstream from the confluence with Sugar Creek (NMFS 1993, later revised in 1999).

In 2016, NMFS released a draft Recovery Plan for the Snake River spring/summer Chinook salmon and Snake River steelhead, which identified recovery strategies by major population group (MPG). The EFSFSR is included in the South Fork Salmon River spring/summer Chinook salmon MPG. Recovery strategies for spring/summer Chinook salmon that affect the EFSFSR watershed include the need to:

- Maintain current wilderness protection and protect pristine tributary habitat
- Provide/improve passage to and from areas with high intrinsic potential through barrier removal
- Reduce and prevent sediment delivery to streams by improving road systems and riparian communities, and rehabilitating abandoned mine sites
- Manage risks from tributary fisheries according to an abundance-based schedule

Chinook salmon are an anadromous fish, which means they spawn in fresh water, but will spend a portion of their lives in the ocean. Snake River spring/summer Chinook salmon exhibit a stream-type life history, which means they rear in fresh water for an extended period (NMFS 2016). Adults (mostly 3- and 4-year-old fish) begin their upstream migration in the Columbia River in late February and early March. Spawning begins in August, and continues through September. Eggs incubate through the fall and winter, and the fry emerge in the late winter and early spring. Most juvenile fish mature in fresh water for one year before they migrate to the ocean in the spring of their second year, and they may be present in the aquatic resources study area year-round.

The EFSFSR population is a summer-run, historically large population, with historic spawning areas throughout the EFSFSR mainstem and in Johnson Creek, which is the population's major tributary (NMFS 2011). Chinook salmon in the upper EFSFSR (upstream from the Yellow Pine pit) were extirpated by the diversion of the EFSFSR into a bypass tunnel for mining operations late in the 1930s, and, after the cessation of mining, the very high-gradient riffle created that precludes fish passage past the Yellow Pine pit into the upper watershed. Most current spawning occurs in Sugar Creek and Johnson Creek. Recently, the IDFG, in cooperation with the Nez Perce Tribe, has planted Chinook salmon above Yellow Pine pit into Meadow Creek, which have spawned successfully (Nez Perce 2009, 2010, 2011). A total of 459 Chinook salmon were released in 2011, 294 in 2012, 130 in 2013, 100 in 2015, and 536 in 2016 (Folsom 2013, McPherson 2013, Felty 2015). No fish were released in 2014.

4.2.2 Snake River Basin Steelhead

Snake River Basin steelhead (*O. mykiss*) are an anadromous form of rainbow trout. These fish were listed as a threatened Evolutionarily Significant Unit (ESU) on August 18, 1997 (NMFS 1997), including all natural-origin populations of steelhead in the Snake River Basin of southeast Washington, northeast Oregon, and Idaho. They were subsequently reclassified as a threatened Distinct Population Segment (DPS) (NMFS 2006). The final rule designating Critical Habitat for steelhead was published by NMFS on September 2, 2005 (NMFS 2005), and took effect on January 2, 2006. Critical Habitat for Snake River Basin steelhead is designated in the EFSFSR up to approximately 1,200 feet upstream from the confluence with Sugar Creek, and includes Sugar Creek as well as Burntlog Creek.

The 2016 Recovery Plan included recovery strategies for the Salmon River steelhead MPG. Recovery strategies for steelhead that effect the EFSFSR watershed include the intent to:

- Collect and analyze population-specific data to accurately determine population status
- Maintain wilderness protection and protect pristine tributary habitat
- Eliminate passage barriers and improve connectivity to historical habitat
- Reduce and prevent sediment delivery to streams by rehabilitating roads and mining sites
- Manage risks from tributary fisheries through updated Fisheries Management Evaluation Plans and Tribal Resource Management Plans, and according to an abundance-based schedule

Snake River Basin steelhead have traditionally been assigned to two groups (A-run and B-run) based on the bimodal timing of passage into the Columbia River (as measured at Bonneville Dam), and by certain life-history characteristics (Busby et al. 1996). Steelhead in the aquatic resources study area are considered B-run steelhead, which means they pass Bonneville Dam after August 25, and tend to return after two years in the ocean (Campbell et al. 2012). Snake River Basin steelhead enter fresh water from June to October and spawn during the following spring from March to May. Their eggs incubate in redds for up to four months before hatching as alevins, a larval life stage dependent on food stored in a yolk sac. Snake River Basin steelhead usually smolt when they reach two or three years of age.

Steelhead occur widely in the EFSFSR, downstream from the Yellow Pine pit. Although spawning is not well documented, redds and adults were identified in 2004 downstream from the town of Yellow Pine. Most of the spawning sites were in small pockets of suitable substrate, rather than in well-developed spawning riffles (Nelson 2004). Some steelhead may also spawn upstream from the town of Yellow Pine. Anadromous steelhead spawning is restricted to below the Yellow Pine pit at Stibnite, coincident with the upstream endpoint for designated critical habitat in the EFSFSR.

4.2.3 Resident Rainbow/Redband Trout

Redband trout (*O. mykiss gairdneri*) is a subspecies of rainbow trout. Because it coexists with other subspecies of rainbow trout, it can compete and hybridize with the other forms of rainbow trout. Redband trout are thought to occur in the aquatic resources study area, though genetic studies have not been conducted to confirm their presence.

In the aquatic resources study area, resident redband or rainbow trout likely spawn between mid-March and early June, with incubation and emergence occurring until mid-July (Miller et al. 2014). This timing overlaps with steelhead in the area, which results in the inability to separate juveniles during field identifications. Because juvenile steelhead, and juvenile and adult redband trout, and rainbow trout cannot be easily distinguished, this report will refer to each as *O. mykiss*.

4.2.4 Columbia River Bull Trout

Columbia River bull trout (*Salvelinus confluentus*) occur in the SFSR watershed. Hereinafter, all references to bull trout are to the Columbia River bull trout. The USFWS designated Critical Habitat for bull trout in the EFSFSR, and in Tamarack, Sugar, Cane, Cinnabar, Meadow, Burntlog, Trapper and Riordan Creeks (USFWS 2010). Within the aquatic resources study area, bull trout occur throughout the EFSFSR and Sugar Creek. Bull trout are found in the control site in Tamarack Creek.

In 2015, the USFWS released the *Recovery Plan for Coterminous United States Population of Bull Trout (Salvelineus confluentus)* (USFWS 2015). They developed recovery unit implementation plans for specific recovery units, including the Upper Snake Recovery Unit, which includes bull trout in the aquatic resources study area. Four strategies were defined for the recovery of bull trout:

- Protect, restore, and maintain suitable habitat conditions
- Minimize demographic threats by restoring connectivity or populations, where appropriate, to promote diverse life-history strategies and conserve genetic diversity
- Prevent and reduce negative effects of non-native fishes and other non-native taxa
- Work with partners to conduct research and monitoring to implement and evaluate recovery activities, consistent with an adaptive-management approach using feedback from implemented, site-specific recovery tasks, and considering the effects of climate change

Bull trout are generally assumed to be in decline across their range, but there are few data on population sizes and trends in the SFSR. In general, the USFS considers the EFSFSR as important bull trout habitat. Bull trout have several life-history strategies: fluvial (i.e., stream and river dwelling, spawning in small tributaries), adfluvial (lake dwelling and river spawning), and anadromous forms. An adfluvial life history uses the Yellow Pine pit for overwintering, with downstream migration to tributaries for spawning (Hogen and Scarnecchia 2006). Migrants stage at the mouths of presumptive spawning tributaries from mid-July to mid-August, then move into tributaries and spawn from mid-August to mid-September. Fluvial populations downstream from the Yellow Pine pit quickly outmigrate as far as the main Salmon River (Hogen and Scarnecchia 2006), or move up to the Yellow Pine pit for overwintering. Populations upstream from the Yellow Pine pit use the fluvial life-history strategy.

Bull trout reach sexual maturity between four and seven years of age, and they are known to live as long as 12 years. They spawn in fall after temperatures drop below 48 degrees Fahrenheit (°F) $(9^{\circ}C)$, in streams with cold, unpolluted water, clean gravel and cobble substrate, and gentle
stream slopes (USFWS 2002). Many spawning areas are associated with cold-water springs or areas where stream flow is influenced by groundwater (USFWS 2002).

Bull trout eggs require a long incubation period (four to five months) compared to other salmon and trout, hatching in late winter or early spring (USFWS 2002). Fry remain in the stream bed for up to three weeks before emerging. Juvenile fish are often found at or near the stream bottom, often in between cobbles and boulders, and among woody-debris complexes (Thurow 1994, Berge and Mavros 2001).

4.2.5 Westslope Cutthroat Trout

Westslope cutthroat trout (*O. clarki lewisi*) are designated by the Regional Forester as a "Sensitive" species. There was a petition to list them as federally threatened (63 FR 31691), but the USFWS determined that such a listing was not warranted (65 FR 20120, April 14, 2000). Westslope cutthroat trout occur in the aquatic resources study area, both upstream and downstream from the Yellow Pine pit.

Westslope cutthroat trout in the aquatic resources study area are either resident or fluvial, living in rivers, but then migrating into the tributaries to spawn. They begin to mature at age three, but they usually spawn first at age four or five. They spawn between March and July when water temperatures are near 50°F (10°C). Emerging fry are around 20 mm in length. Westslope cutthroat trout feed primarily on macroinvertebrates—particularly immature and mature forms of aquatic insects, terrestrial insects, and, in lakes, zooplankton (USFWS 1999).

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SECTION 5 AQUATIC SURVEY RESULTS

This section describes the results of the stream habitat surveys (cobble embeddedness, free matrix, modified McNeil core substrate, PIBO habitat, and water temperature), macroinvertebrate surveys, metals surveys, and fish surveys that were conducted between 2012 and 2016.

5.1 Stream Habitat

The PAF WCI standards were used as the base metrics for the minimum criteria to determine desired habitat conditions in the aquatic resources study area. The results of the cobble embeddedness, free matrix, modified McNeil core substrate, PIBO habitat, and water temperature surveys are described in detail below.

In addition to the surveyed and documented elements in this report, other factors may have contributed to variability within the study area. Observed natural disturbance events (e.g., avalanches, landslides or windfall) and federal management activities (road decommissioning) that occurred within the aquatic resource study area between 2012 and 2016 may also have had an influence on variability between sample years.

5.1.1 Cobble Embeddedness

The average levels of cobble embeddedness for each survey by site are shown in **Figure 5-1** and **Figure 5-2**, along with WCI indicator ranges for reference purposes. In general, cobble embeddedness values for the aquatic resources study area were within the *functioning appropriately* WCI ranges, except for two sites on lower Meadow Creek (MWH-014, MWH-049). The multiple-year averages would currently put these two lower Meadow Creek sites as *functioning as risk*, but they have been steadily decreasing annually.

Summary cobble embeddedness statistics for each site and subwatershed are shown in **Table 5-1** and **Table 5-2.** Because there is climatic and hydrologic variability in natural systems, developing a minimum five-year mean is recommended to create a statistically robust dataset. Five-year means were used to develop WCI indicator values, against which baseline study site data will be compared to indicate future conditions (Nelson and Burns 2005, Zurstadt et al. 2016). A comparison of means, using Tukey's Honestly Significant Differences test or ANOVA, will not be conducted for the surveyed sites until an adequate sample size of data has been attained.

Box and whisker plots were created by year, for each survey site and each subwatershed, and they are shown in **Appendix 1**.

Figure 5-1. Average Cobble Embeddedness by Site for Non-granitic Watersheds in 2013-2016

Green shading = Functioning Appropriately Yellow shading = Functioning at Risk Red shading = Functioning at Unacceptable Risk

Key:

% = percent

A = Functioning Appropriately

CI = confidence interval

CV = coefficient of variation R = Functioning at Risk

UR = Functioning at Unacceptable Risk

WCI = Watershed Condition Indicator

Note:

Subwatersheds are defined as 6th field hydrologic unit code

Annual mean is the percent of cobble embeddedness averaged over all subsamples measured

Key:

% = percent

A = Functioning Appropriately

CI = confidence interval

CV = coefficient of variation

R = Functioning at Risk

UR = Functioning at Unacceptable Risk

WCI = Watershed Condition Indicator

5.1.2 Free Matrix and Surface Fines

The average levels of free matrix and surface fines for each survey site are shown in **Figure 5-3** through **Figure 5-6**, along with WCI indicator ranges for reference purposes.

Free matrix values for the aquatic resources study area are generally considered functioning appropriately, except for select granitic sites on Meadow Creek (MWH-014, MWH-049, and MWH-060), and the two non-granitic sites on Sugar Creek (MWH-008) and Profile Creek (MWH-061). All five sites have multiple-year means that correspond to the *functioning at risk* free matrix threshold (**Table 5-3**).

In contrast with the free matrix results, most surface fines values for the aquatic resources study area were generally either considered functioning at risk or functioning at unacceptable risk. Only the annual means of granitic sites MWH-009, MWH-050, MWH-051, and MWH 054 are functioning appropriately (**Table 5-3**).

Currently, there are no more than four years of free matrix and surface fines data for any specific sample location. To better account for data variability—due to natural (e.g., climate, hydrology), spatial, and temporal factors, as well as measurer error—a minimum of five years of data is recommended to better evaluate baseline conditions. Five-year means were used to develop WCI indicator values, against which baseline study site data will be compared as an indicator of future conditions (Nelson and Burns 2005, Zurstadt et al. 2016). Tiered WCIs have been developed by the PAF to support management decisions regarding substrate information in granitic and nongranitic watersheds (Nelson and Burns 2005, Nelson and Burns 2007, Zurstadt et al. 2016).

Table 5-3 and **Table 5-4** summarize results of the free matrix and surface fines analyses for each site and subwatershed. A comparison of means, using Tukey's Honestly Significant Differences test or ANOVA, will not be conducted until an adequate sample range of a minimum of 5 years of data has been attained for each survey site. Box and whisker plots that were created by year, for each site and each subwatershed, are presented in **Appendix 1**. The plots display sample variation of a statistical population, without making any assumptions of the underlying statistical distribution, and they assist in data interpretation and quality control.

Note: Site locations (ordered downstream to upstream): MWH-033, MWH-032: East Fork of the South Fork of the Salmon River MWH-058* and MWH-061*: Profile Creek (control sites) MWH-017*: Tamarack Creek (control site) MWH-008 and MWH-010: Sugar Creek

Key:

Green shading = Functioning Appropriately

Yellow shading = Functioning at Risk

Red shading = Functioning at Unacceptable Risk

Figure 5-3. Average Free Matrix by Site for Non-Granitic Watersheds in 2013-2016

Note: There were no qualifying substrate material present in the 2014 sampling event for MWH-055, so no annual average is presented.

Site Locations (ordered downstream to upstream):

MWH-006: East Fork Meadow Creek

MWH-009, MWH-059, MWH-013 and MWH-044: East Fork of the South Fork of the Salmon River MWH-049, MWH-014, MWH-060, MWH-047, MWH-016, and MWH-034: Meadow Creek

Key:

Green shading = Functioning Appropriately

Yellow shading = Functioning at Risk

Red shading = Functioning at Unacceptable Risk

Figure 5-4a. Average Free Matrix by Site for Granitic Watersheds in 2013-2016

Note: There were no qualifying substrate material present in the 2014 sampling event for MWH-055, so no annual average is presented.

Site Locations (ordered downstream to upstream):

MWH-051, MWH-050 and MWH-052: Burntlog Creek MWH-054 and MWH-053: Trapper Creek MWH-055: Riordan Creek MWH-056* and MWH-057*: Fourmile Creek and Goat Creek, respectively (control sites) MWH-062: Fiddle Creek

Key: Green shading = Functioning Appropriately

Yellow shading = Functioning at Risk

Red shading = Functioning at Unacceptable Risk

Figure 5-4b. Average Free Matrix by Site for Granitic Watersheds in 2014-2016

Note: Site locations (ordered downstream to upstream):

MWH-008 and MWH-010: Sugar Creek

MWH-017*: Tamarack Creek (control site)

MWH-033, MWH-032: East Fork of the South Fork of the Salmon River MWH-058* and MWH-061*: Profile Creek (control sites)

Key: Green shading = Functioning Appropriately

Yellow shading = Functioning at Risk

Red shading = Functioning at Unacceptable Risk

Note: Site Locations (downstream to upstream):

MWH-006: East Fork Meadow Creek

MWH-009, MWH-059, MWH-013 and MWH-044: East Fork of the South Fork of the Salmon River MWH-049, MWH-014, MWH-060, MWH-047, MWH-016, and MWH-034: Meadow Creek

Key:

Green shading = Functioning Appropriately Yellow shading = Functioning at Risk Red shading = Functioning at Unacceptable Risk

Figure 5-6a. Average Surface Fines by Site for Granitic Watersheds in 2013-2016

Note: Site Locations (downstream to upstream):

MWH-051, MWH-050 and MWH-052: Burntlog Creek MWH-054 and MWH-053: Trapper Creek MWH-055: Riordan Creek MWH-056* and MWH-057*: Fourmile Creek and Goat Creek, respectively (control sites) MWH-062: Fiddle Creek Green shading = Functioning Appropriately

Key:

Yellow shading = Functioning at Risk

Red shading = Functioning at Unacceptable Risk

Figure 5-6b. Average Surface Fines by Site for Granitic Watersheds in 2014-2016

Note:

Annual mean is the percent of free matrix (loose cobble) averaged over all subsamples measured Subwatersheds are defined as 6th field hydrologic unit code

Key:

% = percent

A = Functioning Appropriately

CI = Confidence interval

CV = Coefficient of Variation

EFMC = East Fork Meadow Creek

EFSFSR = East Fork of the South Fork of the Salmon River

R = Functioning at Acceptable Risk

SFSR = South Fork of the Salmon River

UR = Functioning at Unacceptable Risk

WCI = Watershed Condition Indicators

Stream/ Subwatershed	MWH Site ID	Geologic Type	Year	Sample Size (n)	Annual Mean $(\%)$	Standard Deviation	Standard Error	95 % CI $(\%)$	CV	WCI	
		Granitic	2013	30	40	23	4	± 8	0.6	UR	
EFMC / Headwaters EFSFSR			2014	Not Sampled							
	006		2015	30	16	12	$\overline{2}$	±4	0.7	${\sf R}$	
			2016	30	17	24	4	± 8	1.4	${\sf R}$	
	008	Non- granitic	2013	30	10	7	1	± 3	0.7	UR	
Sugar Creek/ Sugar Creek			2014	30	18	23	$\overline{4}$	± 8	1.3	UR	
			2015	30	9	8	1	±3	0.9	UR	
			2016	30	10	9	$\overline{2}$	± 3	1.0	UR	
EFSFSR / No Mans Creek - EFSFSR	009	Granitic	2013	30	8	$\overline{7}$	1	± 3	0.9	Α	
			2014	30	10	8	1	± 3	0.8	Α	
			2015	30	9	9	$\overline{2}$	± 3	1.1	Α	
			2016	30	12	9	$\overline{2}$	±3	0.8	Α	
		Non- granitic	2013	30	7	5	1	±2	0.7	UR	
Sugar Creek/			2014	30	12	8	1	±3	0.7	UR	
Sugar Creek	010		2015	30	8	8	1	±3	1.0	UR	
			2016	30	9	6	1	±2	0.8	UR	
	013	Granitic	2013	30	17	12	$\overline{2}$	±4	0.7	R	
EFSFSR /			2014	Not sampled							
Headwaters EFSFSR			2015	Not Sampled							
			2016	30	11	17	3	±6	1.5	Α	
Meadow Creek/	014	Granitic	2013	30	20	16	3	±6	0.8	UR	
			2014	30	26	12	$\overline{2}$	±4	0.5	UR	
Headwaters			2015	Not Sampled							
EFSFSR			2016	30	22	22	$\overline{4}$	±8	1.0	UR	
Meadow Creek/ Headwaters EFSFSR	016	Granitic	2013	Not Sampled							
			2014	30	25	17	3	±6	0.7	UR	
			2015	30	20	13	$\overline{2}$	± 5	0.7	UR	
			2016	30	18	13	$\overline{2}$	±5	0.7	UR	
Tamarack Creek / Tamarack Creek	017	Non- granitic	2013	30	3	3	1	±1	1.3	Α	
			2014	30	$\overline{2}$	$\overline{4}$	1	± 1	1.6	Α	
			2015	30	3	$\overline{4}$	1	±2	1.5	Α	
			2016	30	11	16	3	±6	1.4	UR	
EFSFSR / No Mans Creek - EFSFSR	032	Non- granitic	2013	30	1	3	$\overline{0}$	±1	3.3	A	
			2014	30	10	9	$\overline{2}$	±3	0.9	UR	
			2015	Not Sampled							
			2016	Not Sampled							
		Non- granitic	2013	30	5	5	1	±2	1.0	R	
EFSFSR / No Mans Creek - EFSFSR	033		2014	30	16	$\overline{7}$	1	±3	0.5	UR	
			2015	Not Sampled							
			2016	$30\,$	$\, 8$	6	$\mathbf{1}$	± 2	0.8	UR	

Table 5-4. Surface Fines Analysis Results by Site in 2013-2016

Note:

Sites are sampled for free matrix only at sites concurrent with a cobble embeddedness survey, or at sites during which a PIBO habitat survey also was conducted. These surveys may not occur every year.

Annual mean is the percent of surface fines averaged over all subsamples measured.

Subwatersheds are defined as 6th field hydrologic unit code.

Key:

% = percent

A = Functioning Appropriately

CI = Confidence interval

CV = Coefficient of Variation

EFMC = East Fork Meadow Creek

EFSFSR = East Fork of the South Fork of the Salmon River

NS = Not sampled

R = Functioning at Acceptable Risk

SFSR = South Fork of the Salmon River

UR = Functioning at Unacceptable Risk

WCI = Watershed Condition Indicators

Subwatershed	Geologic Type	Year	Sample Size (n)	Annual Mean (%)	Standard Deviation	Standard Error	95 % CI (%)	CV	WCI	
Headwaters EFSFSR	Granitic	2013	180	51	25	$\overline{2}$	±4	0.5	A	
		2014	150	21	16	1	±3	0.8	$\mathsf R$	
		2015	210	43	24	$\overline{2}$	± 3	0.6	A	
		2016	300	33	23	$\mathbf{1}$	± 3	0.7	A	
	Non- granitic	2013	60	47	23	1	±2	0.8	R	
Sugar Creek		2014	60	54	21	$\overline{2}$	±4	1.1	$\mathsf R$	
		2015	60	52	19	$\overline{2}$	± 5	0.4	R	
		2016	60	61	14	$\overline{2}$	±4	0.2	Α	
		2013	90	68	16	$\overline{2}$	± 3	1.4	A	
No Mans		2014	90	53	17	\overline{c}	±3	0.8	A	
Creek - EFSFSR	Granitic	2015	30	53	24	$\overline{4}$	± 9	0.5	A	
		2016	60	48	20	\mathfrak{Z}	± 5	0.5	A	
Tamarack Creek	Non- granitic	2013	30	73	13	1	±1	1.3	Α	
		2014	30	66	15	1	± 1	1.6	A	
		2015	30	71	14	\mathfrak{Z}	± 5	0.2	A	
		2016	30	58	14	3	± 5	0.2	A	
Burntlog Creek (Lower)	Granitic	2013	Not Sampled							
		2014	60	54	22	1	±2	0.8	Α	
		2015	60	57	23	3	±6	0.4	Α	
		2016	30	52	15	3	±5	0.3	Α	
	Non- granitic	2013	Not Sampled							
		2014	60	47	18	1	± 1	1.1	$\mathsf R$	
Profile Creek		2015	60	51	22	3	± 6	0.4	${\sf R}$	
		2016	30	53	20	$\overline{4}$	± 7	0.4	$\mathsf R$	
Goat Creek - SFSR	Granitic	2013	Not Sampled							
		2014	30	24	22	4	± 7	1.1	R	
		2015	$30\,$	34	23	4	± 8	0.7	Α	
		2016	30	29	19	$\overline{4}$	±7	0.7	$\mathsf A$	
	Granitic	2013	Not Sampled							
Fourmile		2014	30	25	15	2	±4	0.7	R	
Creek - SFSR		2015	30	46	17	\mathfrak{Z}	± 6	0.4	Α	
		2016	30	34	18	$\overline{3}$	± 7	0.6	A	
		2013	Not Sampled							
Riordan Creek	Granitic	2014	30	$\boldsymbol{0}$	O	$\overline{4}$	±7	0.2	44	
		2015	30	88	28	5	±10	0.3	Α	
		2016	29	100	$\overline{0}$	$\overline{0}$	N/A	N/A	A	

Table 5-5. Free Matrix Analysis Results by Subwatershed in 2013-2016

Table 5-5. Free Matrix Analysis Results by Subwatershed in 2013-2016 (*continued***)**

Note:

Annual mean is the percent of free matrix (loose cobble) averaged over all subsamples measured Subwatersheds are defined as 6th field hydrologic unit code

Key:

A = Functioning Appropriately

CI = Confidence interval

CV = Coefficient of Variation

EFSFSR = East Fork of the South Fork of the Salmon River

R = Functioning at Acceptable Risk

SFSR = South Fork of the Salmon River

UR = Functioning at Unacceptable Risk

WCI = Watershed Condition Indicators

Note:

Annual mean is the percent of surface fines averaged over all subsamples measured

Subwatersheds are defined as 6th field hydrologic unit code

Key:

A = Functioning Appropriately

CI = Confidence interval

CV = Coefficient of Variation

EFSFSR = East Fork of the South Fork of the Salmon River

SFSR = South Fork of the Salmon River

R = Functioning at Acceptable Risk

UR = Functioning at Unacceptable Risk

WCI = Watershed Condition Indicators

5.1.3 Modified McNeil Core Sample

The WCI classification for core sampling efforts requires a 5-year mean. To date, only four years of data have been collected, and as a result, no WCI classification can be identified (Nelson and Burns 2005, Nelson and Burns 2006, Zurstadt et al. 2016). A 5-year mean value for sediment that is less than or equal to 28 percent is considered functioning appropriately. For sediment less than 0.85 mm, an established WCI threshold does not exist. A 5-year mean of less than 10 percent is considered functioning appropriately, as a result of Jensen et al. (2009) modeling efforts, and PAF recommendations (Zurstadt et al. 2016). Jensen et al. (2009) reported that at the 10 percent threshold for fines that are less than 0.85 mm, the odds of salmon and steelhead eggto-fry survival decreased dramatically. **Figure 5-7** compares average depth fines from core samples during the sampling period of 2013 to 2016.

Note:

Bars on columns represent 95 % confidence intervals Key: < = Less than % = percent $mm =$ millimeter

Figure 5-7. Core Sample Average Depth Fines and 95% Confidence Intervals at MWH-033 in the East Fork of the South Fork of the Salmon River by Survey Year

Table 5-7 summarizes the average percentage of depth fines in the EFSFSR at MWH-033 from 2013 to 2016. A comparison of means, using Tukey's Honestly Significant Differences test or ANOVA, will not be conducted until at least five years of data have been recorded (Zurstadt 2014).

Box and whisker plots, created by year for core sampling results, are presented in **Appendix 1**. The plots display sample variation of a statistical population, without making any assumptions of the underlying statistical distribution, and they assist in data interpretation and quality control.

The percent of depth fines material that was less than 6.33 mm for all four years ranged from 15 to 30 percent (**Table 5-7**). Even though a 5-year mean is required to classify the WCI for the core substrate samples, the current 4-year average is 23 percent, and would be considered functioning appropriately. Annual means for the percent of depth fines less than 0.85 mm ranged from 4 to 9 percent (**Table 5-7**). All years would be considered functioning appropriately at the Jensen et al. (2009) recommended threshold of less than 10 percent (**Table 5-7**).

	Sample Year	Sample Size (n)	Annual Mean (%)	Standard Deviation	Standard Error	95 % Confidence Interval (%)	Coefficient of Variation
Percent Fines <6.33 mm	2013	40	30	10	2	± 3	0.3
	2014	40	26	10	$\overline{2}$	± 3	0.4
	2015	40	18	9	1	± 3	0.5
	2016	40	15	11	2	± 3	0.7
Percent Fines < 0.85 mm	2013	40	9	6		±2	0.6
	2014	40	$\overline{7}$	3		±1	0.5
	2015	40	4	3	0.5	±1	0.7
	2016	40	4	6		±2	1.4

Table 5-7. Average Percentage of Depth Fine Material at Site MWH-033 in the East Fork of the South Fork of the Salmon River in 2013-2016

Note: Annual mean is the percent of depth fine material averaged over all subsamples measured.

Key:

< = Less than

 $% =$ percent

mm = millimeter

5.1.4 Channel Morphology and Condition (PIBO Survey)

Table 5-8 presents the overall average parameter for the morphology and conditions of the stream channel recorded at each PIBO survey site for 2012 through 2015. The PAF advised not using WCI indicators yet for PIBO until a larger dataset is collected (Zurstadt 2014). Complete statistical tables for each site and subwatershed are presented in **Appendix 1**.

Three ANOVA tests were performed, to determine if a statistically significant difference exists between survey sites, by using data for three dependent variables (bank angle, maximum pool depth, and WW:WD ratio). Bank angle can be influenced by many variables including stream gradient, geology, flow, the presence of riparian vegetation, and both natural and human-caused disturbances. These influences may be driving ANOVA results among the samples. Pool depth can also be affected by several factors, the most influential being stream stage at the time of sampling. Pool depths change annually, and some pools may deepen and have sediment deposits that can change pool depth. Additionally, due to PIBO protocol criteria for measuring pools, certain pools may not be counted from year-to-year if they are not at least half the wetted width of the channel. However, the differences shown among pool-depth ANOVA test results may be due to the higher maximum depths of pools at the highlighted sites compared to other sites. The WW:WD ratio is driven mainly by stream stage, which is determined by stream flow, at the time data is collected. As stream stage increases, the depth also increases; therefore, the likelihood of the wetted width also increases. For PIBO variables that are summarized at the plot level and not the individual transect level, ANOVA was not performed since only a maximum of two years of data has been collected at any particular PIBO survey location. PIBO variables where ANOVA was not performed include pool frequency, LWD frequency, percent surface fines, substrate mean size (D_{50}) , and bank stability (non-numeric). Summary tables are included in Section 2 of **Appendix 1** for these variables. Additional details and results from the ANOVA testing are discussed in Section 3 of **Appendix 1**.

Table 5-8. PIBO Reach Average Parameter Results for 2012-2015

Table 5-8. PIBO Reach Average Parameter Results for 2012-2015 (*continued***)**

Note:

a = LWD protocols were modified by the PAF during the baseline study period. Large differences between sample years may be as a result of natural causes (e.g., avalanches, windfall in burned areas) or as a result of changed protocols (Zurstadt et al. 2016b).

b = Bank angle not added to PIBO protocol until 2014.

c = In 2012, pool lengths were measured instead of widths, so data are not comparable.

Key:

% = percent

CS = Covered Stable CU = Covered Unstable

D50 = substrate mean size

EFMC = East Fork Meadow Creek

EFSFSR = East Fork of the South Fork of the Salmon River LWD – large woody debris mm = millimeter US = Uncovered Stable UU = Uncovered Unstable WW:WD = wetted width-to-wetted depth

Box and whisker plots created by year for PIBO surveys are presented in Section 1 of **Appendix 1**. The box and whisker plots show variation in samples of a statistical population, without making any assumptions of the underlying statistical distribution, and they assist in data interpretation and quality control.

5.1.5 Water Temperature Evaluation

Table 5-9 and **Appendix 3** present 7-day average maximum water temperature WCIs for spawning and rearing Chinook salmon and steelhead; and bull trout spawning, incubation and rearing based on Appendix B of the Land and Resource Management Plan (PAF 2003). Sites where the thermographs were not installed until 2014 were lost, or were inaccessible during latefall data collection activities, which are shown as data gaps in the figures in **Appendix 3**.

Table 5-9. Water Temperature Watershed Condition Indicators for Chinook Salmon, Steelhead and Bull Trout Based on the 7-Day Average Maximum Temperature

Source: PAF 2003, page B-13 in Appendix B

5.2 Macroinvertebrates

The final SMI scores were consistently high across sites and years because individual scores for most metrics were high, with several consistently exceeding the highest score of 100 (**Table 5-13** and **Appendix 2**). The overall score was *very good* for seven of the 11 survey sites in all four sampling years, attaining the highest designation in the SMI scoring system. Among the seven consistently "very good" sites, there was relatively little variation in SMI scores between years, except for MWH-010 (upper Sugar Creek), which had a decreased score of 83.2 in 2016 as compared to a range of 92.4 to 95.6 from 2012 to 2014. Even though the overall rating at MWH-010 was still considered *very good*" it's worth noting that the departure from previous scores appeared to be driven by lower Trichoptera richness, the percentage of Plecoptera, and lower water quality tolerance (i.e., HBI).

SMI metrics with consistently high scores included total taxa richness, Ephemeroptera taxa (mayflies) richness, Trichoptera taxa (caddisflies) richness, percent 5 dominant taxa (i.e., most abundant), HBI (a measure of water-quality tolerances), scraper taxa richness, and clinger taxa richness. However, lower SMI scores for 2016—compared to previous years' surveys at MWH-007, MWH-010, MWH-015, MWH-016 and MWH-017—were generally the result of lower taxa richness (particularly Plecoptera and Trichoptera), percent Plecoptera, HBI scores, scraper taxa, percent 5 dominant, and clinger taxa. There is some redundancy in these measures, particularly among the richness metrics, so changes in overall diversity may have affected all of these scores.

Plecoptera taxa (stoneflies) richness had somewhat higher variation among sites, but generally it produced high metric scores, except for a sharp decrease over time at site MWH-007 on the EFSFSR, located immediately downstream of the Yellow Pine pit outlet (**Table 5-13** and **Appendix 2**). The cause for this remains unclear considering other metric scores appear healthy.

The lowest metric scores, and widest variation among survey sites, was for percent Plecoptera individuals (**Table 5-13** and **Appendix 2**). Sites MWH-010 (Sugar Creek) and MWH-017 (Tamarack Creek) had the highest average percentage of Plecoptera individuals over the three sampling years, about twice the average metric scores of the other sites, while site MWH-007 had the lowest. In general, most sites exhibited high Plecoptera diversity, but relatively low numbers.

Given the high taxa richness of all the sites, it is difficult to draw firm conclusions regarding relatively low Plecoptera abundance, which could be related to high habitat diversity. Ultimately, relatively low abundance of Plecoptera individuals did not substantially affect the overall SMI scores because of consistently high taxa richness, an abundance of taxa having low water quality tolerance scores, and an abundance of clinger and scraper taxa, which require high-quality stream habitat in the Central and Southern Mountains region, namely clean cobble substrate.

HBI scores were generally lower across all sites in 2016 compared to previous years, indicating that the HBI values were higher. HBI values are the combination of tolerance of organic pollution weighted by the abundance of those taxa in each sample. These higher HBI values (and corresponding lower HBI scores) mean that the types and abundances of taxa collected at those survey sites were skewed towards being more tolerant of organic pollution than in previous years.

Additional metrics results are summarized below, and further described in **Appendix 2**.

Higher MTI values indicate that the taxa composition exhibit greater tolerance to metals. Across all the sites, the MTI values were low, ranging from 1.46 to 3.09, and showed relatively little variation across years with one exception. In 2012, the MTI value was 4.7 at site MWH-015 (a constructed channel on Meadow Creek), but it decreased to 2.42 in 2013, and then to 1.98 by 2014, possibly indicating that metals concentrations within the constructed channel are decreasing, causing a shift towards a greater number of taxa and/or proportion of individuals that are relatively intolerant of metals contamination (**Table 5-13**). In 2016, two sites exhibited MTI values greater than 3.0. These sites were MWH-007 (immediately downstream of the Yellow Pine pit; MTI = 3.49) and at MWH-016 (the furthest upstream Meadow Creek site; MTI = 3.81). The increased MTI score (i.e., lower tolerance to metals) at the upstream Meadow Creek site is a marked departure from earlier years' results. Another site showing a notable departure from previous years was at site MWH-009, the next site downstream from MWH-007. The MTI values at this site from 2012 to 2014 were consistently around 2.0, whereas in 2016 the value was 2.97, indicating possible increased metals contamination at this site.

Intolerant taxa richness was extremely high, for all survey sites and years, at approximately 75 percent of the taxa in individual samples (**Table 5-13**). This metric indicates that all survey sites are supporting many taxa that are intolerant to poor water quality. As with other 2016 richness metrics, MWH observed a notable decrease in intolerant taxa richness at site MWH-016, which was a sharp departure from previous years. This site had the highest and most consistent number of taxa intolerant of organic pollution from 2012 to 2014, but had the lowest number of intolerant taxa in 2016.

The percent of tolerant individuals was low, comprising no more than 4 percent of all individuals at any of the survey sites (**Table 5-13**). This is consistent with other metrics that indicate high water quality among all sites in the project area. In 2016, there was a relatively large increase in percent tolerant individuals at the furthest downstream Meadow Creek site (MWH-014), and at the site immediately downstream (MWH-012), which is located just below the Meadow Creek/EFSFSR confluence. Although the intolerant macroinvertebrates comprised only about 4 to 6 percent of the samples at those sites in 2016, in previous years, intolerant organisms comprised no more than 2 percent of the samples in any given year. This may represent an increase in pollution, particularly when other changes in taxa richness and metals tolerance at Meadow Creek sites are considered.

Shannon-Weaver index values ranged from 2.68 to 3.56, with only a few sites having values of less than 3.0. These values indicate a high diversity among all sites, which is well supported by other measures of richness and evenness for these sites. The lowest values, of all sites and years, were values of 2.68 at MWH-007 (just downstream from the Yellow Pine pit outlet) in 2014, and 2.72 at MWH-014 (Meadow Creek) in 2013 (**Table 5-10**).

The FFG composition of samples across all sites and years was consistent and stable, particularly for FFG richness values across all sites in all survey years, indicating no notable shifts in in habitat conditions. Richness across sites and years was dominated by predators and gatherers, and to a lesser extent shredders and scrapers, along with a low, but stable percentage of filterer taxa.

The data indicates that there is stability and consistency of available habitat and food sources at the study sites, particularly with the apparently rich predator population, which is dependent on diverse and stable food resources in the form of other macroinvertebrates.

Numbers of individuals were dominated by scrapers and gatherers, with a few notable instances of a high proportion of shredders (MWH-013 in 2012, and MWH-015 in 2013). The FFG composition is consistent and evenly distributed among the different FFG designations across sites and sampling years.

The most obvious difference in FFG abundance is the large proportion of filterer individuals at MWH-007, located just downstream of the Yellow Pine pit outlet. Filterers increased steadily from 2012 to 2014 at this site. In 2013 and 2014 this was almost entirely due to a large number of blackflies in the sample. The increase in abundance from 2012 to 2014 may indicate an increase in particulates, including organic/particulate discharge from the Yellow Pine pit. However, the difference may also be due to sampling bias, since blackfly larvae can be found in high densities in such habitat, and therefore, it is possible for a single sampling repetition to inadvertently contribute a disproportionate number of individuals to the combined sample.

The PIBO O/E scores among sampling sites from 2012 through 2016 were consistently rated as being in *good* biological condition, except in three instances where a site was rated as *fair* (**Table 5-11**). A *fair* condition rating was found in 2012 at MWH-007 (Meadow Creek confluence), and in 2013 at MWH-029 (Sugar Creek) and MWH-015 (Meadow Creek). In all three instances, sites with *fair* O/E scores improved to *good* in subsequent sampling years. The O/E condition rating results indicate that the macroinvertebrate assemblages are consistently very comparable to reference conditions in the Western States Ecoregion for which the O/E model was developed (Stoddard et al. 2006).

An overall description of the results, by individual sample locations, is presented in **Appendix 2**.

Table 5-10. Idaho Stream Macroinvertebrate Index Metrics, Scores and Ratings for Sampling Sites in 2012-2014 and 2016

Table 5-10. Idaho Stream Macroinvertebrate Index Metrics, Scores and Ratings for Sampling Sites 2012-2014 and 2016 (*continued***)**

Table 5-10. Idaho Stream Macroinvertebrate Index Metrics, Scores and Ratings for Sampling Sites 2012-2014 and 2016 (*continued***)**

Table 5-10. Idaho Stream Macroinvertebrate Index Metrics, Scores and Ratings for Sampling Sites 2012-2014 and 2016 (*continued***)**

Table 5-11. PIBO Observed/Expected Index Scores for MWH Sampling Sites in 2012-2014 and 2016

Note: Observed/Expected Index is based on results from the RIVPACS statistical model.

In 2012, samples were collected slightly downstream from the current MWH-029, however, the collection location is close and has similar habitat, so is considered the same as MWH-029.

Key:

EFSFSR = East Fork of the South Fork of the Salmon River

O/E = Observed/Expected
5.3 Metals

The results of the metals analysis are discussed below for sediment, macroinvertebrates and fish tissue. Full laboratory results for all metals analyses are presented in **Appendix 4**.

5.3.1 Sediment

Table 5-12 presents sample results for only those metals for which EPA provides screening-level benchmarks (2006). Antimony, arsenic and mercury exceeded their respective benchmarks at most survey sites, generally by one or more orders of magnitude. Iron and manganese exceeded benchmarks at five sites. Cadmium exceeded the benchmark at MWH-007 and MWH-030 (both on the EFSFSR); copper exceeded the benchmark at MWH-017 (Tamarack Creek); and silver exceeded benchmarks at MWH-011 and MWH-012 (both located on the EFSFSR). Selenium concentration exceeded the benchmark at one site (MWH-012), but at all other sites, a comparison could not be made. At those non-comparative sites, the selenium values were reported as less than the Practical Quantitation Limit (PQL) of 4 milligrams per kilogram (mg/kg) dry weight which was twice the benchmark quantitation of 2 mg/kg dry weight.

Table 5-12. Sediment Metals Concentrations Compared to Screening Level Benchmarks

Note: Benchmarks obtained from Environmental Protection Agency website on November 15, 2017, Freshwater Sediment Screening Benchmarks developed for Region 3, at [https://www.epa.gov/risk/freshwater-sediment-screening-benchmarks.](https://www.epa.gov/risk/freshwater-sediment-screening-benchmarks)

*Several samples had higher mercury PQL due to required dilutions.

Shaded cells are higher than the applicable screening benchmark

Key:

mg/kg = milligrams per kilogram

PQL = Practical Quantitation Limit

YPP = Yellow Pine pit

5.3.2 Macroinvertebrates

Table 5-13 presents the sample results for metals concentration in macroinvertebrate tissue. Concentrations of metals in macroinvertebrate tissue were elevated at many sites on the EFSFSR and Sugar Creek, relative to concentrations at other sites—particularly sites on Meadow Creek and Tamarack Creek. For example, the highest tissue concentrations of arsenic were on the EFSFSR at the sites further downstream (MWH-009, MWH-007 and MHW-030). The maximum tissue concentration of arsenic was at MWH-030, with an arsenic concentration of 572 mg/kg dry weight. By contrast, tissue concentration of arsenic at the reference site on Tamarack Creek was 4.38 mg/kg dry weight. The arsenic concentrations were lower at the most upstream EFSFSR site, the most upstream Meadow Creek site, and the lowest Sugar Creek site. The highest tissue concentrations of antimony were also on the EFSFSR at the more downstream sites (particularly MWH-007 and MWH-030). Maximum tissue concentration was found at MWH-007, with an antimony concentration of 27.2 mg/kg dry weight. Antimony concentrations in tissue were lower at the more upstream sites on the EFSFSR, and much lower at the Meadow Creek, Sugar Creek, and Tamarack Creek sites. The minimum tissue concentration of antimony, 0.07 mg/kg dry weight, was at the reference site on Tamarack Creek (MWH-017). By contrast, the highest mercury tissue concentrations were at the two survey sites on Sugar Creek. Mercury concentrations at these sites were approximately 10 times greater than the mercury concentrations at other survey sites.

Table 5-13 and **Table 5-14** present a comparison of the survey sites sampled in 1995, 1996, 1997, and 2016. Comparisons between these years do not indicate any temporal trends that are consistent for all sampled metals and across all sample locations. Differences in concentrations of arsenic and antimony for these years are noted below. Comparisons for mercury are not possible, due to a lack of sampling at many sites in 1995 and 1996, as well as results that were below detection limits at many sites in 1997. Concentrations of arsenic were higher at most sites in 2016 than they were in 1997. The exceptions were MWH-012 and MWH-029, which had slightly lower arsenic concentrations in 2016 relative to 1997. Maximum arsenic concentrations were noted in either 1995 (four sites) or 1996 (four sites). Similarly, maximum antimony concentrations were noted in either 1995 (two sites) or 1996 (six sites). Relative to 1997, antimony concentrations were higher in 2016 at six of the nine sites (the exceptions were MWH-011, MWH-014, and MWH-075, which were lower in 2016 relative to 1997).

As presented in **Table 5-13**, concentrations of arsenic in macroinvertebrate tissue at all sites other than MWH-013 (EFSFSR), MWH-029 (Sugar Creek), and MHW-017 (Tamarack Creek) were greater than the dietary toxicity levels for fish reported in the Biological Opinion for the water quality toxics standards for Idaho (NMFS 2014). In many cases, the concentrations are extremely high. For example, at MWH-074 (EFSFSR) the concentration was over 40 times higher than levels at which effects to fish may occur. By contrast, concentrations of mercury were generally lower than dietary toxicity levels. The exceptions were the sites on Sugar Creek, where mercury concentrations were greater than the dietary toxicity levels reported in NMFS (2014).

NMFS (2014) reports that selenium toxicity has been reported to occur at dietary concentrations (i.e., concentrations in food sources such as macroinvertebrates) of greater than 3 mg/kg dry weight in one study, but at dietary concentrations of greater than 7.6 mg/kg dry weight in another study. This range was used in **Table 5-13**. Concentrations of selenium in macroinvertebrate

tissue at four EFSFSR sites, and one Meadow Creek site, were greater than 3 mg/kg dry weight in 2016, but lower than 7.6 mg/kg dry weight at all other sites.

Table 5-13. Macroinvertebrate Tissue Metals Concentrations

Note: 1 Mercury data in mg/kg wet weight Key: mg/kg = milligrams per kilogram

Creek	MWH Site ID	Antimony				Arsenic				Cadmium			Copper				Iron				
		1995	1996	1997	2016	1995	1996	1997	2016	1995	1996	1997	2016	1995	1996	1997	2016	1995	1996	1997	2016
IEFSFSR	009	4.58	1.2	2.48	8.06	43.8	1117	17.9	100	0.17	0.33	${<}0.1$	0.32	13.7	7.7	14.4	21 .2	1244	3022	490	1600
	007	21	43.7	16.9	27.2	320	333	79.4	331	0.21	0.19	< 0.1	0.14	16.8	22.8	15.3	30.9	4048	4211	1211	3490
	011	14.4	27	12	6.33	49.7	74.9	23	57.8	0.14	< 0.7	0.19	0.34	16	15.6	18.7	15.3	849	1518	744	2130
	1012	57.4	4.9	12.8	13.3	130	75.9	44.5	38.2	0.21	0.41	0.16	0.3	22.8	16.9	23.3	15.3	2086	1599	1137	1720
Meadow ¹⁰¹⁴		36.8	76.8	13.9	.84	209	147	63.5	71.9	0.12	< 0.14	0.11	0.24	18.9	16.3	20	16.	3236	1565	1111	2470
Creek	075	11	23.3	12	0.99	105	291	45	55.5	0.17	< 0.15	0.18	0.25	18.9	20.2	21.3	12.	2533	14975	1329	1930
Sugar	029	.59	$\overline{}$	<0.5	0.39	47.1	$\overline{}$	10.9	9.35	0.18	\sim	< 0.	0.43	14.6	$\overline{}$		13.5	569		246	584
Creek	069	0.7	9.24	< 0.5	.11	5.97	. .54	5.52	24	0.24	0.33	0.22	0.52	16.3	13	19.6	16.5	407	964	433	899
Minimum		0.7	9.24	<0.5	0.39	5.97	.54	5.52	9.35	0.12	< 0.14	< 0.1	0.14	13.7	13		12.	407	964	246	584
<i>I</i> Maximum		57.4	76.8	16.9	27.2	320	117	79.4	331	0.24	0.41	0.22	0.52	22.8	22.8	23.3	30.9	4048	14975	1329	3490

Table 5-14. Comparison of Macroinvertebrate Tissue Metal Concentrations

Table 5-14. Comparison of Macroinvertebrate Tissue Metal Concentrations (*continued***)**

	MWH Site	Lead				Mercury				Selenium				Zinc			
Stream	ID	1995	1996	1997	2016	1995	1996	1997	2016	1995	1996	1997	2016	1995	1996	1997	2016
EFSFSR	009	1.59	2.3	0.5	0.79	0.8	\sim	0.24	0.43	2.09	< 1.64	.69	3.4	145	245	203	234
	007	4.5	3.7	.39	.01	$\overline{}$	$\overline{}$	< 0.25	0.18	3	2.29	2.28	3.4	157	187	179	157
	011	59،،	2.3	0.5	0.48	0.4	$\overline{}$	< 0.25	0.19	2.29	1.9	2.3	3.2	130	190	206	157
	012	17.5	2.98	3.56	0.55	\sim	٠	< 0.25	0.23	2.7	2.03	2.08	2.6	172	203	233	181
Meadow	014	6.47	.68	3.27	0.47	0.59	٠	< 0.25	0.20	3.78	.82	2.88	3.6	132	184	212	133
Creek	075	2.99	1.74	5.5	0.42	0.44	٠	< 0.33	0.18	3.79	.89	2.3	3	139	212	210	129
	029	1.19	\sim	< 0.5	0.31	.09	٠	0.6	2.06	.88	\sim	1.29	2.6	157	٠	234	150
Sugar Creek	069	0.7	< 1.1	< 0.5	0.48	.05	\sim	.64	2.56	1.1	< 1.10	0.69	2.1	165	235	279	155
Minimum		0.7	<1.	< 0.5	0.31	0.4	0	0.24	2.56	1.1	< 1.10	0.69	2.1	130	184	179	129
Maximum		17.5	3.74	5.5	.01	.09	0	.64	0.18	3.79	2.29	2.88	3.6	172	245	279	234

Source: URS 2000 for 1995-1997 data

Note: Concentrations are miligrams per kilogram dry weight

Key: '-' = Not sampled or not analyzed

5.3.3 Fish Tissue

Table 5-15 presents the sample results for metals concentrations in fish tissue. Concentrations of cadmium, silver, and thallium were not detected above the method detection limit (MDL) in fish tissue at any of the sites. In addition, concentrations of chromium, lead, and nickel were only detected above the MDL at a few sites, and even then, at very low concentrations. All other metals were detected at concentrations above the MDL at all sites. Except for mercury, maximum metal concentrations were seen in fish collected in the EFMC and the EFSFSR. However, there was no clear correlation among all metals relative to any sample site (i.e., no one site was elevated for all metals). Concentrations of mercury, antimony, and arsenic did trend higher at the Yellow Pine pit, MWH-011 and MWH-026 sites (all on the EFSFSR). However, concentrations of mercury, antimony, and arsenic were below literature-derived effects thresholds at all sites, except MWH-018 (Sugar Creek). The concentration at MWH-018 was 0.202 mg/kg wet weight, which is essentially at the minimal effects threshold considered in NMFS 2014. The concentration of aluminum at MWH-027 (EFMC) was 16.171 in sample 2G, which is above the lowest observed effects concentration (LOEC) reported in EPA 2015. However, the other two fish sampled from the EFMC had concentrations well below the LOEC.

Metals concentrations in fish tissue from a 1997 study are reported in URS 2000. However, it is unclear how comparable the data is due to differences in data collection (i.e., they analyzed fish fillets and body remains separately, and calculated whole-body concentrations) and sample locations. Therefore, the results are not compared in this baseline report, but the reader is referred to URS 2000 for information on the 1997 data.

Table 5-15. Fish Tissue Metals Concentrations

Note: Effects thresholds obtained from literature derived values in EPA 2015, EPA 2016a, and NMFS 2014.

Shaded cells are higher than the applicable threshold. Bold text cells are the maximum concentrations for each metal.

1 Laboratory values were in mg/kg dry weight, but were converted to mg/kg wet weight for comparison with the effects thresholds, which are typically in mg/kg wet weight. Values were converted using the equation: wet weigh Percent moisture used in the equation was 77.54, which is the value reported for muscle in EPA 2016b. Concentrations below the MDL were not converted to wet weight and the MDLs are as noted below: a Below MDL of 0.7 b Below MDL of 0.046 c Below MDL of 0.11 d Below MDL of 0.039 e Below MDL of 0.15 f Below MDL of 0.035 g Below MDL of 0.055

Key: EFMC = East Fork Meadow Creek EFSFSR = East Fork of the South Fork of the Salmon River mg/kg = milligrams per kilogram WCT = Westslope cutthroat trout

5.4 Fish

The results of the fisheries' studies are described below.

5.4.1 Genetics Studies

2015 – Fish Tissue 5.4.1.1

The tissues collected in 2015 for genetic testing was for determining the level of hybridization between westslope cutthroat trout and *O. mykiss,* as well as distinguishing between bull trout and brook trout. The 2015 genetics-study results showed that most field identifications of fish species were correct, and most of the genetic samples for westslope cutthroat trout and *O. mykiss* were westslope cutthroat trout. **Table 5-16** and **Appendix 5** includes the results of the genetic survey. The cutthroat x *O. mykiss* hybrids are mostly second- and third-generation backcrosses to cutthroat trout.

All the bull trout samples indicated genetically pure bull trout, based on analyses of six diagnostic microsatellite alleles. A seventh allele—that is typically diagnostic—was variable, but the absence of hybridization evidence in the other alleles suggests that this is natural genetic variation, not the result of hybridization. Results are shown in **Appendix 5**.

Creek	MWH Site ID	Number of Cutthroat Trout	Number of O. mykiss	Number of Hybrids	Number of Misidentifications	Notes			
EFSFSR	011	8	$\mathbf 0$	4	\mathfrak{Z}	Hybrid alleles predominantly westslope cutthroat trout Misidentified fish were hybrids and labeled as O. mykiss			
Meadow Creek	014	5	$\mathbf 0$	$\overline{2}$	$\mathbf{1}$	Hybrids alleles predominantly westslope cutthroat trout Misidentified fish was a hybrid and labeled as O. mykiss			
Meadow Creek	016	19	$\mathbf 0$	1	$\overline{2}$	Hybrid alleles predominantly westslope cutthroat trout One misidentified fish was a hybrid and labeled as bull trout One misidentified fish was not a hybrid and mislabeled as O. mykiss			
Sugar Creek	018	$\mathbf{1}$	10	0	$\mathbf 0$				
Cinnabar Creek	019	$\overline{0}$	1	0	$\mathbf 0$				
EFSFSR	025	67	$\mathbf 0$	$\mathsf{O}\xspace$	$\mathbf 0$				
EFSFSR	026	24	$\mathbf 0$	1	1	Hybrid alleles predominantly O. mykiss Hybrid was labeled as bull trout			
Meadow Creek	034	14	$\mathbf 0$	$\mathbf 0$	\mathfrak{Z}	Misidentified fish were labeled as O. mykiss			
Meadow Creek	047	$\overline{4}$	$\overline{0}$	0	$\overline{0}$				
EFMC	N/A	\mathfrak{Z}	$\mathbf 0$	$\boldsymbol{0}$	$\boldsymbol{0}$				
EFSFSR	N/A	6	Ω	$\overline{2}$	$\overline{0}$	One hybrid had predominantly westslope cutthroat trout alleles One hybrid had predominantly O. mykiss alleles			

Table 5-16. Results of the 2015 Genetic Study

Key:

EFMC = East Fork Meadow Creek

EFSFSR = East Fork of the South Fork of the Salmon River

2016 – eDNA 5.4.1.2

The 2016 eDNA results indicate that Midnight Creek, upper Garnet Creek, and at least the upper section of the unnamed tributary to the EFSFSR likely do not support fish.

O. mykiss likely do not occur in the EFSFSR (upstream from the Yellow Pine pit), Fiddle Creek, lower Garnet Creek, Meadow Creek, Fern Creek, or the unnamed tributaries to the EFSFSR. No DNA for any species was detected in Midnight Creek or upper Garnet Creek. The results showed only one *O. mykiss* detection in the upper Meadow Creek Lake. DNA from all 28 samples were amplified with the golden trout marker. Had *O. mykiss* DNA been detected above the Yellow Pine pit, the positive eDNA detections upstream from the pit could be either rainbow trout, redband trout, steelhead golden trout, or golden trout x cutthroat trout hybrids. Because the eDNA study did not show *O. mykiss* DNA upstream from the Yellow Pine pit (apart from upper Meadow Creek Lake), formerly identified *O. mykiss* in the snorkel surveys were changed to cutthroat trout. eDNA samples collected by the Nez Perce in a separate study showed a positive detection for *O. mykiss* in upper meadow section of the EFMC.

The eDNA results also show that westslope cutthroat trout occur throughout the aquatics baseline study area, in nearly every collection site except for Midnight Creek, upper Garnet Creek, upper Fiddle Creek, and the upper site in the unnamed tributary to the EFSFSR. Bull trout were also detected in most of the creeks sampled, but with a more limited range than the westslope cutthroat trout.

Bull trout were found in all sites of the EFSFSR except for the most upstream locations, in the in the lower section of Fiddle Creek closest to the EFSFSR confluence and throughout Meadow Creek except for upper Meadow Creek Lake. They were not detected in the EFMC nor the unnamed tributary to the EFSFSR or EFMC. They also did not occur in Midnight Creek, nor upper Garnet Creek.

Neither brook trout nor Pacific lamprey were detected at any sampled location. **Figure 5-8** through **Figure 5-13** show the detections/non-detections for each sampled site, for all the species identified. Full eDNA results are provided in **Appendix 5**.

The 11 trout that were collected in Meadow Creek Lake were identified as *O. mykiss*, with no subspecies listed.

Figure 5-8. Chinook Salmon eDNA Results

Figure 5-9. Westslope Cutthroat Trout eDNA Results

Figure 5-10. Oncorhynchus mykiss eDNA Results

Figure 5-11. Bull Trout eDNA Results

Figure 5-12. Brook Trout eDNA Results

Figure 5-13. Pacific Lamprey eDNA Results

5.4.2 Population Abundance Study

A population abundance study was conducted in 2015, to obtain a rigorous estimate of fish abundance and size-class structure, characterizing fish populations within the aquatic resources study area. This population abundance study also intended to determine whether efficiency estimates vary substantially by site, and if the measured habitat characteristics were correlated to the efficiency estimates. If efficiency rates vary substantially by site, or with habitat characteristics, then adjusting population estimates based on these variables may be necessary.

The efficiency evaluation, requiring calculations for full comparisons between the counts made in snorkel surveys to those with the combination of the snorkel and electrofishing efforts, has not yet been finalized. These results will be presented in an addendum report.

5.4.3 Snorkel Surveys

2016 Visual Survey 5.4.3.1

Attempts to conduct fish observations were made at two survey sites for each of three locations: Fiddle Creek, Hennessy Creek and Midnight Creek, that are all small tributaries to the EFSFSR (**Table 3-8**). Only two of the six sites were suitable for snorkeling, and three sites were suitable for videography. One site, lower Hennessy Creek, was too shallow, narrow and steep to be able to snorkel or record on video. In Fiddle Creek (site 425-1), one 100-mm cutthroat trout was observed during a snorkel survey. At the second Fiddle Creek site (MWH-024), six fish were observed on video recordings. It is likely that all the fish were cutthroat trout. No fish were observed at any other site. This survey activity was conducted in accordance with a mitigation agreement between Midas and the Nez Perce Tribe.

2012 – 2015 Visual Surveys 5,4,3.2

Results of all snorkel-survey fish counts conducted in 2012 through 2014 are presented in **Appendix 6**.

Tamarack Creek

Tamarack Creek (MWH-017) was selected as the control site because of minimal disturbance to the system. As shown in **Appendix 6**, all key fish species were observed at this site during all survey years (for all day dives and a single night dive), which did not include a survey in 2015. Additionally, whitefish were observed during day dives, and sculpin were observed mostly during the night survey. Water temperatures ranged between 10.1°C to 11.3°C for day dives, and was 11.5°C for the night dive. A full breakdown of the fish counts, and basic habitat information, is presented in **Appendix 6**.

Young-of-the-year (YOY) salmonids are often more active at night, when larger fish, that may feed on the YOY, are less active. Additionally, some observers have claimed that bull trout may also be more active at night (Thurow 1994, Berge and Mavros 2001). The 2013 and 2014 night surveys did not indicate an increase in bull trout activity, except for YOY. Overall, MWH-017 did experience an increase in YOY cutthroat trout and bull trout that were observed in the 2014 night survey.

Both *O. mykiss* and cutthroat trout were the primary species observed during day dives; however, cutthroat trout (primarily YOY) was the species most observed during the night dive (**Table 5- 17**).

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Table 5-17. Percent of Fish Species Composition Observed in Each Size Class and for Each Species in Tamarack Creek (MWH-017) in 2012- 2014

Note: blank cells indicate no fish observed

Key:

 \le = less than

> = greater than

East Fork of the South Fork of the Salmon River

Multiple sites on the EFSFSR were surveyed in late July and early August, between 2012 and 2014, and three sites upstream from the Yellow Pine pit were surveyed in 2015 (**Figure 3-6** and **Table 3-8**). **Appendix 6** presents the total count of salmonids observed in all survey years for all day dives, and night dives at specific locations. Notations were made when observations of whitefish, sculpin and other species were observed. It was noted that sculpin and tadpole observations increased at night. Tadpoles (mostly tailed-frog) were very abundant, particularly at night.

Water temperatures were recorded prior to each snorkel survey, and ranged from 8.3^oC to 15.6^oC in 2012; 9.3°C to 14°C in 2013; and 11.4°C to 13.9°C for the day dives, and 9°C to 14°C for the night dives in 2014 (**Appendix 6**). Water temperatures in 2015 were recorded every 15 minutes with a data logger attached to the block nets, and temperatures ranged between 9°C to 12°C during these dives.

Fish observations at all EFSFSR sites, both upstream and downstream from the Yellow Pine pit, remained relatively consistent in the first two survey years at each site, with Chinook salmon dominating the species observed. The high counts of Chinook salmon were primarily because of the translocation of spawning adult Chinook salmon in Meadow Creek, with the resulting fish observed being the abundant fry in concentrated locations at two EFSFSR sites. However, in 2014, only a few Chinook salmon fry migrated downstream from Meadow Creek by the time the surveys were conducted, and therefore the numbers of Chinook salmon fry observed were lower in the EFSFSR sites (**Table 5-18)**.

Chinook salmon was the dominant species observed in the EFSFSR from 2012 through 2014 (**Table 5-18**). However, when the sites affected by the translocation of spawning adult Chinook salmon in Meadow Creek were excluded from the overall count (MWH-022 and MWH-011), cutthroat trout numbers were slightly higher than Chinook salmon.

Snorkel surveys have historically identified *O. mykiss* during surveys in the EFSFSR above the Yellow Pine pit. As stated above in Section 5.4.1.2; 2016 eDNA results indicate that *O. mykiss* likely do not occur in the EFSFSR upstream from the Yellow Pine pit, Fiddle Creek, lower Garnet Creek, Meadow Creek, Fern Creek, or the unnamed tributaries to the EFSFSR. The results showed only one *O. mykiss* detection in upper Meadow Creek Lake. DNA from all 28 samples were amplified with the golden trout marker. If *O. mykiss* DNA were detected above the Yellow Pine pit, it would indicate that positive eDNA detections above the pit could potentially be rainbow trout, redband trout, steelhead, golden trout, or golden trout x cutthroat trout hybrids. Because the eDNA study did not show *O. mykiss* DNA upstream from the Yellow Pine pit, apart from the one result in upper Meadow Creek Lake, formerly identified *O. mykiss* in the snorkel surveys have been changed to cutthroat trout in all data tables.

The sites downstream from the Yellow Pine pit had a higher diversity of species and age classes than the upstream sites, including non-target species (whitefish and sculpin). Primarily, the sites upstream from the confluence with Meadow Creek (MWH-013, MWH-025, MWH-044, and MWH-026) only had cutthroat trout and bull trout, except in 2012, when Chinook salmon were observed at MWH-013 and MWH-025.

Table 5-18 summarizes size distribution and percent of species composition for fish observed in the EFSFSR across all survey years.

Note:

No snorkel surveys were conducted in 2016.

Blank cells indicate no sample was conducted.

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Species composition in MWH-022 and MWH-011 were substantially affected by offspring from the introduction of adult hatchery Chinook salmon in the previous year. If Chinook had not been introduced, cutthroat trout are most abundant.

Key:

> = greater than < = less than

 $D = Day$

 $N =$ Night

NS = Not Surveyed

Sugar Creek

Four sites were surveyed in Sugar Creek in 2012 and 2013, three sites in 2014, and one in 2015. One night dive occurred in 2013. All target salmonid species were observed (**Appendix 6**), as were whitefish and sculpin.

The fish snorkel site MWH-010 overlaps with the stream habitat survey site MWH-010, but the fish snorkel site MWH-029 is located just upstream from the stream habitat survey site MWH-008. At all dive sites, stream width, depth, and gradient were recorded (**Appendix 6**). Water temperatures recorded at the beginning of each dive ranged between 9°C to 13.7°C in 2012; between 9.9°C to 14.2°C in the day dives, and was 10.3°C during the night dive in 2013; between 11.4°C to 13.5°C in 2014; and was around 8°C in 2015 (**Appendix 6**).

O. mykiss was the dominant species and Chinook salmon was the least abundant salmonid species observed in Sugar Creek (**Table 5-19**). Most size classes were observed, particularly in MWH-010. Fewer bull trout were observed in the lowest site, and zero or very few Chinook salmon and cutthroat trout were observed in the upstream sites (MWH-018 and MWH-020).

Table 5-19. Size Distribution and Percent of Species Composition for Fish Observed in Sugar Creek from 2012-2015

Table 5-19. Size Distribution and Percent of Species Composition for Fish Observed in Sugar Creek from 2012-2015 (*continued***)**

Note:

Blank cells indicate no fish observations. No snorkel surveys were conducted in 2016.

Key:

 \le = less than

> = greater than

 $D = \overrightarrow{D}ay$

N = Night

Cinnabar Creek

Cinnabar Creek, which has a single survey location, was sampled in 2012 through 2015; no night dives were conducted. Pigment anomalies were observed in multiple fish at this site. Several were generally darker than normal over their entire body, while others had darker patterns only on portions of their bodies. For example, in the 2014 survey, one trout (species unidentifiable) had a black head and a black tail, with a golden body that contained no spots or other pigmentation. While these fish had anomalous pigmentation, they appeared otherwise healthy they swam and acted normally and were observed to be the same size and condition as other fish in the same microhabitat.

Bull trout were the most abundant species observed in all survey years (**Table 5-20** and **Appendix 6**). Water temperatures, recorded at the beginning of each dive, were 10.2°C, 9.1°C, and 11°C in 2012, 2013 and 2014, respectively, and were between 7.5 to 8°C in 2015.

Note:

Blank cells indicate no fish observations. No snorkel surveys were conducted in 2016.

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Key:
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 $>$ = greater than \leq = less than

Cane Creek

Cane Creek had a single survey location that was sampled only in 2012 and 2013; no night dives were conducted at this site. Stream habitat parameters recorded included width, depth and gradient. Bull trout were the only species observed (**Table 5-21** and **Appendix 6**). Water temperatures, recorded at the beginning of each dive, were 11.8°C in 2012, and 10.6°C in 2013.

Table 5-21. Size Distribution and Percent of Species Composition for Fish Observed in Cane Creek (MWH-021) in 2012 and 2013

Key:

 \leq = less than

> = greater than

Fiddle Creek

Survey site MWH-023 in lower Fiddle Creek was snorkeled each year; no night dives were conducted. Fiddle Creek had two survey locations; however, difficulty accessing the upstream site (MWH-024) due to thick vegetative overgrowth made sampling a challenge, resulting in a reduced sampling effort. MWH-024 was only spot-checked in 2012, and then video recorded in 2016.

Only cutthroat trout were observed in Fiddle Creek (**Table 5-22** and **Appendix 6**). Water temperatures recorded at the beginning of each annual dive, from 2012 to 2014, were 10.2°C, 10.7°C and 11.7°C, respectively.

Table 5-22. Size Distribution and Percent of Species Composition for Fish Observed in Fiddle Creek (MWH-023) in All Survey Years

Key:

 \leq = less than

> = greater than

Meadow Creek

A total of six sites were surveyed in Meadow Creek between 2012 and 2014, with one surveyed at night in 2013, and one in 2014 (**Appendix 6**). MWH-031, sampled in 2012, was intended to coincide with the stream habitat survey site location MWH-014; however, the initial site location received from the PAF placed the site too far downstream. Thus, in 2013, survey site MWH-031 was replaced with MWH-014, to coincide with the stream habitat survey location. Because the sites are close to one another, data collected at MWH-031 was combined with that collected at MWH-014.

Water temperatures recorded at the beginning of each day dive in 2012 ranged from 10.9 °C to 14.8°C; in 2013 the range was between 10.3°C to 15.5°C; and in 2014, it ranged from 10.2°C to 16.7°C. Night dives began when water temperatures were 11.1°C (MWH-047) and 9.7°C (MWH-015) (**Appendix 6**). Water temperatures during the 2015 dives ranged between 10°C to 11°C at MWH-047, MWH-016 and MWH-034. Water temperatures at MWH‐014 were highly variable during the survey, ranging between 15°C to 16.8°C (**Appendix 3**). It is unclear as to why the temperatures were so variable. One possibility is that the thermograph was placed in an area that was periodically exposed (e.g., it came out of the water when the blocknet moved with the flow, or if the thermograph was placed too high on the net).

In 2012, 559 salmonids were counted during dives conducted during the daytime (**Appendix 6**). The 2013 survey had the highest salmonid observations, totaling 628 salmonids in the day dives, and 38 during the night dive. Of the 628 salmonids recorded, 567 of those were Chinook salmon fry at MWH-014. Tailed-frog tadpoles were common in Meadow Creek, and the observations of them increased during the night dives. Tailed-frog tadpoles were abundant, particularly at MWH-015.

In 2014, a total of 398 salmonids were observed, including nine during the night survey at sites MWH-014 and MWH-015. As with the other sites, tailed-frog tadpoles were common in Meadow Creek, and observations of them increased during the night dives.

During the 2015 population abundance survey, only 26 fish were observed in the Meadow Creek sites combined. Chinook salmon were not translocated to Meadow Creek in 2014, which accounted for the lower fish abundance at MWH-014 and MWH-047 in 2015. The abundance of Chinook salmon in the lower sites (MWH-031 and MWH-014) was strongly affected by the translocation of adult Chinook salmon. Subsequent surveys' counts included the offspring from these released fish. Therefore, the dominant species observed were fry and juvenile Chinook salmon (**Table 5-23** and **Appendix 6**). Yearling Chinook salmon were observed in both 2013 and 2014 in MWH-015 and MWH-047, both of which are located upstream from a long and very high gradient (greater than 5 percent) riffle. These yearlings are the offspring of the released Chinook salmon spawners.

Table 5-23 presents the distribution and percent of species composition for fish observed in Meadow Creek across all survey years.

Table 5-23. Size Distribution and Percent of Species Composition for Fish Observed in All Meadow Creek Sites in 2012-2015, and All Years Combined

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Table 5-23. Size Distribution and Percent of Species Composition for Fish Observed in All Meadow Creek Sites in 2012-2015, and All Years Combined (*continued***)**

Note:

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Blank cells indicate no fish observations.

No snorkel surveys were conducted in 2016.

Species composition in MWH-031and MWH-014 were substantially affected by offspring from the introduction of adult hatchery Chinook salmon in the previous year. If Chinook had not been introduced, cutthroat are the most abundant.

Key:

> = greater than

< = less than

 $D = day$

 $N =$ night NS = not surveyed

East Fork Meadow Creek

Two EFMC sites were snorkeled between 2012 and 2014, including one night dive in 2014. MWH-027 is in a large meadow, upstream from a high-gradient fish barrier created by a 1965 dam failure. The downstream end of MWH-028 is located just upstream from EFMC's confluence with Meadow Creek.

Water temperatures at the beginning of each day dive were: 10.4°C and 11.1°C in 2012; 13.4°C and 14.9°C in 2013; and 14.2°C and 12.7°C in 2014 for MWH-028 and MWH-027, respectively. The water temperature recorded at the beginning of the 2014 night dive at MWH-027 was 11°C.

Fish abundance was affected by the spawning of released Chinook salmon in lower Meadow Creek because the fry from these spawners rear in the lower section of the EFMC.

Species composition in the lower portion of the EFMC is dominated by the offspring of the translocated Chinook salmon which use this stretch as rearing habitat. Without the translocation of Chinook salmon, cutthroat are the most abundant species. **Table 5-24** lists the size and species composition for fish observed in the EFMC.

Table 5-24. Size Distribution and Percent of Species Composition for Fish Observed in All East Fork Meadow Creek Sites for 2012- 2014

Note:

Blank cells indicate no fish observations at that site.

No snorkel surveys were conducted after 2014.

Species composition in MWH-027 was substantially affected by offspring from the introduction of adult hatchery Chinook salmon in the previous year. If Chinook had not been introduced, cutthroat are most abundant.

Key:

> = greater than

 \leq = less than

 $D = day$

 $N =$ night

5.4.4 Discussion

The translocation of spawning adult Chinook salmon to Meadow Creek by IDFG had a distinct effect on species diversity in the EFSFSR. This played a primary role between Meadow Creek and the Yellow Pine pit (at MWH-022 and MWH-011, but also in 2012 at MWH-013 and MWH-025); at the lower Meadow Creek survey sites (MWH-014, MWH-015 and MWH-047); and in the EFMC downstream site (MWH-028). Species and size diversity are greater in the sites downstream than those found in sites upstream from the Yellow Pine pit.

Figure 5-14 through **Figure 5-17** show size distribution for each of the targeted fish species, along with the total number of fish observed. Most of the Chinook salmon were in the smallest size category—less than 80 mm for Chinook salmon. Substantially more large bull trout were observed in 2013 and 2014, possibly due to a larger number of survey sites in the lower EFSFSR for those years. Multiple size classes that are present in a river or stream indicate suitable conditions to support all life stages.

While the surveys did not occur on the same day each year for each site, they all were conducted between late July and mid-August, and were only off by a matter of days to just a few weeks. Because the timing of the surveys is so similar between years, the size classes of each species observed between years is also similar. The different size classes tend to represent different age classes, particularly for resident forms as compared to anadromous forms. Additionally, cooler water temperatures—as occurs in the aquatic resources study area—result in overall slower growth rates. Therefore, the size classes between survey years are comparable.

Note: Number above bars represents total number of fish observed

Key: > = greater than \leq = less than

% = percent

mm = millimeters

Figure 5-14. Chinook Salmon Size Distribution for All Survey Sites Combined (Top Graph), and All Sites Combined, Except Those Affected by the Chinook Salmon Translocation (Bottom Graph)

Note: Number above bars represents total number of fish observed

Key:

- > = greater than
- < = less than
- % = percent
- mm = millimeters

Figure 5-15. Oncorhynchus mykiss Size Distribution for All Survey Sites Combined

Note: Number above bars represents total number of fish observed Key:

> = greater than

< = less than

% = percent

mm = millimeters

Figure 5-16. Bull Trout Size Distribution for All Survey Sites Combined

Note: Number above bars represents total number of fish observed Key: > = greater than

 \leq = less than % = percent mm = millimeters

Figure 5-17. Cutthroat Trout Size Distribution for All Survey Sites Combined

5.5 Aquatics Surveys Summary

This baseline study report characterizes the existing conditions of aquatic resources in the Stibnite Gold project area between 2012 and 2016. The study is based on 4 years of stream habitat survey data, 4 years of macroinvertebrate survey data, 2 years of metals data, 3 years of protocol-level snorkel data, 1 year of 4-pass depletion electrofishing data collected for obtaining unbiased population estimates, 1 year of non-protocol level snorkel survey data or video recordings, and 2 years of genetic data. The combination of these baseline surveys was necessary to characterize the overall condition of the system. Understanding the habitat conditions (e.g., spawning gravel, LWD); water quality conditions (i.e., water temperature, macroinvertebrate taxa diversity and abundance); sediment conditions (i.e., metal content); and biological conditions (i.e., macroinvertebrates and fish distribution and abundance) in combination provides the basis for determining the basin condition.

This report provides an overview of existing stream habitat and biological conditions from data collected on important indicators such as: stream substrate, LWD, water temperature, macroinvertebrate taxa diversity and abundance, sediment and macroinvertebrate metal content, and fish distribution and abundance.

Together, the aquatic resource baseline data permit several general conclusions about the condition of the habitat and biological resources:

• The average streambed interstitial conditions, as measured by both cobble embeddedness and free matrix are considered *functioning appropriately* at nearly all survey sites. Five

sites, including in Meadow Creek, Sugar Creek, Profile Creek, and EFMC, however, are considered *functioning at risk* when averaged over all survey years.

- Average surface fines, as measured by free matrix, are generally considered *functioning at risk* or *functioning at unacceptable risk* at most surveyed sites, however, four sites in the EFSFSR, Burntlog Creek, and Trapper Creek, which are all granitic streambeds, are *functioning appropriately* based on annual means.
- The majority of percent depth fines less than 6.33 mm, as measured by core sampling in the EFSFSR, are considered *functioning appropriately* and the percent has decreased each year since 2013. All annual measurements of percent depth fines less than 0.85 mm, as measured by core sampling, are less than the Jensen et al. (2009) suggested threshold of 10 percent. Jensen et al. indicated odds of salmon and steelhead egg-to-fry survival dropping dramatically above the 10 percent threshold.
- Stream habitat conditions may change over time from factors such as seasonal events (e.g., high runoff flows), fire, weather conditions, erosion, and federal management activities (e.g. road decommissioning). For example, in early 2014, the watershed experienced multiple avalanches and landslides, and the subsequent substrate measurements showed changes in embeddedness, fine sediment and free matrix.
- Water temperatures do not appear to exceed the WCIs substantially enough to impair fish production.
- Macroinvertebrate surveys show that the macroinvertebrate assemblages at all 11 survey sites are generally indicative of high water quality and relatively stable habitat. This is supported by high taxa richness, the presence of many predator and long-lived taxa and taxa that require clean cobble substrates, high proportions of taxa that are intolerant of poor water quality, and low numbers of individuals that are tolerant of poor water quality.

Notable differences in macroinvertebrate composition, as compared between survey sites, included a larger proportion of filterers (in this case blackfly larvae) just downstream of the Yellow Pine pit outlet, somewhat greater numbers of stoneflies at the upstream Sugar Creek and Tamarack Creek sites, and a decrease in metals tolerance index values within the Meadow Creek constructed channel, indicating that metals contamination may be decreasing at that site. PIBO O/E scores among survey sites from 2012 to 2016 were consistently rated as being in *good* biological condition, except in three instances where a site was rated as *fair*. The O/E scores indicated that upper Sugar Creek, the Meadow Creek confluence, and the Meadow Creek constructed channel were initially assessed to be in *fair* condition (2012), but had improved to *good* condition in subsequent sampling years. All other sites were in *good* condition from 2012 to 2014, and in 2016.

- Metals in the sediment—particularly antimony, arsenic, and mercury—frequently exceeded the NMFS and EPA benchmarks at all sampled locations, often by one or more orders of magnitude.
- Concentrations of metals, particularly arsenic in macroinvertebrate tissue, were elevated at sites on the EFSFSR and Sugar Creek, relative to concentrations at other sites particularly those on Meadow Creek and Tamarack Creek. Mercury concentrations in macroinvertebrates were substantially elevated in Sugar Creek. Arsenic and antimony occurred in lower concentrations in sites farther upstream in the EFSFSR, and in Meadow Creek and Sugar Creek. In comparison to previous studies, arsenic concentrations

decreased in all sites except for one Sugar Creek site; however, no clear temporal trend was indicated.

- Except for mercury, fish collected in the EFMC and EFSFSR had maximum metal concentrations, but no one site was elevated for all metals. Concentrations of mercury, antimony and arsenic did trend higher at the Yellow Pine pit, at MWH-011, and MWH-026 (which are all on the EFSFSR). However, concentrations of mercury, antimony and arsenic were below literature-derived effects thresholds at all sites, except for MWH-018 (Sugar Creek).
- Fish distribution and abundance were relatively similar in each survey year, except for the sites directly affected by IDFG's translocation or introduction of adult spawning Chinook salmon in Meadow Creek. Most size classes of all target species were represented throughout the aquatic resources study area, which is indicative of relatively stable fisheries for the entire Stibnite Gold project area.
- Genetic studies indicate that *O. mykiss* likely do not occur upstream from the Yellow Pine pit, except for the previously introduced golden trout in the upper Meadow Creek Lake. Westslope cutthroat trout were found in nearly all sites, and Chinook salmon were found only in the samples taken at the Yellow Pine pit, and the EFSFSR downstream from Meadow Creek. Bull trout were detected in the EFSFSR, lower Fiddle Creek Meadow Creek, and the unnamed tributary to Meadow Creek. Brook trout and Pacific lamprey were not detected in any tested sample, and likely do not occur upstream from the Yellow Pine pit.

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6.2 Glossary

6.3 Abbreviations and Acronyms

SECTION 7 LIST OF PREPARERS

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Appendix 1:

Stream Habitat Survey Results for 2012 -2016 – Statistical Tables, Box and Whisker Plots and ANOVA Tables for Applicable Variables

Appendix 1: Stream Habitat Survey Results for 2012 to 2016 –Box and Whisker Plots, Statistical Tables, and ANOVA Tables for Applicable Variables

Stibnite Gold Project Midas Gold Idaho, Inc.

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> > **April 2017**

Prepared by

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SECTION 1: BOX AND WHISKER PLOTS

A "box and whisker plot" is a way to graphically show where the middle (i.e., median) of the data lies within a dataset. The data is broken into quartiles, each with an equal number of data values. Box and whisker plots display variation in samples of a statistical population without making any assumptions of the underlying statistical distribution. All whiskers are limited to 1.5 times the interquartile range (IQR); this is 1.5 times the distance from quartile to quartile (the distance within the data covered by the boxes). In addition to median, the average or mean is presented for each box and whisker plot with a dashed line. Data that fall outside of the 1.5 IQR appears as a single dot and are considered an outlier. If most of the data is clustered or valued as "0" then no shaded block will appear on the plot. **Figure 1-1** shows the various elements of the box and whisker plots presented in this appendix. Definitions of specific box and whisker components are as followed:

- Median Represents the mid-point of the dataset and is shown by the line that divides the box into two parts. Fifty percent of the data is greater than or equal to the median and the remaining half are less.
- Mean Represents the arithmetic average value of the dataset and is shown by a horizontal dotted line.
- Upper Quartile (Q3) Seventy-five percent of the observed values fall below the upper quartile and twenty-five percent above.
- Lower Quartile $(Q1)$ Twenty-five percent of the observed values fall below the lower quartile and seventy-five percent above.
- Interquartile Range (IQR) Represents the middle fifty percent of the dataset and is defined by the range of the upper and lower quartiles.
- Whiskers The upper and lower whiskers extend to the maximum values within 1.5 times the IQR.
- Outlier Represents observed values that are either greater than or less than 1.5 times the IQR. These values extend beyond the whiskers on the plot.

The wetted channel width-to-wetted stream depth (WW:WD), maximum pool depth, bank angle, bankfull width, maximum bankfull depth, free matrix, surface fines (as measured from free matrix), cobble embeddedness, and depth fines (as measured from core sampling) were evaluated with the box and whisker plots. The results are presented below.

Figure 1-1. Box and Whisker Example Plot

1.1 PIBO Habitat Survey Plots

Methods employed for habitat surveys included habitat sampling following the PAF modified protocol for the Pacific Anadromous Fish Strategy (PACFISH) Inland Fish Strategy (INFISH) Biological Opinion Effectiveness Monitoring Program (PIBO) (USFS 2005, 2013a). PIBO was developed in response to monitoring needs addressed in the Biological Opinions for bull trout (USFWS 1998) and steelhead (NMFS 1998). It provides a consistent framework for monitoring aquatic and riparian resources within the range of the PACFISH/INFISH, to help determine whether land management practices are maintaining or improving riparian and aquatic conditions at both the landscape and watershed scales on federal lands throughout the Upper Columbia River Basin. The results are presented in **Figure 1-2** through **Figure 1-37**.

Dashed lines in each box plot indicate sample mean Key: WW: wetted width of channel
WD: wetted depth of channel

wy.
WW: wetted width of channel
WD: wetted depth of channel

Key:

WW: wetted width of channel WD: wetted depth of channel

Figure 1-4. Box and Whisker Plot for Wetted Width-to-Wetted Depth Ratio for all Sites in 2014

Key:
WW: wetted width of channel
WD: wetted depth of channel

Dashed lines in each box plot indicate sample mean Key: WW: wetted width of channel WD: wetted depth of channel

Figure 1-6. Box and Whisker Plot for Wetted Width-to-Wetted Depth Ratio for all Subwatersheds in 2012

WW: wetted width of channel WD: wetted depth of channel

Figure 1-8. Box and Whisker Plot for Wetted Width-to-Wetted Depth Ratio for all Subwatersheds in 2014

WW: wetted width of channel WD: wetted depth of channel

Figure 1-9. Box and Whisker Plot for Wetted Width-to-Wetted Depth Ratio for all Subwatersheds in 2015

Figure 1-10. Box and Whisker Plot of the Bank Angles for all Survey Sites in 2014

Figure 1-11. Box and Whisker Plot of the Bank Angles for all Survey Sites in 2015

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Figure 1-14. Box and Whisker Plot of the Maximum Pool Depths for all Survey Sites in 2012

Dashed lines in each box plot indicate sample mean

Figure 1-15. Box and Whisker Plot of the Maximum Pool Depths for all Survey Sites in 2013

Figure 1-16. Box and Whisker Plot of the Maximum Pool Depths for all Survey Sites in 2014

Figure 1-17. Box and Whisker Plot of the Maximum Pool Depths for all Survey Sites in 2015

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Dashed lines in each box plot indicate sample mean

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Figure 1-23. Box and Whisker Plot of the Bankfull Widths for all Survey Sites in 2013

Figure 1-24. Box and Whisker Plot of the Bankfull Widths for all Survey Sites in 2014

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Figure 1-26. Box and Whisker Plot of the Bankfull Widths for all Subwatersheds in 2012

Figure 1-28. Box and Whisker Plot of the Bankfull Widths for all Subwatersheds in 2014

Figure 1-29. Box and Whisker Plot of the Bankfull Widths for all Subwatersheds in 2015

Dashed lines in each box plot indicate sample mean

Figure 1-30. Box and Whisker Plot of the Maximum Bankfull Depths for all Survey Sites in 2012

Notes:

Dashed lines in each box plot indicate sample mean

Figure 1-36. Box and Whisker Plot of the Maximum Bankfull Depths for all Subwatersheds in 2014

1.2 Free Matrix and Surface Fines Plots

Figure 1-38 through **Figure 1-53** are box and whisker plots showing free matrix and surface fines results for all sites and subwatersheds. Free matrix measures the amount of loose substrate and surface fine sediment.

Site

Percent fine sediment is measured in the free matrix survey Dashed lines in each box plot indicate sample mean

Figure 1-38. Box and Whisker Plot of the Percent Fines for all Sites in 2013

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Percent fine sediment is measured in the free matrix survey Dashed lines in each box plot indicate sample mean

Figure 1-40. Box and Whisker Plot of the Percent Fines for all Sites in 2015

Note:

Percent fine sediment is measured in the free matrix survey

Dashed lines in each box plot indicate sample mean

Figure 1-41. Box and Whisker Plot of the Percent Fines for all Sites in 2016

Figure 1-42. Box and Whisker Plot of the Percent Fines for all Subwatersheds in 2013

Percent fine sediment is measured in the free matrix survey Dashed lines in each box plot indicate sample mean

Figure 1-43. Box and Whisker Plot of the Percent Fines for all Subwatersheds in 2014

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Figure 1-51. Box and Whisker Plot of the Percent Free Matrix for all Subwatersheds in 2014

Figure 1-52. Box and Whisker Plot of the Percent Free Matrix for all Subwatersheds in 2015

Figure 1-53. Box and Whisker Plot of the Percent Free Matrix for all Subwatersheds in 2016

1.3 Cobble Embeddedness Plots

Cobble embeddedness measures the amount of cobble material that is buried in fine materials. Cobble embeddedness surveys were conducted by the PAF in 2012. The results from both PAF and MWH survey data are presented in **Figures 1-54** through **Figure 1-61**.

1-56

Note:

Figure 1-56. Box and Whisker Plot of the Cobble Embeddedness for all Sites in 2015

Note:

Figure 1-57. Box and Whisker Plot of the Cobble Embeddedness for all Sites in 2016

Dashed lines in each box plot indicate sample mean

Figure 1-59. Box and Whisker Plot of the Cobble Embeddedness for all Subwatersheds in 2014

Dashed lines in each box plot indicate sample mean

Figure 1-60. Box and Whisker Plot of the Cobble Embeddedness for all Subwatersheds in 2015

Dashed lines in each box plot indicate sample mean

Figure 1-61. Box and Whisker Plot of the Cobble Embeddedness for all Subwatersheds in 2016

1.4 Core Sampling (Depth Fines) Plots

The modified McNeil core substrate sampling measures the proportion of the subsurface sediment that are fine materials, which are called depth fines. Percentage of depth fines is an indicator of spawning substrate quality for a reach. The results are presented in **Figure 1-62** and **Figure 1-63**.

1.5 Pool Frequency

Pool frequency is the number of pools per 100 meters of stream at a given site. There is only one data point per PIBO survey, which is insufficient data to develop box-and-whisker plots disaggregated by year and by site or subwatershed.

1.6 Large Woody Debris Frequency

Large Woody Debris (LWD) frequency is the number of LWD pieces per 100 meters of stream at a given site. There is only one data point per PIBO survey, which is insufficient data to develop box-and-whisker plots disaggregated by year and by site or subwatershed.

1.7 Median Particle Size – D₅₀

D50 is the median particle size of a reach, as calculated from PIBO survey pebble count data. There is only one data point per PIBO survey, which is insufficient data to develop box-andwhisker plots disaggregated by year and by site or subwatershed.

1.8 Percent Fine Sediments

Percent of fine sediments is the proportion of pebbles counted during PIBO survey pebble counts that was smaller than 6 millimeters (mm). There is only one data point per PIBO survey, which is insufficient data to develop box-and-whisker plots disaggregated by year and by site or subwatershed. These data are not the same as the surface fine sediment (measured in the free matrix surveys) or depth fines (measured in the modified McNeil core sampling surveys).

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PIBO SUMMARY STATISTIC RESULTS TABLES SECTION 2:

Tables 2-1 through 2-60 summarizes results of the PIBO surveys.

Table 2-1. Summary Statistics for Wetted Width: Wetted Depth Ratio for all Sites in 2012

Site Number	Sample Size (n)	Annual Mean	Standard Deviation	Standard Error	95% Confidence Interval	Coefficient of Variation	Geologic Type
MWH-006	22	43.9	25.93	5.53	±10.83	0.59	Granitic
MWH-016	\overline{O}	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-044	21	18.9	6.54	1.43	±2.8	0.35	Granitic
MWH-047	21	15.0	8.62	1.88	±3.69	0.58	Granitic
MWH-050	0	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-061	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-060	\overline{O}	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-059	\overline{O}	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-057	0	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-056	\overline{O}	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-055	0	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-053	\overline{O}	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-033	23	71.6	42.81	8.93	±17.5	0.60	Non-granitic
MWH-032	24	61.4	32.44	6.62	±12.98	0.53	Non-granitic
MWH-017	22	38.5	24.07	5.13	±10.06	0.63	Non-granitic
MWH-014	21	38.6	16.27	3.55	±6.96	0.42	Granitic
MWH-013	21	39.1	20.64	4.50	±8.83	0.53	Granitic
MWH-010	21	46.1	25.73	5.61	±11	0.56	Non-granitic
MWH-009	21	51.0	22.21	4.85	± 9.5	0.44	Granitic
MWH-008	21	35.8	12.21	2.67	± 5.22	0.34	Non-granitic

Table 2-2. Summary Statistics for Wetted Width:Wetted Depth Ratio for all Sites in 2013

Site Number	Sample Size (n)	Annual Mean	Standard Deviation	Standard Error	95% Confidence Interval	Coefficient of Variation	Geologic Type
MWH-006	22	43.3	20.16	4.30	± 8.42	0.47	Granitic
MWH-016	23	43.5	23.05	4.81	±9.42	0.53	Granitic
MWH-044	21	24.1	11.31	2.47	±4.84	0.47	Granitic
MWH-047	25	18.1	10.23	2.05	±4.01	0.56	Granitic
MWH-050	26	49.4	29.44	5.77	±11.32	0.60	Granitic
MWH-061	\overline{O}	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-060	\overline{O}	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-059	\overline{O}	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-057	24	56.5	33.01	6.74	±13.21	0.58	Granitic
MWH-056	25	47.4	17.05	3.41	±6.68	0.36	Granitic
MWH-055	21	29.6	19.11	4.17	±8.17	0.65	Granitic
MWH-053	22	33.6	17.32	3.69	±7.24	0.52	Granitic
MWH-033	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-032	\overline{O}	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-017	0	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-014	\overline{O}	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-013	\overline{O}	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-010	\overline{O}	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-009	\overline{O}	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-008	Ω	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-062	27	36.6	29.73	5.72	±11.22	0.81	Granitic

Table 2-4. Summary Statistics for Wetted Width:Wetted Depth Ratio for all Sites in 2015

Site Number	Sample Size (n)	Annual Mean (pieces per 100m)	Standard Deviation	Standard Error	95% Confidence Interval	Coefficient of Variation	Geologic Type
MWH-006		2.3	N/A	N/A	N/A	N/A	Granitic
MWH-016	$\mathbf{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-044		3.2	N/A	N/A	N/A	N/A	Granitic
MWH-047		30.2	N/A	N/A	N/A	N/A	Granitic
MWH-050	$\mathbf 0$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-061	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-060	\overline{O}	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-059	\overline{O}	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-057	$\mathbf{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-056	\overline{O}	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-055	\overline{O}	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-053	$\mathbf 0$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-033		0.5	N/A	N/A	N/A	N/A	Non-granitic
MWH-032		1.8	N/A	N/A	N/A	N/A	Non-granitic
MWH-017		8.2	N/A	N/A	N/A	N/A	Non-granitic
MWH-014		0.0	N/A	N/A	N/A	N/A	Granitic
MWH-013		13.1	N/A	N/A	N/A	N/A	Granitic
MWH-010		9.5	N/A	N/A	N/A	N/A	Non-granitic
MWH-009		2.0	N/A	N/A	N/A	N/A	Granitic
MWH-008		6.7	N/A	N/A	N/A	N/A	Non-granitic

Table 2-6. Summary Statistics for Large Woody Debris Frequency for all Sites in 2013

Site Number	Sample Size (n)	Annual Mean (pieces per 100m)	Standard Deviation	Standard Error	95% Confidence Interval	Coefficient of Variation	Geologic Type
MWH-006	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-016	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-044	\overline{O}	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-047	\overline{O}	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-050		16.7	N/A	N/A	N/A	N/A	Granitic
MWH-061		0.0	N/A	N/A	N/A	N/A	Non-granitic
MWH-060		0.0	N/A	N/A	N/A	N/A	Granitic
MWH-059		19.2	N/A	N/A	N/A	N/A	Granitic
MWH-057		31.4	N/A	N/A	N/A	N/A	Granitic
MWH-056		33.9	N/A	N/A	N/A	N/A	Granitic
MWH-055		16.7	N/A	N/A	N/A	N/A	Granitic
MWH-053		34.7	N/A	N/A	N/A	N/A	Granitic
MWH-033		14.8	N/A	N/A	N/A	N/A	Non-granitic
MWH-032	\mathbf{I}	32.3	N/A	N/A	N/A	N/A	Non-granitic
MWH-017	\overline{O}	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-014	$\mathbf{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-013	\overline{O}	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-010	\overline{O}	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-009	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-008	\overline{O}	N/A	N/A	N/A	N/A	N/A	Non-granitic

Table 2-7. Summary Statistics for Large Woody Debris Frequency for all Sites in 2014

Site Number	Sample Size (n)	Annual Mean	Standard Deviation	Standard Error	95% Confidence Interval	Coefficient of Variation	Geologic Type
MWH-006	4	0.3	0.09	0.04	±0.09	0.35	Granitic
MWH-016	8	0.4	0.08	0.03	±0.06	0.19	Granitic
MWH-044	$\overline{7}$	0.5	0.15	0.06	±0.11	0.32	Granitic
MWH-047	5	0.5	0.27	0.12	±0.23	0.50	Granitic
MWH-050	9	0.6	0.19	0.06	±0.12	0.31	Granitic
MWH-061	0	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-060	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-059	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-057	13	0.3	0.08	0.02	±0.04	0.23	Granitic
MWH-056	5	0.5	0.15	0.07	±0.13	0.29	Granitic
MWH-055	6	0.6	0.20	0.08	±0.16	0.31	Granitic
MWH-053	$\overline{4}$	0.4	0.05	0.03	±0.05	0.12	Granitic
MWH-033	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-032	0	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-017	$\mathbf{0}$	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-014	0	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-013	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-010	$\mathbf 0$	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-009	0	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-008	$\mathbf{0}$	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-062	12	0.3	0.08	0.02	±0.05	0.28	Granitic

Table 2-8. Summary Statistics for Large Woody Debris Frequency for all Sites in 2015

Site Number	Sample Size (n)	Annual Mean	Standard Deviation	Standard Error	95% Confidence Interval	Coefficient of Variation	Geologic Type
MWH-006	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-016	11	0.4	0.10	0.03	±0.06	0.22	Granitic
MWH-044	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-047	0	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-050	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-061	Ω	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-060	0	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-059	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-057	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-056	0	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-055	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-053	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-033	0	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-032	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-017	8	0.5	0.14	0.05	±0.1	0.32	Non-granitic
MWH-014	$\overline{7}$	0.4	0.12	0.05	±0.09	0.29	Granitic
MWH-013	9	0.5	0.16	0.05	±0.11	0.34	Granitic
MWH-010	6	0.5	0.17	0.07	±0.14	0.32	Non-granitic
MWH-009	$\overline{7}$	0.6	0.27	0.10	±0.2	0.43	Granitic
MWH-008	14	0.5	0.19	0.05	±0.1	0.40	Non-granitic

Table 2-9. Summary Statistics for Maximum Pool Depth at all Sites in 2012

Site Number	Sample Size (n)	Annual Mean	Standard Deviation	Standard Error	95% Confidence Interval	Coefficient of Variation	Geologic Type
MWH-006	3	0.3	0.08	0.04	±0.09	0.22	Granitic
MWH-016	0	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-044	6	0.5	0.12	0.05	±0.1	0.24	Granitic
MWH-047	8	0.8	0.21	0.07	±0.15	0.28	Granitic
MWH-050	0	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-061	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-060	0	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-059	0	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-057	$\mathbf 0$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-056	0	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-055	0	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-053	$\mathbf 0$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-033	6	0.7	0.18	0.07	±0.15	0.26	Non-granitic
MWH-032	8	1.2	0.30	0.10	±0.21	0.25	Non-granitic
MWH-017	16	0.5	0.14	0.03	±0.07	0.31	Non-granitic
MWH-014	10	0.5	0.12	0.04	±0.08	0.25	Granitic
MWH-013	$\overline{7}$	0.5	0.13	0.05	±0.1	0.25	Granitic
MWH-010	12	0.5	0.15	0.04	±0.08	0.27	Non-granitic
MWH-009	6	0.7	0.12	0.05	±0.1	0.17	Granitic
MWH-008	18	0.6	0.16	0.04	± 0.07	0.28	Non-granitic

Table 2-10. Summary Statistics for Maximum Pool Depth at all Sites in 2013

Site Number	Sample Size (n)	Annual Mean	Standard Deviation	Standard Error	95% Confidence Interval	Coefficient of Variation	Geologic Type
MWH-006	4	0.3	0.09	0.04	±0.09	0.35	Granitic
MWH-016	8	0.4	0.08	0.03	±0.06	0.19	Granitic
MWH-044	$\overline{7}$	0.5	0.15	0.06	±0.11	0.32	Granitic
MWH-047	5	0.5	0.27	0.12	±0.23	0.50	Granitic
MWH-050	9	0.6	0.19	0.06	±0.12	0.31	Granitic
MWH-061	0	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-060	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-059	$\mathbf 0$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-057	13	0.3	0.08	0.02	±0.04	0.23	Granitic
MWH-056	5	0.5	0.15	0.07	±0.13	0.29	Granitic
MWH-055	6	0.6	0.20	0.08	±0.16	0.31	Granitic
MWH-053	4	0.4	0.05	0.03	±0.05	0.12	Granitic
MWH-033	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-032	0	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-017	$\mathbf{0}$	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-014	0	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-013	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-010	$\mathbf 0$	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-009	0	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-008	$\mathbf{0}$	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-062	12	0.3	0.08	0.02	±0.05	0.28	Granitic

Table 2-12. Summary Statistics for Maximum Pool Depth at all Sites in 2015

Table 2-13. Summary Statistics for Pool Frequency at all Sites in 2012

Site Number	Sample Size (n)	Annual Mean	Standard Deviation	Standard Error	95% Confidence Interval	Coefficient of Variation	Geologic Type
MWH-006	3	2.3	0.00	0.00	N/A	N/A	Granitic
MWH-016	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-044	6	4.8	0.00	0.00	N/A	N/A	Granitic
MWH-047	8	6.3	0.00	0.00	N/A	N/A	Granitic
MWH-050	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-061	Ω	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-060	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-059	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-057	0	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-056	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-055	0	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-053	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-033	6	1.4	0.00	0.00	N/A	N/A	Non-granitic
MWH-032	8	2.4	0.00	0.00	N/A	N/A	Non-granitic
MWH-017	16	8.7	0.00	0.00	± 0	0.00	Non-granitic
MWH-014	10	6.0	0.00	0.00	± 0	0.00	Granitic
MWH-013	$\overline{7}$	4.2	0.00	0.00	± 0	0.00	Granitic
MWH-010	12	7.1	0.00	0.00	± 0	0.00	Non-granitic
MWH-009	6	2.4	0.00	0.00	N/A	N/A	Granitic
MWH-008	18	8.6	0.00	0.00	± 0	0.00	Non-granitic

Table 2-14. Summary Statistics for Pool Frequency at all Sites in 2013

Site Number	Sample Size (n)	Annual Mean	Standard Deviation	Standard Error	95% Confidence Interval	Coefficient of Variation	Geologic Type
MWH-006	0	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-016	0	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-044	0	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-047	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-050	6	2.0	0.00	0.00	N/A	N/A	Granitic
MWH-061	3	1.5	0.00	0.00	N/A	N/A	Non-granitic
MWH-060		0.5	N/A	N/A	N/A	N/A	Granitic
MWH-059	$\overline{2}$	0.8	0.00	0.00	N/A	N/A	Granitic
MWH-057	12	4.5	0.00	0.00	N/A	N/A	Granitic
MWH-056	$\overline{4}$	1.7	0.00	0.00	N/A	N/A	Granitic
MWH-055	8	4.8	0.00	0.00	N/A	N/A	Granitic
MWH-053	6	3.4	0.00	0.00	N/A	N/A	Granitic
MWH-033	3	0.6	0.00	0.00	N/A	N/A	Non-granitic
MWH-032	5	1.7	0.00	0.00	N/A	N/A	Non-granitic
MWH-017	Ω	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-014	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-013	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-010	Ω	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-009	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-008	0	N/A	N/A	N/A	N/A	N/A	Non-granitic

Table 2-15. Summary Statistics for Pool Frequency at all Sites in 2014

Site Number	Sample Size (n)	Annual Mean	Standard Deviation	Standard Error	95% Confidence Interval	Coefficient of Variation	Geologic Type
MWH-006	4	3.0	0.00	0.00	N/A	N/A	Granitic
MWH-016	8	4.3	0.00	0.00	± 0	0.00	Granitic
MWH-044	$\overline{7}$	5.6	0.00	0.00	± 0	0.00	Granitic
MWH-047	5	4.0	0.00	0.00	± 0	0.00	Granitic
MWH-050	9	3.1	0.00	0.00	N/A	N/A	Granitic
MWH-061	Ω	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-060	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-059	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-057	13	4.9	0.01	0.00	±0.01	0.00	Granitic
MWH-056	5	2.2	0.00	0.00	N/A	N/A	Granitic
MWH-055	6	3.6	0.00	0.00	± 0	0.00	Granitic
MWH-053	4	2.3	0.00	0.00	N/A	N/A	Granitic
MWH-033	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-032	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-017	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-014	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-013	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-010	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-009	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-008	\overline{O}	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-062	12	7.1	0.00	0.00	± 0	0.00	Granitic

Table 2-16. Summary Statistics for Pool Frequency at all Sites in 2015

Site Number	Sample Size (n)	Annual Mean	Standard Deviation	Standard Error	95% Confidence Interval	Coefficient of Variation	Geologic Type
MWH-006	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-016	0	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-044	0	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-047	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-050	26	124	31.77	6.23	±12.21	0.26	Granitic
MWH-061	25	113	36.71	7.34	±14.39	0.32	Non-granitic
MWH-060	25	160	3.10	0.62	±1.21	0.02	Granitic
MWH-059	25	142	20.77	4.15	±8.14	0.15	Granitic
MWH-057	24	109	27.65	5.64	±11.06	0.25	Granitic
MWH-056	25	123	30.97	6.19	±12.14	0.25	Granitic
MWH-055	21	91	27.22	5.94	±11.64	0.30	Granitic
MWH-053	22	108	31.96	6.81	±13.36	0.30	Granitic
MWH-033	27	105	64.24	12.36	±24.23	0.61	Non-granitic
MWH-032	21	147	26.15	5.71	±11.18	0.18	Non-granitic
MWH-017	Ω	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-014	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-013	0	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-010	Ω	N/A	N/A	N/A	N/A	N/A	Non-granitic
MWH-009	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-008	0	N/A	N/A	N/A	N/A	N/A	Non-granitic

Table 2-17. Summary Statistics for Bank Angle for all Sites in 2014

Note: Bank angles were not included in the PIBO protocol during the 2012 and 2013 surveys.

Note: Bank angles were not included in the PIBO protocol during the 2012 and 2013 surveys.

Site Number	CS	UU	US	${\sf C}{\sf U}$
MWH-006	$\mathbf 0$	$\mathbf 0$	$\mathbf 0$	$\overline{0}$
MWH-008	14	$\mathbf{1}$	$\overline{4}$	$\overline{2}$
MWH-009	13	5	$\mathbf{1}$	$\overline{2}$
MWH-010	6	\mathfrak{Z}	$\overline{7}$	6
MWH-013	16	1	$\overline{4}$	$\mathbf 0$
MWH-014	15	$\overline{2}$	\mbox{O}	3
MWH-016	20	$\mathbf 0$	$\mathbf{1}$	$\overline{2}$
MWH-017	21	$\mathbf{1}$	$\mathbf{1}$	$\mathbf 0$
MWH-032	$\mathbf 0$	$\mathbf 0$	$\mathsf{O}\xspace$	$\overline{0}$
MWH-033	$\mathbf 0$	$\mathbf 0$	\mbox{O}	$\mathsf{O}\xspace$
MWH-044	$\mathbf 0$	$\mathbf 0$	$\boldsymbol{0}$	$\mathsf{O}\xspace$
MWH-047	$\mathbf 0$	$\mathbf 0$	$\mathbf 0$	$\mathsf{O}\xspace$
MWH-050	$\mathbf 0$	$\mathbf 0$	\mbox{O}	$\mathsf{O}\xspace$
MWH-053	$\mathbf 0$	$\mathbf 0$	\mbox{O}	$\boldsymbol{0}$
MWH-055	$\mathbf 0$	$\mathbf 0$	\mbox{O}	$\boldsymbol{0}$
MWH-056	$\mathbf 0$	$\mathbf 0$	$\mathsf{O}\xspace$	$\mathbf 0$
MWH-057	$\mathbf 0$	$\mathbf 0$	$\mathsf{O}\xspace$	$\mathsf{O}\xspace$
MWH-059	$\boldsymbol{0}$	$\mathbf 0$	$\mathsf{O}\xspace$	$\mathsf{O}\xspace$
MWH-060	$\mathbf 0$	$\mathbf 0$	\mbox{O}	$\mathsf{O}\xspace$
MWH-061	$\mathbf 0$	$\mathbf 0$	$\mathsf{O}\xspace$	$\mathsf{O}\xspace$

Table 2-19. Summary Statistics for Bank Stability at all Sites in 2012

Covered = greater than 50 percent covered by vegetation, roots or rocks

Uncovered = less than 50 percent covered by vegetation, roots or rocks

Stable = minimal or no visible fractures, cracks or slumping Unstable = visible fractures, cracks or slumping

Key:

 $CS =$ Covered stable

UU = Uncovered unstable

US = Uncovered stable

Site Number	$\mathsf{CS}\xspace$	UU	US	CU
MWH-006	4	12	$\boldsymbol{6}$	$\mathsf{O}\xspace$
MWH-016	$\mathbf 0$	$\mathbf 0$	$\,0\,$	$\mathsf{O}\xspace$
MWH-044	21	$\mathbf 0$	$\,0\,$	$\mathsf{O}\xspace$
MWH-047	18	$\mathbf 0$	$\sqrt{3}$	$\mathsf{O}\xspace$
MWH-050	$\mathbf 0$	$\mathbf 0$	$\boldsymbol{0}$	$\boldsymbol{0}$
MWH-061	$\mathbf 0$	$\mathbf 0$	$\,0\,$	$\mathsf{O}\xspace$
MWH-060	$\mathbf 0$	$\mathbf 0$	$\,0\,$	\mbox{O}
MWH-059	$\mathbf 0$	$\mathbf 0$	$\boldsymbol{0}$	$\mathsf{O}\xspace$
MWH-057	$\mathbf 0$	$\mathsf{O}\xspace$	$\,0\,$	$\mathsf{O}\xspace$
MWH-056	$\mathbf 0$	$\mathbf 0$	$\mathbf 0$	$\mathsf{O}\xspace$
MWH-055	$\mathbf 0$	$\mathbf 0$	$\boldsymbol{0}$	$\mathsf{O}\xspace$
MWH-053	$\mathbf 0$	$\mathbf 0$	$\boldsymbol{0}$	$\mathsf{O}\xspace$
MWH-033	22	$\mathbf 0$	$\mathbf{1}$	$\mathbf 0$
MWH-032	24	$\mathbf 0$	$\mathbf 0$	\mbox{O}
MWH-017	21	$\mathbf 0$	$\mathbf{1}$	$\mathsf{O}\xspace$
MWH-014	19	$\mathbf 0$	\overline{a}	$\mathsf{O}\xspace$
MWH-013	17	$\mathbf 0$	$\overline{4}$	$\mathsf{O}\xspace$
MWH-010	12	$\mathbf 0$	9	\mbox{O}
MWH-009	17	$\mathbf{1}$	3	$\boldsymbol{0}$
MWH-008	20	$\mathbf 0$	1	$\mathbf 0$

Table 2-20. Summary Statistics for Bank Stability at all Sites in 2013

Covered = greater than 50 percent covered by vegetation, roots or rocks Uncovered = less than 50 percent covered by vegetation, roots or rocks

Stable = minimal or no visible fractures, cracks or slumping

Unstable = visible fractures, cracks or slumping

Key:

CS = Covered stable

UU = Uncovered unstable

US = Uncovered stable

Site Number	CS	UU	US	CU
MWH-006	$\boldsymbol{0}$	$\mathbf 0$	$\mathbf 0$	$\boldsymbol{0}$
MWH-016	$\mathbf 0$	$\mathbf 0$	$\mathsf{O}\xspace$	$\mathsf{O}\xspace$
MWH-044	$\boldsymbol{0}$	$\mathbf 0$	$\boldsymbol{0}$	$\boldsymbol{0}$
MWH-047	$\boldsymbol{0}$	$\mathbf 0$	$\mathsf{O}\xspace$	$\boldsymbol{0}$
MWH-050	19	$\overline{2}$	5	$\boldsymbol{0}$
MWH-061	25	$\mathbf 0$	$\mathbf 0$	$\boldsymbol{0}$
MWH-060	25	$\mathbf 0$	$\mathsf{O}\xspace$	$\mathbf 0$
MWH-059	23	$\overline{2}$	$\mathsf{O}\xspace$	$\mathbf 0$
MWH-057	21	$\mathbf 0$	\mathfrak{Z}	$\mathsf{O}\xspace$
MWH-056	17	$\overline{2}$	5	$\mathbf 0$
MWH-055	20	$\mathbf 0$	$\mathbf{1}$	$\mathbf 0$
MWH-053	18	3	$\mathbf{1}$	$\mathbf 0$
MWH-033	14	$\boldsymbol{6}$	$\boldsymbol{6}$	$\mathbf{1}$
MWH-032	20	$\mathbf 0$	$\mathbf{1}$	\mbox{O}
MWH-017	$\mathsf{O}\xspace$	$\mathbf 0$	$\mathsf{O}\xspace$	$\mathbf 0$
MWH-014	$\mathbf 0$	$\mathbf 0$	$\mathsf{O}\xspace$	$\mathbf 0$
MWH-013	$\mathbf 0$	$\mathbf 0$	$\mathsf{O}\xspace$	$\mathsf{O}\xspace$
MWH-010	$\boldsymbol{0}$	$\mathbf 0$	$\mathsf{O}\xspace$	$\mathbf 0$
MWH-009	$\boldsymbol{0}$	$\mathbf 0$	$\mathsf{O}\xspace$	$\mathbf 0$
MWH-008	$\mathbf 0$	$\mathbf 0$	$\mathbf 0$	$\mathbf 0$

Table 2-21. Summary Statistics for Bank Stability at all Sites in 2014

Covered = greater than 50 percent covered by vegetation, roots or rocks

Uncovered = less than 50 percent covered by vegetation, roots or rocks Stable = minimal or no visible fractures, cracks or slumping

Unstable = visible fractures, cracks or slumping

Key:

 $CS =$ Covered stable

UU = Uncovered unstable

US = Uncovered stable

Site Number	CS	UU	US	CU
MWH-006	8	$\overline{0}$	14	$\overline{0}$
MWH-016	21	$\mathsf{O}\xspace$	$\overline{2}$	$\mathsf{O}\xspace$
MWH-044	17	$\mathbf 0$	$\overline{4}$	$\mathsf{O}\xspace$
MWH-047	23	$\mathsf{O}\xspace$	$\mathbf 0$	$\overline{2}$
MWH-050	19	$\mathbf 0$	$\overline{7}$	$\mathbf 0$
MWH-061	$\mathbf 0$	0	$\mathbf 0$	$\mathsf{O}\xspace$
MWH-060	$\mathbf 0$	$\mathsf{O}\xspace$	$\boldsymbol{0}$	$\mathsf{O}\xspace$
MWH-059	$\mathbf 0$	$\mathbf 0$	$\mathbf 0$	$\mathbf 0$
MWH-057	16	0	8	$\mathbf 0$
MWH-056	19	1	5	$\mathbf 0$
MWH-055	18	$\mathsf{O}\xspace$	\mathfrak{Z}	$\mathsf{O}\xspace$
MWH-053	14	1	$\overline{7}$	$\mathbf 0$
MWH-033	$\boldsymbol{0}$	0	$\mathbf 0$	$\mathsf{O}\xspace$
MWH-032	$\mathbf 0$	0	$\mathbf 0$	$\mathsf{O}\xspace$
MWH-017	$\mathbf 0$	0	$\mathbf 0$	$\mathbf 0$
MWH-014	$\mathbf 0$	$\mathbf 0$	$\mathbf 0$	$\mathbf 0$
MWH-013	$\mathbf 0$	$\mathsf{O}\xspace$	$\mathbf 0$	$\mathsf{O}\xspace$
MWH-010	$\mathbf 0$	$\mathsf{O}\xspace$	$\mathbf 0$	$\mathbf 0$
MWH-009	$\mathbf 0$	$\mathbf 0$	$\mathbf 0$	$\mathbf 0$
MWH-008	$\mathbf 0$	$\mathbf 0$	$\mathbf 0$	$\mathbf 0$
MWH-062	21	1	5	$\mathbf 0$

Table 2-22. Summary Statistics for Bank Stability at all Sites in 2015

Covered = greater than 50 percent covered by vegetation, roots or rocks Uncovered = less than 50 percent covered by vegetation, roots or rocks

Stable = minimal or no visible fractures, cracks or slumping

Unstable = visible fractures, cracks or slumping

Key:

CS = Covered stable

UU = Uncovered unstable

US = Uncovered stable

Site Number	Sample Size (n)	Annual Mean	Standard Deviation	Standard Error	95% Confidence Interval	Coefficient of Variation	Geologic Type
MWH-006	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-016	23	4.5	0.96	0.20	±0.39	0.22	Granitic
MWH-044	$\mathbf 0$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-047	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-050	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-061	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Non- granitic
MWH-060	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-059	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-057	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-056	$\mathbf 0$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-055	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-053	$\mathbf 0$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-033	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Non- granitic
MWH-032	$\mathbf 0$	N/A	N/A	N/A	N/A	N/A	Non- granitic
MWH-017	23	7.6	1.91	0.40	±0.78	0.25	Non- granitic
MWH-014	20	6.2	0.83	0.19	±0.36	0.13	Granitic
MWH-013	21	5.7	1.45	0.32	±0.62	0.25	Granitic
MWH-010	22	7.4	2.88	0.61	±1.2	0.39	Non- granitic
MWH-009	21	10.7	1.69	0.37	±0.72	0.16	Granitic
MWH-008	21	6.6	1.46	0.32	±0.62	0.22	Non- granitic

Table 2-23. Summary Statistics for Bankfull Widths for all Sites in 2012

Site Number	Sample Size (n)	Annual Mean	Standard Deviation	Standard Error	95% Confidence Interval	Coefficient of Variation	Geologic Type
MWH-006	22	6.4	1.90	0.41	±0.8	0.30	Granitic
MWH-016	$\mathbf 0$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-044	21	4.0	1.27	0.28	±0.54	0.32	Granitic
MWH-047	21	3.9	1.44	0.32	±0.62	0.37	Granitic
MWH-050	$\boldsymbol{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-061	$\mathbf 0$	N/A	N/A	N/A	N/A	N/A	Non- granitic
MWH-060	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-059	$\mathbf 0$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-057	0	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-056	$\mathbf 0$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-055	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-053	$\mathbf 0$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-033	23	19.4	6.45	1.34	±2.63	0.33	Non- granitic
MWH-032	24	16.3	4.17	0.85	±1.67	0.26	Non- granitic
MWH-017	22	7.3	1.53	0.33	±0.64	0.21	Non- granitic
MWH-014	21	5.8	0.56	0.12	±0.24	0.10	Granitic
MWH-013	21	4.8	1.32	0.29	±0.56	0.27	Granitic
MWH-010	21	6.8	1.91	0.42	±0.81	0.28	Non- granitic
MWH-009	21	10.6	1.74	0.38	±0.74	0.16	Granitic
MWH-008	21	6.6	1.13	0.25	±0.48	0.17	Non- granitic

Table 2-24. Summary Statistics for Bankfull Widths for all Sites in 2013

Site Number	Sample Size (n)	Annual Mean	Standard Deviation	Standard Error	95% Confidence Interval	Coefficient of Variation	Geologic Type
MWH-006	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-016	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-044	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-047	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-050	26	9.6	2.66	0.52	±1.02	0.28	Granitic
MWH-061	25	8.1	1.42	0.28	±0.56	0.17	Non- granitic
MWH-060	25	7.0	0.62	0.12	±0.24	0.09	Granitic
MWH-059	25	7.6	1.30	0.26	±0.51	0.17	Granitic
MWH-057	24	7.3	1.72	0.35	±0.69	0.24	Granitic
MWH-056	25	7.1	1.91	0.38	±0.75	0.27	Granitic
MWH-055	21	5.9	1.77	0.39	±0.76	0.30	Granitic
MWH-053	22	7.0	2.08	0.44	±0.87	0.30	Granitic
MWH-033	27	18.2	5.76	1.11	±2.17	0.32	Non- granitic
MWH-032	21	16.6	2.61	0.57	±1.12	0.16	Non- granitic
MWH-017	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Non- granitic
MWH-014	$\mathbf 0$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-013	$\mathbf 0$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-010	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Non- granitic
MWH-009	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-008	$\mathbf 0$	N/A	N/A	N/A	N/A	N/A	Non- granitic

Table 2-25. Summary Statistics for Bankfull Widths for all Sites in 2014

Site Number	Sample Size (n)	Annual Mean	Standard Deviation	Standard Error	95% Confidence Interval	Coefficient of Variation	Geologic Type
MWH-006	22	6.4	1.82	0.39	±0.76	0.29	Granitic
MWH-016	23	5.9	2.13	0.44	±0.87	0.36	Granitic
MWH-044	21	3.9	0.66	0.14	±0.28	0.17	Granitic
MWH-047	25	4.4	1.76	0.35	±0.69	0.41	Granitic
MWH-050	26	10.0	2.85	0.56	± 1.09	0.28	Granitic
MWH-061	$\mathsf{O}\xspace$	N/A	N/A	N/A	N/A	N/A	Non- granitic
MWH-060	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-059	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-057	24	7.7	1.83	0.37	±0.73	0.24	Granitic
MWH-056	25	8.1	1.88	0.38	±0.74	0.23	Granitic
MWH-055	21	5.7	1.73	0.38	±0.74	0.30	Granitic
MWH-053	22	7.0	2.09	0.44	±0.87	0.30	Granitic
MWH-033	$\mathbf 0$	N/A	N/A	N/A	N/A	N/A	Non- granitic
MWH-032	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Non- granitic
MWH-017	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Non- granitic
MWH-014	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-013	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-010	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Non- granitic
MWH-009	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-008	Ω	N/A	N/A	N/A	N/A	N/A	Non- granitic
MWH-062	27	2.5	0.85	0.16	±0.32	0.34	Granitic

Table 2-26. Summary Statistics for Bankfull Widths for all Sites in 2015

Site Number	Sample Size (n)	Annual Mean	Standard Deviation	Standard Error	95% Confidence Interval	Coefficient of Variation	Geologic Type
MWH-006	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-016	23	0.6	0.14	0.03	±0.06	0.24	Granitic
MWH-044	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-047	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-050	$\mathbf 0$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-061	$\mathbf 0$	N/A	N/A	N/A	N/A	N/A	Non- granitic
MWH-060	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-059	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-057	$\mathbf 0$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-056	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-055	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-053	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
MWH-033	$\mathbf 0$	N/A	N/A	N/A	N/A	N/A	Non- granitic
MWH-032	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Non- aranitic
MWH-017	23	0.5	0.30	0.06	±0.12	0.56	Non- granitic
MWH-014	20	0.5	0.10	0.02	±0.04	0.20	Granitic
MWH-013	21	0.6	0.18	0.04	±0.08	0.31	Granitic
MWH-010	22	0.6	0.20	0.04	±0.08	0.35	Non- granitic
MWH-009	21	0.8	0.14	0.03	±0.06	0.18	Granitic
MWH-008	21	0.6	0.17	0.04	±0.07	0.29	Non- granitic

Table 2-27. Summary Statistics for Maximum Bankfull Depths for all Sites in 2012

Subwatershed	Sample Size (n)	Annual Mean Number per 100 meters	Standard Deviation	Standard Error	95% Confidence Interval	Coefficient of Variation	Geologic Type
Upper EFSFSR	3	6.1	5.37	3.10	±6.07	0.88	Granitic
Lower Burntlog	Ω	$- -$	$-$	$- -$	$- -$	$- -$	Granitic
Profile	$\mathbf 0$	$- -$	--	$- -$	$\qquad \qquad -$		Non-granitic
Goat	Ω	$- -$	--	$- -$	$\qquad \qquad -$	$- -$	Granitic
Fourmile	$\mathbf 0$		--		$- -$		Granitic
Riordan	$\mathbf 0$	$- -$	$-$	$- -$	$- -$		Granitic
Trapper	$\mathbf 0$	$- -$	$-$	$- -$	$- -$	$- -$	Granitic
No Man-Boulder	Ω	$- -$	$-$	$- -$	$- -$	$- -$	Non-granitic
Salt and Pepper	1	2.4	--	$- -$	$- -$	$- -$	Granitic
Tamarack	1	11.4			$- -$		Non-granitic
Sugar ϵ	$\overline{2}$	6.5	0.29	0.21	±0.41	0.05	Non-granitic

Table 2-35. Summary Statistics for Large Woody Debris Frequency in all Subwatersheds in 2012

Table 2-37. Summary Statistics for Large Woody Debris Frequency in all Subwatersheds in 2014

Key:

Subwatershed	Sample Size (n)	Annual Mean	Standard Deviation	Standard Error	95% Confidence Interval	Coefficient of Variation	Geologic Type
Upper EFSFSR	36	0.4	0.16	0.03	±0.05	0.42	Granitic
Lower Burntlog	9	0.6	0.19	0.06	±0.12	0.31	Granitic
Profile	0	N/A	N/A	N/A	N/A	N/A	Non-granitic
Goat Creek	13	0.3	0.08	0.02	±0.04	0.23	Granitic
Fourmile Creek	5	0.5	0.15	0.07	±0.13	0.29	Granitic
Riordan Creek	6	0.6	0.20	0.08	±0.16	0.31	Granitic
Trapper Creek	4	0.4	0.05	0.03	±0.05	0.12	Granitic
No Man-Boulder	0	N/A	N/A	N/A	N/A	N/A	Non-granitic
Salt and Pepper	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
Tamarack	Ω	N/A	N/A	N/A	N/A	N/A	Non-granitic
Sugar	Ω	N/A	N/A	N/A	N/A	N/A	Non-granitic

Table 2-42. Summary Statistics for Maximum Pool Depth in all Subwatersheds in 2015

Subwatershed	Sample Size (n)	Annual Mean	Standard Deviation	Standard Error	95% Confidence Interval	Coefficient of Variation	Geologic Type
Upper EFSFSR	36	5.3	1.49	0.25	±0.49	0.28	Granitic
Lower Burntlog	9	3.1	0.00	0.00	N/A	N/A	Granitic
Profile	0	N/A	N/A	N/A	N/A	N/A	Non-granitic
Goat Creek	13	4.9	0.01	0.00	±0.01	0.00	Granitic
Fourmile Creek	5	2.2	0.00	0.00	N/A	N/A	Granitic
Riordan Creek	6	3.6	0.00	0.00	± 0	0.00	Granitic
Trapper Creek	4	2.3	0.00	0.00	N/A	N/A	Granitic
No Man-Boulder	0	N/A	N/A	N/A	N/A	N/A	Non-granitic
Salt and Pepper	0	N/A	N/A	N/A	N/A	N/A	Granitic
Tamarack	0	N/A	N/A	N/A	N/A	N/A	Non-granitic
Sugar	0	N/A	N/A	N/A	N/A	N/A	Non-granitic

Table 2-46. Summary Statistics for Pool Frequency in all Subwatersheds in 2015

Note: Bank angles were not part of the PIBO protocol during the 2012 and 2013 surveys.

Note: Bank angles were not part of the PIBO protocol during the 2012 and 2013 surveys.

Note:

Covered = greater than 50 percent covered by vegetation, roots or rocks Uncovered = less than 50 percent covered by vegetation, roots or rocks

Stable = minimal or no visible fractures, cracks or slumping

Unstable = visible fractures, cracks or slumping

Key:

CS = Covered stable

UU = Uncovered unstable

US = Uncovered stable

CU = Covered unstable

Note:

Covered = greater than 50 percent covered by vegetation, roots or rocks Uncovered = less than 50 percent covered by vegetation, roots or rocks

Stable = minimal or no visible fractures, cracks or slumping

Unstable = visible fractures, cracks or slumping

Key:

 $CS =$ Covered stable

UU = Uncovered unstable

US = Uncovered stable CU = Covered unstable

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Note:

Covered = greater than 50 percent covered by vegetation, roots or rocks

Uncovered = less than 50 percent covered by vegetation, roots or rocks

Stable = minimal or no visible fractures, cracks or slumping Unstable = visible fractures, cracks or slumping

Key:

CS = Covered stable

UU = Uncovered unstable

US = Uncovered stable

CU = Covered unstable

Table 2-52. Frequency Distribution of Bank Stability in all Subwatersheds in 2015

Note:

Covered = greater than 50 percent covered by vegetation, roots or rocks $Uncovered =$ less than 50 percent covered by vegetation, roots or rocks

Stable = minimal or no visible fractures, cracks or slumping

Unstable = visible fractures, cracks or slumping

Key:

 $CS =$ Covered stable

UU = Uncovered unstable

US = Uncovered stable

CU = Covered unstable

Subwatershed	Sample Size (n)	Annual Mean	Standard Deviation	Standard Error	95% Confidence Interval	Coefficient οf Variation	Geologic Type
Upper EFSFSR	64	5.4	1.33	0.17	± 0.33	0.25	Granitic
Lower Burntlog	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
Profile	Ω	N/A	N/A	N/A	N/A	N/A	Non- granitic
Goat Creek	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
Fourmile Creek	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
Riordan Creek	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
Trapper Creek	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
No Man-Boulder	Ω	N/A	N/A	N/A	N/A	N/A	Non- granitic
Salt and Pepper	21	10.7	1.69	0.37	±0.72	0.16	Granitic
Tamarack	23	7.6	1.91	0.40	±0.78	0.25	Non- granitic
Sugar	43	7.0	2.31	0.35	±0.69	0.33	Non- granitic

Table 2-53. Summary Statistics for Bankfull Widths in all Subwatersheds in 2012

Table 2-54. Summary Statistics for Bankfull Widths in all Subwatersheds in 2013

Subwatershed	Sample Size (n)	Annual Mean	Standard Deviation	Standard Error	95% Confidence Interval	Coefficient οf Variation	Geologic Type
Upper EFSFSR	50	7.3	1.04	0.15	±0.29	0.14	Granitic
Lower Burntlog	26	9.6	2.66	0.52	±1.02	0.28	Granitic
Profile	25	8.1	1.42	0.28	±0.56	0.17	Non- granitic
Goat Creek	24	7.3	1.72	0.35	±0.69	0.24	Granitic
Fourmile Creek	25	7.1	1.91	0.38	±0.75	0.27	Granitic
Riordan Creek	21	5.9	1.77	0.39	±0.76	0.30	Granitic
Trapper Creek	22	7.0	2.08	0.44	±0.87	0.30	Granitic
No Man-Boulder	27	18.2	5.76	1.11	±2.17	0.32	Non- granitic
Salt and Pepper	21	16.6	2.61	0.57	±1.12	0.16	Granitic
Tamarack	Ω	N/A	N/A	N/A	N/A	N/A	Non- granitic
Sugar	0	N/A	N/A	N/A	N/A	N/A	Non- granitic

Table 2-55. Summary Statistics for Bankfull Widths in all Subwatersheds in 2014

Table 2-56. Summary Statistics for Bankfull Widths in all Subwatersheds in 2015

Subwatershed	Sample Size (n)	Annual Mean	Standard Deviation	Standard Error	95% Confidence Interval	Coefficient οf Variation	Geologic Type
Upper EFSFSR	64	0.6	0.15	0.02	±0.04	0.26	Granitic
Lower Burntlog	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
Profile	Ω	N/A	N/A	N/A	N/A	N/A	Non- granitic
Goat Creek	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
Fourmile Creek	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
Riordan Creek	Ω	N/A	N/A	N/A	N/A	N/A	Granitic
Trapper Creek	$\overline{0}$	N/A	N/A	N/A	N/A	N/A	Granitic
No Man-Boulder	Ω	N/A	N/A	N/A	N/A	N/A	Non- granitic
Salt and Pepper	21	0.8	0.14	0.03	± 0.06	0.18	Granitic
Tamarack	23	0.5	0.30	0.06	±0.12	0.56	Non- granitic
Sugar	43	0.6	0.18	0.03	± 0.05	0.32	Non- granitic

Table 2-57. Summary Statistics for Maximum Bankfull Depths in all Subwatersheds in 2012

Table 2-58. Summary Statistics for Maximum Bankfull Depths in all Subwatersheds in 2013

Subwatershed	Sample Size (n)	Annual Mean	Standard Deviation	Standard Error	95% Confidence Interval	Coefficient οf Variation	Geologic Type
Upper EFSFSR	50	0.7	0.15	0.02	±0.04	0.23	Granitic
Lower Burntlog	26	0.7	0.20	0.04	± 0.08	0.30	Granitic
Profile	25	0.9	0.29	0.06	±0.11	0.32	Non- granitic
Goat Creek	24	0.8	0.18	0.04	±0.07	0.22	Granitic
Fourmile Creek	25	0.5	0.16	0.03	± 0.06	0.30	Granitic
Riordan Creek	21	0.9	0.14	0.03	± 0.06	0.16	Granitic
Trapper Creek	22	0.6	0.20	0.04	± 0.08	0.30	Granitic
No Man-Boulder	27	0.8	0.27	0.05	±0.1	0.33	Non- granitic
Salt and Pepper	21	1.0	0.22	0.05	±0.09	0.22	Granitic
Tamarack	Ω	N/A	N/A	N/A	N/A	N/A	Non- granitic
Sugar	0	N/A	N/A	N/A	N/A	N/A	Non- granitic

Table 2-59. Summary Statistics for Maximum Bankfull Depths in all Subwatersheds in 2014

Table 2-60. Summary Statistics for Maximum Bankfull Depths in all Subwatersheds in 2015

PIBO ANOVA RESULTS TABLES SECTION 3:

Tables 3-1 through 3-3 present results from analysis of variance (ANOVA) tests for PIBO variables with sufficient data collected (bank angle, maximum pool depth, and WW:WD ratio). ANOVA tests compare the means of sampled groups to determine if any of those means are different from one another. Specifically, ANOVA tests the null hypothesis:

Null Hypothesis: The null hypothesis in ANOVA is that the means of the sites are equal for the parameter being tested. If the null hypothesis is not rejected, it means that the different sites are of the same sampling distribution.

Alternate Hypothesis: If the null hypothesis is rejected, the implication is the opposite – that the means of the sites are different.

A significance value (p) of 0.05 was used, indicating that we are 95 percent certain that the test has identified an actual difference when there was one. Assumptions of an ANOVA analysis include:

- Each group sample is from a normally distributed population
- Homogeneity of variances
- Independence of observations

To determine which sites when a difference in means was indicated, additional tests were then performed using Tukey's confidence intervals correction method at 95 percent confidence level $(p=0.05)$. Tukey's method was used because it is robust as it does not underestimate the least significant difference between the means of two sites.

In **Table 3-1** through **Table 3-3**, a difference between two sites – as determined by the ANOVA at the $p=0.05$ level – is indicated by a colored square at the cross-section of the two sites. A white square at the cross-section of two sites indicates no difference in means between the sites.

Note: Highlighted cells denote a statistically significant difference between sites Key:

ANOVA = Analysis of Variance

EFMC = East Fork Meadow Creek

EFSFSR = East Fork of the South Fork of the Salmon River

Table 3-2. ANOVA Results for Pool Depth for 2012-2015

		EFMC	EFSFSR (downstream to upstream)			Sugar Creek Meadow Creek (downstream (downstream to upstream to upstream)			Creek Bumtlog	Trapper Creek	Riordan Creek	Tamarack Creek	Fourmile Creek	Goat Creek	Fiddle Creek				
	Stream/MWH ID	006	009	013	032	033	044	008	010	014	016	047	050	053	055	017	056	057	062
	EFMC/006																		
	009																		
	013																		
EFSFSR	032																		
	033																		
	044																		
	008																		
Sugar Creek	010																		
	014																		
Meadow Creek	016																		
	047																		
	Burntlog Creek/050																		
	Trapper Creek/053																		
	Riordan Creek/055																		
	Tamarack Creek/017																		
	Fourmile Creek/056																		
	Goat Creek/057																		
	Fiddle Creek/062 Note: Highlighted cells denote a statistically significant difference between sites																		

Key: ANOVA = Analysis of Variance

EFMC = East Fork Meadow Creek

EFSFSR = East Fork of the South Fork of the Salmon River

Note: Highlighted cells denote a statistically significant difference between sites

Key: ANOVA = Analysis of Variance EFMC = East Fork Meadow Creek EFSFSR = East Fork of the South Fork of the Salmon River

Appendix 2: Macroinvertebrate Survey Report for 2012 to 2016

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Stibnite Gold Project Midas Gold Idaho, Inc.

April 2017

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SECTION 1:INTRODUCTION

EcoAnalysts, Inc. and MWH Americas, Inc. (MWH) conducted macroinvertebrate surveys in the summers of 2012 through 2014 and 2016 in the East Fork of the South Fork of the Salmon River (EFSFSR) and selected tributaries as part of an aquatic resources baseline study prior to the start of proposed mining operations at the Stibnite Gold Project in central Idaho. Macroinvertebrates are often used as biological indicators of water quality, being used to identify impaired waterways, assess aquatic life stressors, and indicate improvements. Macroinvertebrates provide a localized assessment of their response to stream conditions, making them a better indicator of water quality conditions than fish, which are typically more mobile. In addition to their value as a water quality indicator, macroinvertebrates are also the primary prey for most of the fish species in the aquatic resources study area.

The project, proposed by Midas Gold Idaho, Inc. (Midas Gold), is located in the Stibnite-Yellow Pine Mining District near Yellow Pine, Idaho and lies within the Salmon River Mountains near the Frank Church River of No Return Wilderness. The mining district is administered by the Krassel Ranger District of the Payette National Forest (PAF).

The terrain within the project area consists of narrow valleys surrounded by steep mountains. Elevations range from 6,000 to 6,600 feet above mean sea level along valley floors to more than 8,500 feet above mean sea level in the surrounding mountains. The EFSFSR is the main drainage basin in the project area.

Macroinvertebrates

Organisms without backbones that are visible without the aid of a microscope. Aquatic macroinvertebrates live on, under, and around rocks and sediment on the bottoms of lakes, rivers, and streams.

The purpose of these surveys was to evaluate stream health and water quality based on macroinvertebrates found at the sampling sites, and to allow future repeatable studies that can track any changes to habitat conditions that may result from mining activity. Species composition and abundances at each site were used to calculate quantitative measures of biological condition (metrics) that are based on habitat preferences, feeding strategies, waterquality tolerances, and expected occurrences of different macroinvertebrates. The magnitude and variation in these metrics provide useful information for understanding the water quality and habitat characteristics of streams in the project area.

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SECTION 2:METHODS

2.1 Field Methods

Benthic macroinvertebrate samples were collected from 11 sites (**Figure 2-1**) in the summers of 2012, 2013, 2014, and 2016 using protocols of the Pacific Anadromous Fish Strategy (PACFISH) Inland Fish Strategy (INFISH) Biological Opinion Effectiveness Monitoring Program (PIBO) (Heitke et al. 2011). Macroinvertebrates were collected at these sites with a 500 micron mesh Hess sampler $(2012 – 2014)$ or D-net (2016) , using a targeted composite method. Two individual samples were taken at randomized locations in each of four consecutive riffle/run habitat units at each sampling site, for a total of eight individual samples per site. These eight samples were then composited in the field into a single benthic macroinvertebrate sample per site, preserved in 95 percent ethanol, and transported to the EcoAnalysts, Inc. laboratory in Moscow, Idaho for sorting, subsampling, and taxonomic identification.

2.2 Laboratory Methods

Benthic Macroinvertebrates

Organisms that live in or on the bottom sediments of rivers, streams, and lakes. The benthic invertebrate community is strongly affected by its environment, including sediment composition and quality, water quality, and hydrological factors that influence the physical habitat. Because the benthic community is so dependent on its surroundings, it serves as a biological indicator that reflects the overall condition of the aquatic environment.

A subsample of 500 organisms was randomly sorted from each composite sample by rinsing the sample, mixing it to homogenize the composition, and spreading the mixed sample into a gridded tray. All organisms were picked from a random selection of grids until at least 500 organisms were removed from the composited sample. Once a grid was selected, all organisms were picked from that grid so that no partial grids were sorted. Therefore, subsamples typically exceeded 500 organisms, to a greater or lesser extent, depending on how many organisms were in the last grid sorted to achieve the minimum target of 500 organisms.

All organisms in the subsample were then identified and enumerated to the lowest practical level – generally by genus and species. If an individual subsample could not be identified to the lowest practical target resolution (e.g., due to poor condition, immaturity, or another indeterminate reason), it was identified to the lowest taxonomic resolution possible for that individual specimen. All individual organisms not identified to target resolution were noted as being unique or non-unique relative to the others in the sample.

Figure 2-1. Macroinvertebrate Survey Site Locations since 2012

2.3 Data Analysis

2.3.1 Idaho Stream Macroinvertebrate Index

The Idaho Stream Macroinvertebrate Index (SMI) was used to assess conditions for all sites and sampling years. Developed by the Idaho Department of Environmental Quality, the SMI is a methodology for using different macroinvertebrate population indicators to evaluate stream conditions and is used extensively to assess stream conditions throughout Idaho (Grafe 2002). The SMI is used to convert macroinvertebrate population measurements (metrics) into scores that are then combined into an overall final score and stream condition rating based on the best available regional reference conditions that were used to develop the SMI.

The macroinvertebrate population includes insects and non-insects, such as clams, snails, and worms. These populations are indicators of the condition of existing stream habitat as well as changes over time since macroinvertebrates have different water quality, habitat, and feeding requirements. Changes in habitat conditions can also be detected over relatively short time periods because of the short live cycles of insects and quick response to changing habitat. Therefore, describing the macroinvertebrate community also describes many aspects of stream conditions.

The Idaho SMI has been calibrated for different Idaho regions to account for varying organism composition and natural habitat conditions. For this study, the scoring formulas for the Idaho SMI Central and Southern Mountains region (where the Stibnite Gold Project is located) were used. Non-unique taxa were excluded from all richness calculations because inclusion of nonunique taxa would falsely inflate the number of unique macroinvertebrates found at the survey sites. See Grafe (2002) for detailed information on development and calculation of the SMI.

When reading through the study results, it is important to understand the distinction between metric *values* and metric *scores*. The SMI scoring system is designed to convert metric *values* into metric *scores* so that higher metric *scores* always respond positively to greater habitat quality or function regardless of whether a metric *value* (used to calculate the score) responds positively or negatively to perturbations. Following are descriptions of SMI metrics including the predicted response of metric *values* to perturbations in habitat and/or water quality such as sedimentation, organic pollution, or chemical pollution as described by Barbour (1999) and Grafe (2002). Bold text indicates how macroinvertebrate populations and associated metric *values* generally respond to such perturbations.

Total Taxa Richness: Total number of identifiably distinct taxa in a sample. The biodiversity of a stream declines as flow regimes are altered, habitat is lost, chemicals are introduced, energy cycles are disrupted, and alien taxa invade. Total taxa richness includes all aquatic invertebrates including insects, clams, snails, and worms. *Metric response to habitat/water quality perturbations = decrease*.

Ephemeroptera Taxa Richness: Total number of identifiably distinct taxa in the insect order Ephemeroptera (mayflies). The diversity of mayflies declines in response to most types of human influence. Many mayflies graze on algae and are particularly sensitive to chemical pollution. Mayflies may disappear when heavy metal concentrations are high while caddisflies and stoneflies are unaffected. Mayflies may also increase in numbers in response to nutrient enrichment. *Metric response to habitat/water quality perturbations = decrease, but can be variable*.

Plecoptera Taxa Richness: Total number of identifiably distinct taxa in the insect order Plecoptera (stoneflies). Stoneflies are often some of the first taxa to disappear from a stream due to disturbance. They require as clean substrate, and cool highly-oxygenated water. Most Plecoptera are predators, but several taxa are shredders that feed on leaf litter that drops from an overhanging tree canopy. *Metric response to habitat/water quality perturbations = decrease*.

Percent Plecoptera: Relative abundance of stoneflies, expressed as a percent of the total number of individuals in the sample. *Metric response to habitat/water quality perturbations = decrease*.

Trichoptera Taxa Richness: Total number of identifiably distinct taxa in the insect order Trichoptera (caddisflies). Caddisflies are an ecologically diverse group of insects and the variety of taxa declines as the variety and complexity of habitat declines. *Metric response to habitat/water quality perturbations = decrease*.

Hilsenhoff Biotic Index (HBI): An index of community tolerance to organic pollution and is a single value derived from the pollution tolerance values associated with each type of macroinvertebrate weighted by the abundance of those macroinvertebrates. Values range from 0 to 10, with higher values indicating more organic influence at a site. The index is influenced primarily by organic enrichment/compounds (i.e., organic pollution), but other factors, including temperature and sediment, can also have an influence. Different agencies may use different tolerance values for each species depending on geographical region. The default values used in this evaluation are the same as currently in use in Region 10 (Pacific Northwest) of the United States Environmental Protection Agency (EPA) (Barbour 1999). *Metric response to habitat/water quality perturbations = increase*.

Percent 5 Dominant Taxa: The percentage of individuals in a sample comprised of the five most abundant taxa. This is a measure of taxa evenness, and as diversity declines, the tendency is for fewer numbers of taxa to make up a greater proportion of overall abundance. *Metric response to habitat/water quality perturbations = increase*.

Scraper Taxa Richness: The total number of distinctly identifiable taxa in a sample whose primary feeding strategy is to scrape attached periphyton and other particulates. (Periphyton is a complex mixture of algae, cyanobacteria, heterotrophic microbes, and detritus that is attached to submerged surfaces in most aquatic ecosystems.) *Metric response to habitat/water quality perturbations = decrease.*

Clinger Taxa Richness: The total number of clinger taxa in a sample. Clingers having morphological and behavioral adaptations allowing them to hold on to smooth, stable substrates in flowing water. Clinger taxa are sensitive to sediment deposition, which can fill in crevice habitat between rocks in the streambed. *Metric response to habitat/water quality perturbations = decrease*.

2.3.2 Additional Selected Metrics

When developing a biological index such as the SMI, the final list of metrics are selected from a larger list of informative candidate metrics as those that best evaluate regional stream conditions. In fact, no two regionally-developed biological indices use identical metrics because of differences in regional conditions and also the scoring methods that researchers used to calculate the index. Basically, all candidate metrics provide useful information on stream condition whether or not they "make the cut" and are incorporated into the final index calculation. This section describes five additional non-SMI metrics selected as additional and informative

descriptors of macroinvertebrate populations, habitat and water quality conditions. The fact that these other metrics are not included in the SMI is not a reflection of the limitations of the SMI metrics, but rather a reflection of the strength of the final SMI metrics at evaluating and scoring stream conditions in Idaho. The additional metrics described below were chosen to supplement the SMI and provide a better overall picture of the streams. Bold text indicates how macroinvertebrate populations and associated metric *values* generally respond to such perturbations.

Long-lived Taxa Richness: The total number of taxa in a sample that require more than one year to complete their life cycle. These taxa must overwinter in the stream and are exposed to any natural and human disturbance that may occur during the winter. They are not usually found in ephemeral streams or streams prone to severe flooding. *Metric response to habitat/water quality perturbations = decrease.*

Metals Tolerance Index (MTI): McGuire's Metals Tolerance Index (MTI) was developed using benthic invertebrate communities and associated metal data (copper) in Montana's Clark Fork River (McGuire 1998). It is an index of community tolerance to metal contamination that ranks taxa according to their sensitivity to metals. It is calculated as a weighted average of tolerance values assigned to each taxon, similar to the HBI. As with the HBI, values range from 0 to 10, with higher values indicating that organisms are more tolerant of metals contamination. Values of 0 to 3 generally indicate little to no impact of metals on the macroinvertebrate community (i.e., relatively intolerant of metals contamination). *Metric response to habitat/water quality perturbations = increase.*

Intolerant Taxa Richness: The number of taxa in a sample with HBI values of 0 to 2 (i.e., highly intolerant of organic pollution). Intolerant taxa are the most sensitive and the first to disappear as disturbance increases. They typically represent 5 to 10 percent of all taxa in a region, with the remainder being either moderately or very tolerant. *Metric response to habitat/water quality perturbations = decrease.*

Percent Tolerant Individuals: The relative abundance of all individuals in a sample having HBI values of 8 to 10 (i.e., highly tolerant of organic pollution), expressed as a percentage of the total number of individuals in the sample. Although tolerant taxa may be found in many sites, they do not usually become dominant unless a disturbance removes intolerant taxa. The tolerant taxa can then dominate a community and fill in the niches left as intolerant taxa disappear. *Metric response to habitat/water quality perturbations = increase.*

Shannon-Weaver H' (log e): A community diversity index that describes the degree of taxonomic richness and how evenly the counts of those taxa are distributed in a sample. Diversity and evenness metrics are good indicators of the ability of a system to support varied taxa. The most common diversity index is the Shannon-Weaver index; also known as the Shannon-Weiner index (Shannon and Weaver 1949). This index can be a useful measure to compare invertebrate communities; however it is often redundant with other measures of diversity such as taxa richness and percent dominance. Higher values indicate greater species diversity and evenness. *Metric response to habitat/water quality perturbations = decrease.*

2.3.3 Functional Feeding Groups

The functional feeding group (FFG) composition of each sample was calculated as a percentage of individuals and number of taxa. The FFGs represented in the insect community are a direct function of available food and habitat resources. Some FFGs such as scraper taxa and predator

taxa are commonly used in multimetric models because they are good indicators of habitat quality and stability. The expected or probable response to perturbations are listed below for each FFG, according to Barbour (1999) and Grafe (2002). The response of FFG distribution to perturbations is more variable that other metrics, but FFG distribution can be an indicator of habitat diversity and/or diversity of food resources within and between sampling sites. Bold text indicates how macroinvertebrate populations and associated metric *values* generally respond to such perturbations.

Percent Filterers: The relative abundance of all individuals in a sample whose primary feeding mechanism is to filter suspended fine particulates, expressed as a percentage of the total number of individuals in the sample. *Metric response to habitat/water quality perturbations = variable.*

Percent Gatherers: The relative abundance of all individuals in a sample whose primary feeding mechanism is to gather deposited fine particulates, expressed as a percentage of the total number of individuals in the sample. *Metric response to habitat/water quality perturbations = variable.*

Percent Predators: The relative abundance of all individuals in a sample whose primary feeding mechanism is to pierce or engulf other invertebrates, expressed as a percentage of the total number of individuals in the sample. Predators represent the top of the food web and depend on stable sources of other invertebrates to eat. The percentage of predators is an indicator of community complexity supported by a site. In general, less disturbed sites support a greater diversity of prey items and habitats. *Metric response to habitat/water quality perturbations = variable, but often a decrease due to loss of habitat stability and diverse community structure.*

Percent Scrapers: The relative abundance of all individuals in a sample whose primary feeding strategy is to scrape attached periphyton and other particulates, expressed as a percentage of the total number of individuals in the sample. *Metric response to habitat/water quality perturbations = decrease.*

Percent Shredders: The relative abundance of all individuals in a sample whose primary feeding mechanism is to shred coarse particulate organic matter (CPOM), expressed as a percentage of the total number of individuals in the sample. *Metric response to habitat/water quality perturbations = variable.*

Percent Unclassified: The relative abundance of all individuals in a sample whose primary feeding mechanism is unknown or unclassified, expressed as a percentage of the total number of individuals in the sample. A taxa can be unclassified if the feeding group is unknown, undefined, or the taxa represents multiple feeding groups. This metric can usually be ignored due to usually low numbers of such taxa. *Metric response to habitat/water quality perturbations = variable.*

Filterer Richness: The total number of distinctly identifiable taxa in a sample in which the primary feeding mechanism is to filter suspended fine particulates. *Metric response to habitat/water quality perturbations = variable.*

Gatherer Richness: The total number of distinctly identifiable taxa in a sample in which the primary feeding mechanism is to gather deposited fine particulates. *Metric response to habitat/water quality perturbations = variable.*

Predator Richness: The total number of distinctly identifiable taxa in a sample in which the primary feeding mechanism is to pierce or engulf other invertebrates. Predators represent the top of the food web and depend on a stable sources of other invertebrates to eat. The percentage of predators is an indicator of community complexity supported by a site. In general, less disturbed sites support a greater diversity of prey items and habitats. *Metric response to habitat/water*

quality perturbations = variable, but often decrease due to loss of habitat stability and diverse community structure.

Scraper Richness: The total number of distinctly identifiable taxa in a sample in which the primary feeding strategy is to scrape attached periphyton and other particulates. (This is also an SMI metric – see additional description in SMI metric section.) *Metric response to habitat/water quality perturbations = decrease.*

Shredder Richness: The total number of distinctly identifiable taxa in a sample in which the primary feeding mechanism is to shred CPOM. This is an indicator of terrestrial vegetation input. *Metric response to habitat/water quality perturbations = variable.*

Unclassified richness: The total number of distinctly identifiable taxa in a sample in which the primary feeding mechanism is unknown or unclassified. A taxa can be unclassified if the feeding group is unknown, undefined, or the taxa represents multiple feeding groups. This metric can often be ignored due to typically low numbers of such taxa. *Metric response to habitat/water quality perturbations = variable.*

2.3.4 PIBO Observed/Expected Model

The PIBO observed/expected (O/E) index, derived from the statistical model RIVPACS (River Invertebrate Prediction and Classification System) was used to assess the biological condition of sites that were sampled in 2012, 2013, 2014, and 2016. The O/E index calculations were performed by Dr. Scott W. Miller, Director of the National Aquatic Monitoring Center at Utah State University, who developed RIVPACS model in 2009 in partnership with Robert Al-Chochachy of PIBO (PACFISH/INFISH Biological Opinion, U.S. Forest Service) and Chuck Hawkins of the Western Monitoring Center, Utah State University.

O/E models compare the macroinvertebrate taxa observed at sites of unknown biological condition (i.e., "test sites") to the populations expected to be found in the absence of humancaused stressors (Hawkins et al. 2000). The PIBO O/E model is based on 174 reference conditions throughout the western United States grouped into 10 distinct classes based on the similarity of macroinvertebrate populations among sites following the standard methods of Hawkins et al. (2000) and described in detail in the western streams and rivers statistical summary of the EPA's Environmental Monitoring and Assessment Program (Stoddard et al. 2006). The PIBO O/E model incorporates watershed area, long-term precipitation average, and maximum temperature to predict the expected macroinvertebrate populations for comparison to test sites.

Biological condition for this study was defined by how closely the measured macroinvertebrate community matched the expected community. Biological condition was assessed based on the precision of the western stream reference site data used to develop the PIBO O/E model (mean = 0.95, standard deviation $(SD) = 0.16$). In general, the amount a PIBO score differs from a score of 1.0 indicates the amount that the macroinvertebrate composition at a test site differs from what is expected under less disturbed conditions. Test sites scoring less than 1.0 SD below the mean of reference sites were rated in *good* biological condition (i.e., comparable to reference conditions); sites scoring between 1.0 SD and 2.0 SD below the mean in *fair* biological condition; and sites scoring more than 2.0 SD below the mean of reference sites in *poor* biological condition. All PIBO O/E scores greater than 1.0 are also rated as *good* because it indicates a higher than expected diversity of macroinvertebrate taxa in the test site relative to western reference sites, which is unlikely to occur under disturbed conditions.

PIBO O/E scores and rating categories are independent of SMI scores and rating categories. PIBO O/E scores and ratings evaluate only taxa richness in test sites based on 174 reference condition sites throughout the western United States. The SMI evaluates not only taxa richness, but incorporates other ecological information to come up with scores and rating categories based only on Idaho reference conditions.
SECTION 3:RESULTS AND DISCUSSION

3.1 Idaho Stream Macroinvertebrate Index (SMI) – Central and Southern Mountains

This section describes results of the analysis of stream conditions at all sites in all sampling years. The SMI scores shown in **Table 3-1** were calculated using scoring formulas for the Central and Southern Mountains region. A detailed description of the SMI methodology is found in Grafe (2002). The scoring methodology includes the following parameters (see details in Section 2.3.1):

- *Metric Values* are those values calculated most directly from the composition of macroinveterbrates in each sample. For example, Total Taxa simply equals the total number of unique taxa in a sample. Detailed descriptions and the responses of each metric value to habitat/water quality perturbations are described in **Section 2.3.1**.
- *Metric Scores* are calculated using on the Metric Valuesand always respond positively to higher quality habitat/water quality conditons.The equations for calculating each Metric Score were formulated based on Idaho reference conditions so that a score of 100 or greater indicates that a site is equivelent to the best reference sites used to develop the SMI in Idaho for that metric. Metric Scores cannot be less than 0, but there no maximum possible score.
- *SMI Rating* is the qualitative rating based the final SMI score. The levels for each rating category are indicated in **Figure 3-1**. The rating categories are as follows:
	- o Very Poor (0 to 19)
	- o Poor (20 to 39)
	- o Fair (40 to 58)
	- o Good (59 to 79)
	- o Very Good (80 to 100)
- *Metric Scores 100 Maximum* are simply the Metric Scores adjusted so that any value over 100 is reset down to 100 for the purposes of standardizing the scores to calculate a final average SMI score that ranges from 0-100. All metric scores less than 100 remain unadjusted. This adjustement is necessary for the calculation of a final overall score, but it is also useful to know which metrics "max-out" at 100 because those are the metrics that are equivelent to, or better than the Idaho reference condition sites used to develop the SMI. Any score of 100 or greater have been highlighted grey in **Table 3-1** to aid in interpretation of which metrics are driving the overall score for each stream. In **Figures 3-2** to **3-10**, a line at a value of 100 illustrates which metric scores "max-out" for each sampling site.
- *SMI Score* is the average of the Metric Scores that have been adjusted down to 100 where necessary. This average is the final overall SMI score for each site and the SMI rating is based on this final score.

In general, the final SMI scores were consistently high across sites and years because individual scores for most metrics were high, with several consistently exceeding the highest score of 100 (**Table 3-1**, **Figures 3-2** through **3-10**). The corresponding overall rating was *very good* for 7 of the 11 sample sites in all four sampling years (the highest category in the SMI scoring system) (**Table 3-1**, **Figure 3-1**). Among these seven consistently *very good* sites, there was relatively little variation in SMI scores between years with the exception of MWH-010 (upper Sugar Creek) which had a decreased score of 83.2 in 2016 as compared to 92.4 to 95.6 from 2012 to 2014. Even though the overall rating at MWH-010 was still considered *very good*, it's worthy to note that the departure from previous scores appeared to be driven by lower Trichoptera richness, percent Plecoptera, and lower water quality tolerance (i.e., HBI).

Only six occurrences of scores lower than the highest SMI condition category were observed, all of which were the next highest SMI category of *good*. Two of these occurrences were at MWH-007 in 2014 and 2016 (immediately downstream from the Yellow Pine Pit on the EFSFSR), and two were at MWH-015 in 2012 and 2016 (the constructed channel on Meadow Creek). However, in all four of these instances the SMI scores were similar across years and changes in SMI rating category represented only slight changes in the SMI scores.

The remaining two occurrences of *good* SMI ratings were in 2016 at MWH-016 (the furthest upstream Meadow Creek site) and MWH-017 (Tamarack Creek, a control site), both of which were departures from previous years. In particular, the furthest upstream Meadow Creek site (MWH-016) had a much lower overall SMI score than in previous years due largely to dramatically lower taxa richness (particularly Trichoptera and Plecoptera richness) and greater water tolerance values (corresponding to lower HBI scores) for those taxa collected. Reasons for a decrease in taxa richness at MWH-016 are unclear while the HBI scores were lower in 2016 across most sites as compared to previous years (see discussion below). The decreased SMI score at MWH-017 was driven by somewhat lower scores for nearly all metrics. MWH-017 is a designated control site and MWH-016 is upstream of the Stibnite project area where there is little human activity within and upstream of the site.

SMI metrics that generally scored high across survey sites and years included total taxa richness, Ephemeroptera taxa richness (mayflies), Trichoptera taxa richness (caddisflies), percent 5 dominant taxa (the five most abundant taxa), HBI (a measure of water quality tolerance), scraper taxa richness, and clinger taxa richness. However, as discussed, lower 2016 SMI scores for MWH-007, MWH-010, MWH-015, MWH-016, and MWH-017 were generally the result of lower taxa richness (particularly Plecoptera and Trichoptera richness), percent Plecoptera, HBI scores, scraper taxa, percent-5 dominant, and clinger taxa. There is some redundancy in these measures, particularly among the richness metrics, so changes in overall diversity could affect all of these scores.

Plecoptera taxa richness (stoneflies) exhibited higher variation among all sites and years, but still produced fairly high metric scores (**Figure 3-4**, **Table 3-1**). A notable exception was relatively low stonefly richness at site MWH-007 on the EFSFSR immediately downstream of the Yellow Pine Pit outlet. While suggestive of some aspect of water quality such as temperature, the low stonefly richness remains unclear considering that other metric scores appear healthy. Another possibility to explain the low stonefly richness at this site is sampling bias and potential inconsistency between sampling events, as MWH-007 is extremely difficult to sample because of high gradients, high current velocities, and very large substrate making the site difficult to effectively sample with nets. Plecoptera richness was lower at several sites in 2016 than in previous years.

The lowest metric scores and widest variation among sites was for percent Plecoptera individuals (**Table 3-1**, **Figure 3-6**). Sites MWH-010 (upper Sugar Creek) and MWH-017 (Tamarack Creek) had the highest average percentage of Plecoptera individuals (stoneflies) over the four sampling years (about twice the average scores of the other sites), but these average scores were largely driven by particularly high scores at these sites in 2012, and were markedly lower in 2016. Site MWH-007 (just below Yellow Pine Pit) had the lowest stonefly abundance, but was still generally similar to low abundances seen at all other sites. As with Plecoptera richness, this may suggest that the site has different water quality properties than other sites, but sampling bias cannot be ruled out considering the difficulty in sampling the site, as previously discussed. In summary, most sites exhibited moderate to high Plecoptera diversity, but relatively low numbers.

Given the high taxa richness of all the sites, it is difficult to draw firm conclusions regarding low Plecoptera abundance. Ultimately, relatively low abundance of Plecoptera individuals did not significantly affect the overall SMI scores because of consistently high taxa richness, an abundance of taxa having low water quality tolerance scores, and an abundance of clinger and scraper taxa which require what is considered high-quality stream habitat in the Central and Southern Mountains region, namely clean cobble substrate.

While there was some variation, the overall HBI scores were fairly high (generally above 75) for most sites and years, which detracted little from the overall SMI scores, particularly in light of additional metrics presented in the next section that demonstrate the low numbers of tolerant individuals and very high numbers of intolerant taxa found at all sites.

It is notable that HBI scores were generally lower across all sites in 2016 than in previous years, indicating that the HBI *values* were higher, with HBI values being the combination of tolerance of organic pollution weighted by the abundance of those taxa in each sample. These higher HBI values (and corresponding lower HBI scores) mean that the types and abundances of taxa collected at those sites were skewed towards being more tolerant of organic pollution than in previous years and may warrant further investigation to see if such trends hold into the future. Similarly, there appeared to be a lower number of clinger and scraper taxa at a number of sites in 2016. A possible explanation for this observed trend in HBI and some taxa richness scores across sites in 2016 is that sampling was conducted using a D-net in 2016 rather than a Hess sampler as in 2012-2014. However, it is unclear whether sampling method affected 2016 results in a systematic way. Equal effort (i.e., time and depths sampled) and similar randomization of replicate locations were employed in all years regardless of sampling gear used. In addition, a consistent trend in scores was not found for all metrics and both methods are commonly approved, and are known to similarly collect, and methods the representative macroinvertebrate community within riffle habitat.

Note: Shaded cells are scores that were set to the maximum allowable value of 100 because the calculated metric scores exceeded 100

*Values for scraper taxa generated in the standard metrics output is generally lower that values calculated for SMI scores because, when designating FFGs for the SMI metrics, taxa that are partially designated as scraper taxa are included in the FFG, but excluded from the standard metrics output.

Note: The SMI scores are calculated using formulas for the *Central and Southern Mountains* ecoregion.

Note: Higher scores indicate better overall stream condition as evaluated by the SMI

Figure 3-1: Idaho Stream Macroinvertebrate Index (SMI) Final Composite Scores

Note: Higher scores indicate greater numbers of all macroinvertebrate taxa

Figure 3-2: Idaho SMI Metric Scores: Total Taxa Richness

Note: Higher scores indicate greater numbers of Ephemeroptera taxa

Figure 3-4: Idaho SMI Metric Scores: Plecoptera Taxa Richness

Note: Higher scores indicate greater numbers of Trichoptera taxa

Note: Higher scores indicate that Plecoptera comprise a greater proportion of the macroinvertebrate abundance

Figure 3-6: Idaho SMI Metric Scores: Percent Plecoptera

Note: Higher scores indicate that macroinvertebrates are less tolerant of organic pollution

Figure 3-7: Idaho SMI Metric Scores: Hilsenhoff Biotic Index (HBI)

Note: Higher scores indicate greater numbers of taxa that exhibit scraper feeding behavior which is indicative of quality substrate habitat

Figure 3-8: Idaho SMI Metric Scores: Scraper Taxa Richness

Note: Higher scores indicate greater taxa diversity and that it is less likely that a few taxa comprise most of the macroinvertebrate population

Note: Higher scores indicate greater numbers of taxa that exhibit clinger behavior which is indicative of quality substrate habitat

Figure 3-10: Idaho SMI Metric Scores: Clinger Taxa Richness

3.2 Additional Selected Metrics

This section describes results of the analysis of the five additional metrics for all sites and sampling years. As described in **Section 2.3.2**, these were added to the SMI metrics because they provide valuable additional data for describing overall habitat quality.

The values and scores discussed below and shown in **Table 3-2** were calculated and include the following parameters:

- Feeding Groups are listed by percentage of the relative abundance of all individuals in a sample based on feeding mechanism.
- Richness Scores (including unclassified species) are a measurement of the total number of identifiably distinct organisms in a sample.
- Long-lived Taxa Richness lists the total number of what are considered long-lived taxa in a sample.
- Metals Tolerance Index is a weighted average of tolerance to metal contamination assigned to each taxon, with values range from 0 to 10; higher values indicate greater tolerance to metals.
- Intolerant Taxa Richness lists the number of taxa in a sample with low tolerance values of 0 to 2; indicating that those taxa are considered highly intolerant of organic pollution at a site.
- Percent Tolerant Individuals shows the percentage of all individuals in a sample having tolerance values of 8 to 10 and considered highly tolerant of poor water quality.
- Shannon-Weaver H' (log e) is an index indicating the ability of a system to support species diversity, with higher values indicating greater species diversity and evenness.

Long-lived Taxa Richness (**Table 3-2, Figure 3-11***):* All sites from 2012 through 2014 sample years had long-lived taxa abundances of 5 to 10, indicating habitat availability is relatively stable. This also held true in 2016 except for MWH-016, which is the farthest upstream site on Meadow Creek. However, taxa richness among many groups was lower at this site than in previous years, so slightly fewer long-lived taxa in 2016 may not be the result of factors that can disproportionately affect long-lived taxa such as habitat instability or overwintering conditions.

Metals Tolerance Index (MTI) (**Table 3-2, Figure 3-12**): MTI values of 0 to 3 indicate "no impact" according to McGuire (1993). Across sites from 2012 to 2014, the MTI values were low, ranging from 1.46 to 3.09, and show relatively little variation across years, with one exception. In 2012, the MTI value was 4.7 at site MWH-015 (constructed channel on Meadow Creek), and decreased to 2.42 in 2013, and 1.98 by 2014. This indicates a shift toward a greater number of taxa and/or proportion of individuals that are relatively intolerant of metals contamination. In this case, the most notable change in metals-intolerant taxa was an increase in Tanytarsini midges from 2012 (0.19 percent) to 2014 (2.12 percent), suggesting that metals concentrations within the constructed channel may be decreasing over time. In 2016, two sites exhibited MTI values greater than 3.0. These sites were MWH-007 (immediately downstream of the Yellow Pine Pit; $MTI = 3.49$) and at MWH-016 (the furthest upstream Meadow Creek site; MTI = 3.81). In particular, the increased MTI score (i.e., lower tolerance to metals) at the upstream Meadow Creek site is a marked departure from earlier year's results. Another site showing a notable departure from previous years is at site MWH-009 (EFSFSR), the next site downstream from

MWH-007. The MTI values at this site from 2012 to 2014 were consistently around 2.0, whereas in 2016 the value was 2.97, again indicating possible increased metals contamination at this site.

Intolerant taxa richness (**Table 3-2**, **Figure 3-13**): Intolerant taxa richness was extremely high for all sites and years – about 75 percent of the taxa in individual samples. Typically, these taxa represent 5 to 10 percent of all taxa in a region, with the remainder being either moderately or very tolerant. As with other 2016 richness metrics, we observed a notable decrease in intolerant taxa richness at site MWH-016 that was a sharp departure from previous years. In fact, this site had the highest and most consistent number of taxa intolerant of organic pollution from 2012 to 2014, but had the lowest number of intolerant taxa in 2016. While intolerant taxa still made up a large proportion of the macroinvertebrate community, this decrease along with the other observed decreases in taxa richness warrants further monitoring to determine if this trend continues.

Percent Tolerant Individuals (**Table 3-2**, **Figure 3-14**): As expected with such abundance of intolerant taxa in the samples, the percent of tolerant individuals was quite low – comprising no more than four percent of all individuals in any sites from 2012 to 2014. This was consistent with other metrics in those years that indicate high water quality among all sites in the project area. In 2016, there was a relatively large increase in percent tolerant individuals at the furthest downstream Meadow Creek site (MWH-014) and at the site immediately downstream from there, Site MWH-12 (just below the Meadow Creek/EFSF confluence). Although the intolerant macroinvertebrates comprised only about four to six percent of the samples at those sites in 2016, in previous years intolerant organisms comprised no more than two percent of the samples in any given year. This may represent an increase in pollution, particularly when other changes in taxa richness and metals tolerance at Meadow Creek sites are taken into account.

Shannon-Weaver Diversity Index (**Table 3-2**, **Figure 3-15**): From 2012 to 2014, Shannon-Weaver index values ranged from 2.68 to 3.56, with few sites having values less than 3.0. This indicates high diversity and evenness among all sites, which is well-supported by other measures of richness and evenness for these sites. The lowest values of any site during those years were 2.68 at MWH-007 (just below the Yellow Pine Pit) in 2014 and 2.72 at MWH-014 (farthest downstream site in Meadow Creek) in 2013. The relatively low index value at MWH-007 in 2014 was likely due to a large number of blackfly larvae in the sample (see description of functional feeding group below). The index value for the downstream Meadow Creek site in 2013 was lower than 2012 and 2014, so no trend was observed. No site had consistently lower diversity index values than other sites, and no directional trends in diversity were found at any sites from 2012 to 2014. In 2016, the index values were comparable to previous years at all sites except for notably higher values at MWH-011 (SW Index = 3.58) and MWH-012 (SW Index = 3.73), which are the two sites immediately below the Meadow Creek/EFSFSR confluence. In general, the consistently good Shannon-Weaver index values can be explained by high taxa richness and the abundances of taxa being spread evenly among those taxa found at the sites rather than only a few taxa dominating the samples.

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*Values for scraper taxa generated in the standard metrics output is generally lower that values calculated for SMI scores because, when designating FFGs for the SMI metrics, taxa that are partially designated as scraper taxa are included in the FFG but excluded from the standard metrics output.

Note: Higher values indicate greater numbers of long-lived taxa which require stable and diverse habitat

Figure 3-11: Additional Selected Metrics: Long-lived Taxa Richness

Note: Higher values indicate that the macroinvertebrate community is more tolerant of metals

Figure 3-12: Additional Selected Metrics: Metals Tolerance Index (MTI) Values

Note: (higher values indicate greater numbers of taxa that are intolerant of poor water quality)

Note: Low values indicate low numbers of individuals that are considered tolerant of poor water quality

Figure 3-4: Additional Selected Metrics: Percentage of Tolerant Individuals

Note: Higher scores indicate greater macroinvertebrate species diversity

3.3 Functional Feeding Groups

The FFG composition of samples across all sites and years was remarkably consistent and stable, particularly for richness values across all years at sites in all survey years, which is an indication of no notable shifts in in habitat conditions (**Table 3-2**, **Figure 3-16** and **Figure 3-17**). Richness across sites and years was dominated by gatherers and predators, to a lesser extent shredders and scrapers, and also a low, but stable percentage of filterer taxa.

Once again, the data indicate a high level of stability and consistency of available habitat and food sources at the study sites, particularly with the apparently rich predator population, which is dependent on diverse and stable food resources in the form of other macroinvertebrates.

Functional feeding group composition measured as a percentage of individuals was not as consistent as taxa richness, but largely mirrored the observed disposition of taxa FFGs. Less consistency in FFG abundances is expected since it is easier to consistently sample taxa richness at a site than to consistently sample both abundance and taxa richness due to spatial variation in macroinvertebrate numbers within a stream. In terms of abundance, numbers of individuals were dominated by scrapers and gatherers, with a few notable instances of a high proportion of shredders (MWH-013 in all years, and MWH-014 in 2013).

The most obvious difference in FFG abundance was the larger proportion of filterers at MWH-007 (downstream from the Yellow Pine Pit outlet) that was observed at other sites, which was almost entirely due to a large number of blackflies in the sample. This site is classic habitat for blackflies, which thrive in areas of high-gradient, cold-water, abundant clean substrate, and where a lake outlet provides abundant particulates that the blackfly larvae feed on (Kim and Merritt 1987).

The increase in filterer abundance from 2012 to later sampling years at MWH-007 may indicate an increase in organic and particulate discharge from the Yellow Pine Pit, which likely warrants further investigation or monitoring because of the potential implications for stream conditions further downstream. However, the change in abundances may also be due to sampling bias. Because blackfly larvae can be found in extremely high densities in such habitat, it is possible for even a single sampling repetition to inadvertently contribute a disproportionate number of individuals to a combined sample. Another source of potential bias would be that MWH-007 was the most difficult of all the sites to sample due to the large substrate size and fast current as previously discussed. Therefore, high densities of blackflies may have been present in 2012, but were not well-represented in the sample.

Note: each bar represents a single sample for each sampling site per year and the colors correspond to the either the percent of all individuals or number of taxa within each feeding group in that sample

Figure 3-6: Functional Feeding Group Composition for Macroinvertebrates Collected Between 2012 and 2016 Expressed as the Percentage of Individuals Per Sample

Note: each bar represents a single sample for each sampling site per year and the colors correspond to the either the percent of all individuals or number of taxa within each feeding group in that sample

Figure 3-77: Functional Feeding Group Composition for Macroinvertebrates Collected Between 2012 and 2016 Expressed as the Number of Taxa Per Sample

3.4 PIBO Observed/Expected Index (RIVPACS)

This section describes results of the analysis of the sampled sites in 2012, 2013, 2014, and 2016 using the PIBO O/E index from the RIVPACS statistical model (see Section 2.3.4).

The O/E scores discussed below and shown in **Table 3-3** are based on reference site data from throughout the West, with a mean of 0.95 and a standard deviation of 0.16. In general, departures from an O/E ratio of 1.0 indicate that the taxonomic composition in a stream sample differs from that expected under less disturbed conditions. The biological condition rating for each sampling site is a qualitative rating compared to reference sites ranging from:

- Poor (more than 2 standard deviations below the mean of reference sites)
- Fair (between 1 and 2 standard deviations below the mean of reference sites)
- Good (either less than one standard deviation below the mean of reference sites *or* a value greater than 1.0)

The O/E scores among sampling sites from 2012 through 2016 consistently rated as being in *good* biological condition except in three instances where a site was rated as *fair* (**Table 3-3**). A *fair* condition rating was found in 2012 at MWH-007 (Meadow Creek confluence) and in 2013 at sites MWH-029 (downstream Sugar Creek site) and MWH-015 (Meadow Creek constructed channel). In all three instances, sites with *fair* O/E scores improved to *good* in subsequent sampling years. The O/E condition rating results indicate that the macroinvertebrate assemblages are consistently very comparable to reference conditions in the western states ecoregion for which the O/E model was developed (Stoddard et al. 2006).

Table 3-3. PIBO Observed/Expected Index (RIVPACS) Scores for MWH Survey Sites, 2012-2016

3.5 Results by Sampling Site

All sites were similar in many respects. They all had high overall SMI scores, indicating that the streams were in *very good* or *good* condition in all sampling years relative to the regional reference sites used to develop the Central and Southern Mountains SMI. These high scores were consistently the result of high taxa richness, diversity/evenness measures, high HBI scores (indicating an abundance of taxa requiring high water quality), and large numbers of scraper and clinger taxa that require high quality substrate. Abundance of Plecoptera was generally lower than expected across all sites in all years. As previous described, in 2016 many sites had notably lower taxa richness for some groups as well as lower HBI scores indicating a greater tolerance of organic pollution. Several other metrics consistently supported the SMI metric calculations namely high Shannon-Weaver diversity values; many long-lived taxa that require stable habitats; high numbers of intolerant taxa; low numbers of tolerant individuals; and an even distribution of several FFGs, most notably a large proportion of predator taxa. Metals tolerance also fluctuated somewhat among sites and years. Specific results for each site are presented in the following site summaries.

3.5.1 MWH-007

This site, located on the EFSFSR immediately downstream of the Yellow Pine Pit outlet, is characterized by very large substrate sizes and high gradient relative to the other macroinvertebrate sampling sites. It is unique among the sites because it is most directly influenced by an impoundment, which likely affects water quality in terms of temperature, chemical properties and productivity. MWH-007 had SMI ratings of *very good* in 2012 and 2013, and lower ratings of *good* in 2014 and 2016. The overall SMI scores were similar across years and the change in condition category represented only a slight decrease in some metrics that comprise the overall SMI score. Notable, were lower scraper taxa richness and HBI scores in 2016, which were observed at several sites. As with most sites in almost all years, the percent of Plecoptera in samples was lower than expected.

Most richness and evenness metrics tended to be more variable and have lower values than other sites, which is expected given the likely influence of the upstream impoundment. Filterer abundance is notably greatest at this site because it is ideal habitat for blackflies, which can occur in very high densities. An increasing trend in blackfly abundance at this site may indicate changes in water quality; however sampling bias is possible since the site is difficult to sample and lake outlets commonly have large blackfly populations. A notable change from previous years was an increase in the MTI in 2016.

PIBO O/E scores indicated that this site was in *fair* condition in 2012 and was rated as *good* in 2013, 2014, and 2016.

3.5.2 MWH-009

This site is located on the EFSFSR approximately one mile downstream of the Yellow Pine Pit. MWH-009 had a SMI rating of *very good* in all sampling years, but lower scores for scraper taxa, HBI, and percent 5 dominant taxa were observed in 2016. Metrics at this site are similar to most other sites, with high taxa richness, taxa indicative of good water quality, and stable diverse habitat availability. As with most sites in almost all years, the percent of Plecoptera in samples

was lower than expected. In addition, as at MWH-007 immediately upstream, the MTI was somewhat elevated in 2016 relative to previous years.

PIBO O/E scores indicated that this site was rated as *good* in all sampling years.

3.5.3 MWH-010

This is the upstream site on Sugar Creek and was rated as *very good* in all sampling years, although the overall SMI score was lower in 2016 (SMI score = 83.2) as compared scores ranging from 92.4 to 95.6 from 2012 to 2014. This departure from previous years was driven by lower scores for several metrics including Plecoptera richness, Trichoptera richness, HBI, clinger taxa, and a large decrease in percent Plecoptera,. Lower numbers of intolerant taxa and longlived taxa were also found in 2016. Prior to 2016, the only notable difference in the macroinvertebrate community relative to other sites was a higher average number of Plecoptera. In this respect, the upstream Sugar Creek site was most similar to the Tamarack Creek control site (MWH-017). In general, it is difficult to draw any firm conclusions regarding potential changes in habitat over time considering the high scores for most metrics and the overall SMI scores across all years. However, the observed decrease in some metric scores seen in 2016, particularly the decrease in Plecoptera numbers, warrants future observation to see if these trends continue.

PIBO O/E scores indicated that this site was rated as *good* in all sampling years.

3.5.4 MWH-029

This site is the downstream site on Sugar Creek immediately upstream of the confluence with the EFSFSR and was rated as *very good* in all sampling years. The metric scores at this site were similar to most other sites and characterized by high taxa richness, taxa indicative of good water quality, and stable diverse habitat availability. In 2016, somewhat lower HBI scores and elevated MTI values relative to previous years were observed. As with most sites in almost all years, the percent of Plecoptera in samples was lower than expected.

PIBO O/E scores in 2013 indicated that this site was in *fair* condition and was rated as *good* in 2012, 2014 and 2016.

3.5.5 MWH-011

This site is located on the EFSFSR just downstream of, and in close proximity to, the mining camp at the project station. MWH-011 was rated as *very good* with extremely similar overall SMI scores across sampling years. Metrics at this site were similar to most other sites, with high taxa richness, taxa indicative of good water quality, and stable diverse habitat availability. As with most sites in almost all years, the percent of Plecoptera in samples was lower than expected, and in 2016 we observed somewhat lower HBI scores than in previous years. Consistent with a slightly lower HBI score in 2016 was a slight increase in numbers of macroinvertebrates that are highly tolerant of organic pollution.

PIBO O/E scores indicated that this site was rated as *good* in all sampling years.

3.5.6 MWH-012

This site is located on the EFSFSR immediately downstream of the confluence with Meadow Creek and upstream of the bridge crossing of the EFSFR near the project station. MWH-012 was rated as *very good* in all sampling years based on generally high taxa richness, taxa indicative of good water quality, and stable diverse habitat availability. As with most sites in almost all years, the percent of Plecoptera in samples was lower than expected, and in 2016 we observed somewhat lower HBI scores than in previous years. Also observed in 2016 was a large relative increase in the percentage of macroinvertebrates that are highly tolerant of organic pollution.

PIBO O/E scores indicated that this site was rated as *good* in all sampling years.

3.5.7 MWH-013

This is the farthest upstream site on the EFSFSR and was rated as *very good* across all sampling years. Metrics were similar to most other sites, with high taxa richness, taxa indicative of good water quality, and stable diverse habitat availability. As with most sites in almost all years, the percent of Plecoptera in samples was lower than expected, but HBI scores were not lower in 2016 than in previous years as observed at many other sites.

PIBO O/E scores indicated that this site was rated as *good* in all sampling years.

3.5.8 MWH-014

This is the farthest downstream site located on Meadow Creek and was rated as *very good* across all sampling years with little change among all SMI metrics. Metrics were similar to most other sites, with high taxa richness, taxa indicative of good water quality, stable diverse habitat availability, and lower than expected Plecoptera abundance. However, in 2016 there was a large relative increase in the percentage of macroinvertebrate individuals that are highly tolerant of organic. This result is similar to what was also observed MWH-012 immediately downstream from this site.

PIBO O/E scores indicated that this site was rated as *good* in all sampling years.

3.5.9 MWH-015

This site is located within the constructed channel on Meadow Creek. It was rated as *very good* sites in 2013 and 2104 and rated as *good* in 2012 and 2016. However, in all four of these instances the SMI scores were similar across years and changes in SMI rating category represented only slight changes in the actual SMI scores. Most richness and evenness metrics tended to be more variable and have lower values at this site than other sites, but the metrics are still similar to most other sites including lower than expected Plecoptera abundance. A sharp change in MTI value was found at this site, decreasing from high 2012 levels to values similar to other sampling sites in 2013, 2014, and 2016 indicating that metals contamination may have decreased after 2012. There was also a notable decrease in observed Trichoptera richness in 2016 as compared to previous years. HBI scores were not lower in 2016 than in previous years as found at many other sites.

PIBO O/E scores in 2013 indicated that this site was in *fair* condition, but was rated as *good* in 2012, 2014 and 2016.

3.5.10 MWH-016

This is the furthest upstream site located on Meadow Creek. MWH-016 was rated as *very good* from 2012 through 2014, but had a substantially lower SMI score and rating in 2016. From 2012 to 2014, metrics were similar to most other sites, with high taxa richness, taxa indicative of good water quality, stable diverse habitat availability, and lower than expected Plecoptera abundance. A decreased overall SMI score/condition rating was observed in 2016 that was determined by decreased taxa richness of several taxa groups, decreased taxa evenness, fewer clinger taxa, and a lower HBI score (indicating that the taxa found had greater tolerance of organic pollution). Similar to the change in HBI values, there was a large decrease in the number of taxa that are highly intolerant of organic pollution. In addition, the MTI value in 2016 was notably higher than in previous years indicating the macroinvertebrate community sampled in 2016 was more tolerant of metals contamination. Reasons for such changes in metric scores are unclear since MWH-016 is upstream of the Stibnite project area where little human activity would be expected.

PIBO O/E scores indicated that this site was rated as *good* in all sampling years.

3.5.11 MWH-017

This control site is located on Tamarack Creek. MWH-017 rated as *very good* from 2012 through 2014 and showed a slight decrease in overall SMI score in 2016 which was *good*. The decreased 2016 SMI score was driven by somewhat lower scores for virtually all metrics. In general though, metric scores were were similar to most other sites, with high taxa richness, taxa indicative of good water quality, stable diverse habitat availability. One distinction between MWH-017 and most other sites was a somewhat higher abundance of Plecoptera, which is most similar to the upstream site on Sugar Creek in this respect.

PIBO O/E scores indicated that this site was rated as *good* in all sampling years.

SECTION 4:CONCLUSIONS

The macroinvertebrate assemblages at all MWH macroinvertebrate sites are generally indicative of high water quality and relatively stable habitat availability. This is indicated by high taxa richness, the presence of many predator and long-lived taxa, taxa that require clean cobble substrate, low MTI values, high HBI scores, high proportions of taxa that are intolerant of poor water quality, and low numbers of individuals that are tolerant of poor water quality. Lower than expected numbers of Plecoptera were found among all sites in all years. In 2016, HBI scores were consistently lower across most sites indicating a general lower tolerance of organic pollution or some undetermined aspect of sampling bias. Richness of some taxonomic groups was also lower in 2016, but this did not occur in a consistent pattern across sites as would be expected if sampling bias was a factor.

High proportions of clinger and scraper taxa were also indicative of abundant clean substrate, although some sites in 2016 had lower scraper and clinger taxa richness relative to previous years.

Notable differences in macroinvertebrate composition included a larger proportion of filterers (in this case blackfly larvae) just downstream of the Yellow Pine Pit outlet at MWH-007, somewhat greater numbers of stoneflies at the upstream Sugar Creek and Tamarack Creek sites, and a sharp decrease in MTI values within the Meadow Creek constructed channel, indicating that metals contamination may have decreased since 2012 at MWH-015. In 2016, a decrease in the overall SMI score was observed at the furthest upstream Meadow Creek site (MWH-016). Additionally an increased MTI value and a decrease in the number of taxa intolerant of organic pollution relative to previous years was observed at MWH-016. Other sites exhibiting potential decreases in water quality include an increase in MTI values at sites most immediately downstream of the Yellow Pine Pit (MWH-007 and MWH-009) and increases in the number of intolerant individuals near the confluence of Meadow Creek and EFSFSR.

PIBO O/E scores rated all sites across all sampling years as *good* with respect to regionally expected macroinvertebrate populations with only three exceptions across sampling years. These exceptions included *fair* ratings just below the Yellow Pine Pit (MWH-007) in 2012, lower Sugar Creek in 2013 (MWH-029), and the Meadow Creek constructed channel (MWH-015) in 2013.

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SECTION 5:REFERENCES

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5.2 Abbreviations and Acronyms

SECTION 6:LIST OF PREPARERS

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Appendix 3: Water Temperature Watershed Condition Indicators and Monitoring Results

Appendix 3: Water Temperature Watershed Condition Indicators and Monitoring Results

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WATER TEMPERATURE MONITORING RESULTS AND WATERSHED CONDITION INDICATORS

The following figures are plots of 7-day average maximum water temperature overlain on the water temperature Watershed Condition Indicator (WCI) ranges for spawning and rearing Chinook salmon and steelhead, and bull trout spawning, incubation and rearing. **Table 1** summarizes the water temperature WCIs from Appendix B of the Land and Resource Management Plan (PAF 2003).

Table 1. Water Temperature Watershed Condition Indicators for Chinook Salmon, Steelhead and Bull Trout Based on the 7-Day Average Maximum Temperature

Source: PAF 2003, page B-13 in Appendix B

Sites where the thermographs were not installed until 2014, were lost, or were inaccessible during late fall data collections activities, show as data gaps in the figures below. USGS gage data for 5 stations are also included. The WCIs are shown in all figures by a color bar:

- Green represents 'Functioning Appropriately'
- Yellow represents 'Functioning at Risk'
- Red represents 'Functioning at Unacceptable Risk'

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Figure 2. Water Temperature Monitoring Results and Watershed Condition Indicators for
Steelhead in the EFSFSR, MWH-001

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Figure 4. Water Temperature Monitoring Results and Watershed Condition Indicators for Chinook
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Spawning

Figure 8. Water Temperature Monitoring Results and Watershed Condition Indicators for
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Figure 9. Water Temperature Monitoring Results and Watershed Condition Indicators for Bull Trout

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Figure 19. Water Temperature Monitoring Results and Watershed Condition Indicators for Chinook
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Figure 20. Water Temperature Monitoring Results and Watershed Condition Indicators for
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River **at the U.S. Geological Survey Gage, 13311250 in the East Fork of the South Fork of the Salmon**

Figure 43. Water Temperature Monitoring Results and Watershed Condition Indicators for Chinook
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Figure 48. Water Temperature Monitoring Results and Watershed Condition Indicators for Bull Trout
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River **at the U.S. Geological Survey Gage, 13310800 in the East Fork of the South Fork of the Salmon**

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Appendix 3 - Water Temperature Watershed Condition Indicators and Monitoring Results

Appendix 3 - Water Temperature Watershed Condition Indicators and Monitoring Results

Appendix 3 - Water Temperature Watershed Condition Indicators and Monitoring Results

Figure 54. Water Temperature Monitoring Results and Watershed Condition Indicators Bull Trout at

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Appendix 4: Sediment, Macroinvertebrate, and Fish Tissue Metal Concentration Results

Appendix 4: Sediment, Macroinvertebrate, and Fish Tissue Metal Concentration Laboratory Reports

Stibnite Gold Project Midas Gold Idaho, Inc.

April 2017

Prepared by

MWH Americas, Inc. 727 E. Riverpark Lane, Suite 150 Boise, Idaho 83706

Sediment Metal Concentration Results

www.svl.net

One Government Gulch - PO Box 929 Kellogg ID 83837-0929 (208) 784-1258 Fax (208) 783-0891

1620 W. Fountainhead Pkwy - Suite 202 Tempe, AZ 85282 06-Sep-16 10:59 Reported:

Work Order: **W6H0498** MWH Global (AZ) **Project Name: Stibnite Sediment 2016**

ANALYTICAL REPORT FOR SAMPLES

Solid samples are analyzed on an as-received, wet-weight basis, unless otherwise requested.

Sample preparation is defined by the client as per their Data Quality Objectives.

This report supercedes any previous reports for this Work Order. The complete report includes pages for each sample, a full QC report, and a notes section.

The results presented in this report relate only to the samples, and meet all requirements of the NELAC Standards unless otherwise noted.

Case Narrative: W6H0498

08/23/16 HJG - Standing water was decanted from all samples prior to analysis per client directive.

1 e50 M. Fo

: **holds the following certifications SVL** :C573 **WA**):ID000192015-1, **TNI** (**UT** :ID000192007A, **NV** :ID00019 & ID00965 (Microbiology), **ID** :2080, **CA** :0538, **AZ**

John Kern

-;P76

Laboratory Director

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06-Sep-16 10:59 Tempe, AZ 85282 Reported:

Client Sample ID: **MWH-009**

Sampled: 09-Aug-16 14:30 $19-8nV$ -61 Received: Sampled By: JNE

ork Order: W **W6H0498**

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1620 W. Fountainhead Pkwy - Suite 202

Tempe, AZ 85282

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Page 1 of 1 Sample Report Sample ID: SVL

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Sampled: 10-Aug-16 14:00 $19-8nV$ -61 Received: Sampled By: JNE

> Reported: : Work Order **W6H0498**

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Batch

Laboratory Director

John Kern

-;P76

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I 620 W. Fountainhead Pkwy - Suite 202 06-Sep-16 10:59 Tempe, AZ 85282 Reported:

Client Sample ID: **MWH-012**

ork Order: W **W6H0498**

Sampled: 10-Aug-16 17:30 Received: 19-Aug-16 Sampled By: JNE

NDL major Batch Analyst bosts Notes

Batch

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Metals (Total) by EPA 6000/7000 Methods SMB 09/02/16 07:19 W636023 4.9 8.0 mg/kg 5470 **Aluminum** EPA 6010C الله المسلمية المسلمي
المسلمية المسلمية ا SMB 09/02/16 07:19 W636023 0.6 2.5 mg/kg 200 **Arsenic** EPA 6010C الله المسلم المسلم المعرفي المسلم المعرفي المسلم المسلم المسلم المسلم المسلم المسلم المسلم المسلم المسلم المسلم
المسلم المسلم المس SMB 09/02/16 07:19 W636023 0.07 0.20 mg/kg 0.37 **Beryllium** EPA 6010C الكات المسلمية المسلمية (1020-1120) 61-2003 1170/60 BMS وتمام 1120.09 61.70 Mg/kg f70000 61.20 amax f70109 vd الكات المسابق ا SMB 09/02/16 07:19 W636023 0.06 0.20 mg/kg 0.30 **Cadmium** EPA 6010C SMB 09/02/16 07:19 W636023 3.5 10.0 mg/kg 3310 **Calcium** EPA 6010C الكات المواقع ا
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المواقع المواقع الموا SMB 09/02/16 07:19 W636023 0.16 1.00 mg/kg 6.88 **Copper** EPA 6010C التاب المسابق ا SMB 09/02/16 07:19 W636023 4.0 10.0 mg/kg 8640 **Iron** EPA 6010C SMB 09/02/16 07:19 W636023 0.32 0.50 mg/kg 18.5 **Lanthanum** EPA 6010C SMB 09/02/16 07:19 W636023 0.3 0.8 mg/kg 12.2 **Lead** EPA 6010C SMB 09/02/16 07:19 W636023 0.9 2.0 mg/kg 9.1 **Lithium** EPA 6010C SMB 09/02/16 07:19 W636023 8.7 20.0 mg/kg 2520 **Magnesium** EPA 6010C الكات المستقعة المستق
المستقعة المستقعة ا SMB 09/02/16 07:19 W636023 0.13 0.80 mg/kg 1.11 **Molybdenum** EPA 6010C المجموع التي التي التي التي يتم التي ال
التي يتم التي يتم ا SMB 09/02/16 07:19 W636023 1.1 5.0 mg/kg 1020 **Phosphorus** EPA 6010C SMB 09/02/16 07:19 W636023 13.0 50.0 mg/kg 1530 **Potassium** EPA 6010C SMB 09/02/16 07:19 W636023 0.13 0.20 mg/kg 1.33 **Scandium** EPA 6010C SMB 09/02/16 07:19 W636023 1.4 4.0 mg/kg 4.3 **Selenium** EPA 6010C SMB 09/02/16 07:19 W636023 0.14 0.50 mg/kg 2.00 **Silver** EPA 6010C SMB 09/02/16 07:19 W636023 4.7 50.0 mg/kg 89.9 **Sodium** EPA 6010C الات المسلم العبدية التي تقويم العبدية التي تقويم العبدية التي تقويم التي تقويم العبدية التي تقويم التي تقويم ا
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المستقام المستقام ا SMB 09/02/16 07:19 W636023 0.15 0.50 mg/kg 11.1 **Vanadium** EPA 6010C SMB 09/02/16 07:19 W636023 0.5 1.0 mg/kg 21.9 **Zinc** EPA 6010C D2 08/26/16 11:26 MWD W635142 10 0.038 0.330 mg/kg 11.6 **Mercury** EPA 7471B **Classical Chemistry Parameters** D2,M2 09/02/16 17:02 SM W636152 2 0.22 1.00 % 11.5 **Silica (SiO2)** ASTM D-3682 **TCLP Extraction Parameters** Si CO HG HG ESB 08/30/16 09:15 W635088 % 71.8 **% Dry Solids** EPA 1311 **TCLP Leachates (Metals) Extracted: 08/30/16 09:15** SMB 09/02/16 06:15 W636112 0.008 0.050 mg/L Extract 0.246 **Arsenic** EPA 6010C كال 1.60/07 BMB 21.1969W والمالي المستقل المستقل المستقل المستقل المستقل المستقل المستقل المستقل المستقل المستق الله المواجهة المواجه الي المواقع ال الاق العباد العباد العباد المعلم العبد المستقل المعلم المعلم المعلم العبد المعلم العبد المعلم العبد المعلم الع
العبد العبد ال الكان المواجب ال الحقار المجموعين المج MWD 09/01/16 11:11 W636175 0.000053 0.00020 mg/L Extract < 0.00020 Mercury EPA 7470A This data has been reviewed for accuracy and has been authorized for release by the Laboratory Director or designee.

Analyte Dilution Amaric Disample of the Management Analyte Method School (Analyte Method School of the Method

John Kern Laboratory Director

-;P76

: **holds the following certifications SVL** :C573 **WA**):ID000192015-1, **TNI** (**UT** :ID000192007A, **NV** :ID00019 & ID00965 (Microbiology), **ID** :2080, **CA** :0538, **AZ**

I 620 W. Fountainhead Pkwy - Suite 202 **of Official Anamibis Stiffe Indianal (AZ) Reduced Set CAS Properties Stiffe Indianal (SA) Reduced Stiffe Indianal (SA) Reduced Stiffe Indianal (SA) Reduced Stiffe Indianal (SA) Reduced Stiffe Indianal (SA) R**

06-Sep-16 10:59 Tempe, AZ 85282 Reported:

TCLP Extraction Parameters

TCLP Leachates (Metals) Extracted: 08/30/16 09:15

Client Sample ID: **MWH-013**

ork Order: W **W6H0498**

Sampled: 10-Aug-16 16:00 Received: 19-Aug-16 Sampled By: JNE **L** Sample ID: **W6H0498-05 (Sediment)** \blacksquare **(100)** \blacksquare **CO-8640H3W** \blacksquare GI \blacksquare slqms \blacksquare \blacksquare \blacksquare

Analyte Dilution Amaric Disample of the Management Analyte Method School (Analyte Method School of the Method NDL major Batch Analyst bosts Notes Batch **Metals (Total) by EPA 6000/7000 Methods** الاست المسموعية التي تم ين المركز المعاملة التي تم الت
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et: e0 a1\06\80 are assumed assumed assumed to the study of \log ESB 08/30/16 09:15 W635088 % 52.9 **% Dry Solids** EPA 1311

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This data has been reviewed for accuracy and has been authorized for release by the Laboratory Director or designee. **John Kern** -;P76

Laboratory Director

Project Name: Stingles Stimulus Stips Sti

1620 W. Fountainhead Pkwy - Suite 202 06-Sep-16 10:59 AZ 85282 Tempe, Reported:

Client Sample ID: **MWH-014**

Sampled: 11-Aug-16 10:45

: Work Order **W6H0498**

 $19-8nV$ -61 Received: Sampled By: JNE

Page 1 of 1 of 1 of 10 of 10 cm space 10 cm space 1 cm $\frac{1}{2}$ **cm** $\frac{1}{2$ NDL Diluion Batch Analyst Description Notes Batch

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John Kern Laboratory Director

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: **SVL holds the following certifications** ork order Report Page 7 of 23 W :C573 **WA**):ID000192015-1, **TNI** (**UT** :ID000192007A, **NV** :ID00019 & ID00965 (Microbiology), **ID** :2080, **CA** :0538, **AZ**

1620 W. Fountainhead Pkwy - Suite 202 **Project Names: Stibology Area (AZ)** μ (AZ) μ (Sediment 2016 MWH Global (Sediment 2016 MWH Global (Sediment 2016 MWH Global HWM σ)

06-Sep-16 10:59 AZ 85282 Tempe, Reported:

Client Sample ID: **MWH-016**

Sampled: 11-Aug-16 16:45

Received: 19-Aug-16

: Work Order **W6H0498**

NDL major Batch Analyst bosts Notes Sampled By: JNE

Batch

Analyte Dilution Amaric Disample of the Management Analyte Method School (Analyte Method School of the Method **Sample ID: W6H0498-07 (Sediment) (11 of 12 of 13 sets)** Separation Sample ID: 1 of 1

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تاریخ الله به الله بی الله به ا **Classical Chemistry Parameters** D2 09/02/16 17:02 SM W636152 2 0.22 1.00 % 13.9 **Silica (SiO2)** ASTM D-3682 **TCLP Extraction Parameters** Si = Si Fe \text ESB 08/30/16 09:15 W635088 % 67.8 **% Dry Solids** EPA 1311 **TCLP Leachates (Metals) Extracted: 08/30/16 09:15** SMB 09/02/16 06:31 W636112 0.008 0.050 mg/L Extract < 0.050 Arsenic EPA 6010C SMB 091/20/60 BINIS 211969M 00.000 00.1 12 Marked 20.00 00.1 and 00.1 minus transmitted to 20100 A93 SMB 091/70/60 HWS 711969M 60000 60000 001000 penxs 178 00100 00100 mg/menyed munimped الابان المواجع الاق العباد العباد المسابق المسلس ال الكان المواجب المسابر المواجب الحقاب 16:90 AMB 211969 MG 09:00 002010 002016 002016 16:30 MG 002016 16:30 Mg/L E190 91/70/60 BMS 211969MG 16 ل التي يتم المسلم الموارد التي تم المسلم التي يتم التي تم التي
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John Kern Laboratory Director

-;P76

Project Name: Stingles Stimulus Stips Sti

1620 W. Fountainhead Pkwy - Suite 202 06-Sep-16 10:59 AZ 85282 Tempe, Reported:

Client Sample ID: **MWH-017**

Sampled: 09-Aug-16 12:30 $19-8nV$ -61 Received:

: Work Order **W6H0498**

Sampled By: JNE

NDL Diluion Batch Analyst Description Notes

Batch

Page 1 of 1 of 1 of 10 of 10 cm space 10 cm space 1 cm $\frac{1}{2}$ **cm** $\frac{1}{2$

Analyte Dilution Amaric Disample of the Management Analyte Method School (Analyte Method School of the Method

John Kern Laboratory Director

-;P76

This data has been reviewed for accuracy and has been authorized for release by the Laboratory Director or designee.

I 620 W. Fountainhead Pkwy - Suite 202 06-Sep-16 10:59 Tempe, AZ 85282 Reported: **of Official Strange Sediment 2016 Properties Sediment 201**

ork Order: W **W6H0498**

Sampled: 09-Aug-16 17:00 $19-8$ ug-16 Received: Sampled By: JNE **Page 1 of 19 of 19 of 19 of 11 of 10 of 10 of 11 of 10 of** Client Sample ID: **MWH-029**

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This data has been reviewed for accuracy and has been authorized for release by the Laboratory Director or designee. -;P76

John Kern Laboratory Director

:C573 **WA**):ID000192015-1, **TNI** (**UT** :ID000192007A, **NV** :ID00019 & ID00965 (Microbiology), **ID** :2080, **CA** :0538, **AZ**

: **holds the following certifications SVL**

Project Names: Stibology Area (AZ) Redoit (AS) Redoit Settlem What (XA) Islands 30 HWM (XA) Islands 30 HWM

1620 W. Fountainhead Pkwy - Suite 202 06-Sep-16 10:59 AZ 85282 Tempe, Reported:

Sampled: 13-Aug-16 14:30

Batch

NDL Diluion Batch Analyst Description Notes

 $19-8nV$ -61 Received: Sampled By: JNE

: Work Order **W6H0498**

W6H0498-10 (Sediment) Page 1 of 1 Sample Report Sample ID: SVL Client Sample ID: **MWH-030**

SMB 09/02/16 06:41 W636112 0.0016 0.0500 mg/L Extract < 0.0500 Silver EPA 6010C BE BN 09/11/16 12:36 WMD 10/16 UMN SLAGE ASSOCIATE COOOL 12:36 WERE 200000 0.00020 MMD 10/16 10:36 WERE 91/10/60 MMD 12:36

Analyte Dilution Amaric Disample of the Management Analyte Method School (Analyte Method School of the Method

This data has been reviewed for accuracy and has been authorized for release by the Laboratory Director or designee. -;P76

John Kern Laboratory Director

 F ax (208) 783-0891 F ax 923-0891 F 83831-09291 F 83831-09291 F 83831-09291 F 63891-09 F 63891 F

1620 W

John Kern Laboratory Director

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-;P76

: **SVL holds the following certifications**

Project Name: Stingles Stimulus Stips Sti

1620 W. Fountainhead Pkwy - Suite 202 06-Sep-16 10:59 AZ 85282 Tempe, Reported:

Client Sample ID: **MWH-070**

Sampled: 12-Aug-16 17:30 $19-8nV$ -61 Received:

: Work Order **W6H0498**

Sampled By: JNE

NDL Dinion Batch Analyst Notes

Batch

Sample ID: W6H0498-12 (Sediment) example S9 of 1 of 1 of 1 of 1 of 1

Analyte Dilution Amaric Disample of the Management Analyte Method School (Analyte Method School of the Method

John Kern -;P76

This data has been reviewed for accuracy and has been authorized for release by the Laboratory Director or designee.

Laboratory Director

ork order Report Page 13 of 23 W :C573 **WA**):ID000192015-1, **TNI** (**UT** :ID000192007A, **NV** :ID00019 & ID00965 (Microbiology), **ID** :2080, **CA** :0538, **AZ**

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 1620 W. Fountainhead Pkwy - Suite 202 06-Sep-16 10:59 Tempe, AZ 85282 Reported:

Client Sample ID: **MWH-071**

Sampled: 12-Aug-16 17:00 $19-8nV$ -61 Received:

ork Order: W **W6H0498**

Sampled By: JNE

NDL Diluion Batch Analyst Description Notes

Batch

S μ **PROFICES C-8641000 (1000) C-864000 (1000) Page 1** of 1

Analyte Dilution Amaric Disample of the Management Analyte Method School (Analyte Method School of the Method

Laboratory Director

John Kern

-;P76

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: **holds the following certifications SVL** :C573 **WA**):ID000192015-1, **TNI** (**UT** :ID000192007A, **NV** :ID00019 & ID00965 (Microbiology), **ID** :2080, **CA** :0538, **AZ**

MUD 16:11 91/10/60 UMN SL1969M 6200000 6200000 020000 122113H 6200000 0.00020 123113H

Client Sample ID: **MWH-072**

Project Name: Stingles Stimulus Stips Sti

1620 W. Fountainhead Pkwy - Suite 202 06-Sep-16 10:59 AZ 85282 Tempe, Reported:

Sampled: 12-Aug-16 14:30

NDL Dilution Batch Analyst Reserved MDL 100tes

Batch

 $19-8nV$ -61 Received: Sampled By: JNE

: Work Order **W6H0498**

Sample ID: W6H0498-14 (Sediment) consider the Sample Report Bage 1 of 1

Analyte Dilution Amaric Disample of the Management Analyte Method School (Analyte Method School of the Method

John Kern tu pur

Laboratory Director

ork order Report Page 15 of 23 W :C573 **WA**):ID000192015-1, **TNI** (**UT** :ID000192007A, **NV** :ID00019 & ID00965 (Microbiology), **ID** :2080, **CA** :0538, **AZ**

1620 W. Fountainhead Pkwy - Suite 202 **Project Names: Stibology Area (AZ)** Redoit (AS) Redoit Settlem What (XA) Islands 30 HWM (XA) Islands 30 HWM

06-Sep-16 10:59 AZ 85282 Tempe, Reported:

Client Sample ID: **MWH-073**

Sampled: 12-Aug-16 16:30

 $01-guA-01$: bevised Sampled By: JNE

: Work Order **W6H0498**

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This data has been reviewed for accuracy and has been authorized for release by the Laboratory Director or designee. **John Kern** -;P76

Laboratory Director

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 1620 W. Fountainhead Pkwy - Suite 202 06-Sep-16 10:59 Tempe, AZ 85282 Reported:

Sampled: 13-Aug-16 17:00

ork Order: W **W6H0498**

 $19-8nV$ -61 Received: Sampled By: JNE

NDL Diluion Batch Analyst Description Notes

Batch

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Analyte Dilution Amaric Disample of the Management Analyte Method School (Analyte Method School of the Method

This data has been reviewed for accuracy and has been authorized for release by the Laboratory Director or designee. -;P76

John Kern Laboratory Director

BE:11 91/10/60 CIMN SL19E9M ES0000.0 02000.0 02000.0 pp and 1736 0.00000 0.00020 mm ANGLET ETA 760LET VOLT ES

I 620 W. Fountainhead Pkwy - Suite 202 **of Official Anamibis Stiffe Indianal (AZ) Reduced Set CAS Properties Stiffe Indianal (SA) Reduced Stiffe Indianal (SA) Reduced Stiffe Indianal (SA) Reduced Stiffe Indianal (SA) Reduced Stiffe Indianal (SA) R**

06-Sep-16 10:59 Tempe, AZ 85282 Reported:

Client Sample ID: **MWH-075**

ork Order: W **W6H0498**

Batch

Sampled: 13-Aug-16 12:00 Received: 19-Aug-16 Sampled By: JNE **S L Sample ID: W6H049417 (Sediment) (119mibe S) T1-8640H3W** : CII blqms 2JV2

Analyte Dilution Amaric Disample of the Management Analyte Method School (Analyte Method School of the Method

NDL major Batch Analyst bosts Notes **Metals (Total) by EPA 6000/7000 Methods** الله المسموعي المسموع
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المسابق المسابق المساب SMB 09/02/16 08:25 W636023 4.0 10.0 mg/kg 12300 **Iron** EPA 6010C الله المسابق المسموعي المسموع
المسموعي المسموعي ال الاوات المعالي ا الكات المسابق ا
المسابق المسابق المسا SMB 09/02/16 08:25 W636023 8.7 20.0 mg/kg 2520 **Magnesium** EPA 6010C الكات المجموعة المعلمية المجموعة المجموعة المجموعة المجموعة المجموعة المجموعة المجموعة المجموعة المجموعة المجموع
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التي يقوم التي يقوم SMB 09/02/16 08:25 W636023 1.1 5.0 mg/kg 842 **Phosphorus** EPA 6010C الكات المستوفية المستو
المستوفية المستوفية المستوفية المستوفية المستوفية المستوفية المستوفية المستوفية المستوفية المستوفية المستوفي الله المسلم المسلم المعلمية المسلمين التي تقارب المسلمين المسلمين المسلمين المسلمين المسلمين المسلمين المسلمين ا
المسلمين المسلمين الله المسابق ال SMB 09/02/16 08:25 W636023 0.14 0.50 mg/kg 0.95 **Silver** EPA 6010C التار التي تي ا
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التي تم التي ت الله المسابر ا الله المسابق ا الله المسموعية المسمو
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المسابق المسابق المسا الله المسابق ال
المسابق المسابق المسا MWD 08/26/16 11:21 W635142 0.004 0.033 mg/kg 0.447 **Mercury** EPA 7471B **Classical Chemistry Parameters** D2 09/02/16 17:02 SM W636152 2 0.22 1.00 % 14.6 **Silica (SiO2)** ASTM D-3682 **TCLP Extraction Parameters** et: e0 a1\06\80 are assumed assumed assumed to the study of the study of the participate of the participate of $\frac{1}{2}$ ESB 08/30/16 09:15 W635088 % 58.0 **% Dry Solids** EPA 1311 **TCLP Leachates (Metals) Extracted: 08/30/16 09:15** SMB 09/02/16 07:26 W636112 0.008 0.050 mg/L Extract 0.061 **Arsenic** EPA 6010C الكات المواقع ا الله المسابق ا الله المسابق المسترق المستقل المسابق المستقل المستقل المستقل المستقل المستقل المستقل المستقل المستقل المستقل ا الاق العباد العباد العباد المعلم العبد المستقل المعلم المعلم المعلم العبد المعلم العبد المعلم العبد المعلم الع الكات المواجب المستقرر المواجب المواج الحقار المجموع التي يوم المجموع ال MWD 09/01/16 11:38 W636175 0.000053 0.00020 mg/L Extract < 0.00020 Mercury EPA 7470A

John Kern Laboratory Director

This data has been reviewed for accuracy and has been authorized for release by the Laboratory Director or designee.

-;P76

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One Government Gulch - PO Box 929 Kellogg ID 83837-0929 (208) 784-1258 Fax (208) 783-0891

1620 W. Fountainhead Pkwy - Suite 202 Tempe, AZ 85282 06-Sep-16 10:59 Reported:

Work Order: **W6H0498** MWH Global (AZ) **Project Name: Stibnite Sediment 2016**

SVL holds the following certifications:

AZ:0538, **CA**:2080, **ID**:ID00019 & ID00965 (Microbiology), **NV**:ID000192007A, **UT**(**TNI**):ID000192015-1, **WA**:C573 Work order Report Page 19 of 23

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1620 W. Fountainhead Pkwy - Suite 202 Tempe, AZ 85282 06-Sep-16 10:59 Reported: Work Order: **W6H0498** MWH Global (AZ) **Project Name: Stibnite Sediment 2016**

SVL holds the following certifications:

AZ:0538, **CA**:2080, **ID**:ID00019 & ID00965 (Microbiology), **NV**:ID000192007A, **UT**(**TNI**):ID000192015-1, **WA**:C573 Work order Report Page 20 of 23

Macroinvertebrate Metal Concentration Results

ALS Group USA, Corp 1317 South 13th Avenue Kelso, WA 98626 +1 360 577 7222 **T : F** : +1 360 636 1068 ALS Environmental **www.alsglobal.com**

October 10, 2016 **Analytical Report for Service Request No:** K1609659

Jeremy Collyard MWH Global, Inc 1620 West Fountainhead Parkway Suite 202 Tempe, AZ 85282

RE: Metals in Sediment and Tissue 2016 / 10509465.2000

Dear Jeremy,

For your reference, these analyses have been assigned our service request number **K1609659.** Enclosed are the results of the sample(s) submitted to our laboratory August 19, 2016

Analyses were performed according to our laboratory's NELAP-approved quality assurance program. The test results meet requirements of the current NELAP standards, where applicable, and except as noted in the laboratory case narrative provided. For a specific list of NELAP-accredited analytes, refer to the certifications section at www.alsglobal.com. All results are intended to be considered in their entirety, and ALS Group USA Corp. dba ALS Environmental (ALS) is not responsible for use of less than the complete report. Results apply only to the items submitted to the laboratory for analysis and individual items (samples) analyzed, as listed in the report.

Please contact me if you have any questions. My extension is 3376. You may also contact me via email at gregory.salata@alsglobal.com.

Respectfully submitted,

ALS Group USA, Corp. dba ALS Environmental

yang xhilit

Gregory Salata, Ph.D. Senior Project Manager

www.alsglobal.com ALS Environmental **F :** +1 360 636 1068 **T :** +1 360 577 7222 Kelso, WA 98626 1317 South 13th Avenue ALS Group USA, Corp

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Acronyms

Inorganic Data Qualifiers

- ***** The result is an outlier. See case narrative.
- **#** The control limit criteria is not applicable. See case narrative.
- B The analyte was found in the associated method blank at a level that is significant relative to the sample result as defined by the DOD or NELAC standards.
- E The result is an estimate amount because the value exceeded the instrument calibration range.
- J The result is an estimated value.
- U The analyte was analyzed for, but was not detected ("Non-detect") at or above the MRL/MDL. *DOD-QSM 4.2 definition* : Analyte was not detected and is reported as less than the LOD or as defined by the project. The detection limit is adjusted for dilution.
- i The MRL/MDL or LOQ/LOD is elevated due to a matrix interference.
- X See case narrative.
- Q See case narrative. One or more quality control criteria was outside the limits.
- H The holding time for this test is immediately following sample collection. The samples were analyzed as soon as possible after receipt by the laboratory.

Metals Data Qualifiers

- **#** The control limit criteria is not applicable. See case narrative.
- J The result is an estimated value.
- E The percent difference for the serial dilution was greater than 10%, indicating a possible matrix interference in the sample.
- M The duplicate injection precision was not met.
- N The Matrix Spike sample recovery is not within control limits. See case narrative.
- S The reported value was determined by the Method of Standard Additions (MSA).
- U The analyte was analyzed for, but was not detected ("Non-detect") at or above the MRL/MDL.
- *DOD-QSM 4.2 definition* : Analyte was not detected and is reported as less than the LOD or as defined by the project. The detection limit is adjusted for dilution.
- W The post-digestion spike for furnace AA analysis is out of control limits, while sample absorbance is less than 50% of spike absorbance.
- i The MRL/MDL or LOQ/LOD is elevated due to a matrix interference.
- X See case narrative.
- + The correlation coefficient for the MSA is less than 0.995.
- Q See case narrative. One or more quality control criteria was outside the limits.

Organic Data Qualifiers

- ***** The result is an outlier. See case narrative.
- **#** The control limit criteria is not applicable. See case narrative.
- A A tentatively identified compound, a suspected aldol-condensation product.
- B The analyte was found in the associated method blank at a level that is significant relative to the sample result as defined by the DOD or NELAC standards.
- C The analyte was qualitatively confirmed using GC/MS techniques, pattern recognition, or by comparing to historical data.
- D The reported result is from a dilution.
- E The result is an estimated value.
- J The result is an estimated value.
- N The result is presumptive. The analyte was tentatively identified, but a confirmation analysis was not performed.
- P The GC or HPLC confirmation criteria was exceeded. The relative percent difference is greater than 40% between the two analytical results.
- U The analyte was analyzed for, but was not detected ("Non-detect") at or above the MRL/MDL. *DOD-QSM 4.2 definition* : Analyte was not detected and is reported as less than the LOD or as defined by the project. The detection limit is adjusted for dilution.
- i The MRL/MDL or LOQ/LOD is elevated due to a chromatographic interference.
- X See case narrative.
- Q See case narrative. One or more quality control criteria was outside the limits.

Additional Petroleum Hydrocarbon Specific Qualifiers

- F The chromatographic fingerprint of the sample matches the elution pattern of the calibration standard.
- L The chromatographic fingerprint of the sample resembles a petroleum product, but the elution pattern indicates the presence of a greater amount of lighter molecular weight constituents than the calibration standard.
- H The chromatographic fingerprint of the sample resembles a petroleum product, but the elution pattern indicates the presence of a greater amount of heavier molecular weight constituents than the calibration standard.
- O The chromatographic fingerprint of the sample resembles an oil, but does not match the calibration standard.
- Y The chromatographic fingerprint of the sample resembles a petroleum product eluting in approximately the correct carbon range, but the elution pattern does not match the calibration standard.
- Z The chromatographic fingerprint does not resemble a petroleum product.

Page 4 of 538

ALS Group USA Corp. dba ALS Environmental (ALS) - Kelso State Certifications, Accreditations, and Licenses

Analyses were performed according to our laboratory's NELAP-approved quality assurance program. A complete listing of specific NELAP-certified analytes, can be found in the certification section at www.ALSGlobal.com or at the accreditation bodies web site.

Please refer to the certification and/or accreditation body's web site if samples are submitted for compliance purposes. The states highlighted above, require the analysis be listed on the state certification if used for compliance purposes and if the method/anlayte is offered by that state.

Case Narrative

ALS Environmental—Kelso Laboratory 1317 South 13th Avenue, Kelso, WA 98626 Phone (360)577-7222 Fax (360)636-1068 www.alsglobal.com

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ALS ENVIRONMENTAL

Client: MWH Global, Inc **Service Request No.:** K1609659 **Project:** Metals in Sediment and Tissue 2016/ **Date Received:** 08/19/16 10509465.2000 **Sample Matrix:** Animal Tissue

Case Narrative

All analyses were performed consistent with the quality assurance program of ALS Environmental. This report contains analytical results for samples designated for Tier IV validation deliverables including summary forms and all of the associated raw data for each of the analyses. When appropriate to the method, method blank results have been reported with each analytical test.

Sample Receipt

Thirteen animal tissue samples were received for analysis at ALS Environmental on 08/19/16. The samples were received in good condition and consistent with the accompanying chain of custody form. The samples were stored frozen at –20ºC upon receipt at the laboratory.

Lipids

No anomalies associated with the analysis of these samples were observed.

Total Metals

Matrix Spike Recovery Exceptions:

The control criteria for matrix spike recovery of Aluminum, Calcium, Iron, Magnesium, Manganese, Potassium, and Sodium for sample MWH-011 were not applicable. The analyte concentration in the sample was significantly higher than the added spike concentration, preventing accurate evaluation of the spike recovery.

The matrix spike recovery of Antimony for sample MWH-011 was outside control criteria. Recovery in the Laboratory Control Sample (LCS) was acceptable, which indicated the analytical batch was in control. The matrix spike outlier suggested a potential low bias in this matrix. No further corrective action was appropriate.

Relative Percent Difference Exceptions:

The Relative Percent Difference (RPD) for the replicate analysis of Iron in sample MWH-011 was outside the normal ALS control limits (23% RPD versus a control limit of 20%). The samples were homogenized, freeze dried, then ground prior to digestion, however this was not sufficient to achieve a completely uniform distribution of Iron in the tissue.

No other anomalies associated with the analysis of these samples were observed.

Approved by **August Allette**

Chain of Custody

ALS Environmental—Kelso Laboratory 1317 South 13th Avenue, Kelso, WA 98626 Phone (360)577-7222 Fax (360)636-1068 www.alsglobal.com

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J.

Total Solids

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ALS Group USA, Corp. dba ALS Environmental

Analytical Report

Client: Animal Tissue **Project:** Metals in Sediment and Tissue 2016/10509465.2000 **Date Collected:** 08/09/16 - 08/13/16 MWH Global, Inc **Sample Matrix:** Freeze Dry **Prep Method:** None **Analysis Method:**

08/19/16 **Date Received: Service Request:** K1609659 Date Collected: $08/09/16 - 08/13/16$ **Units:** Percent

Basis: Wet

Total Solids

LIPIDS

ALS Environmental—Kelso Laboratory 1317 South 13th Avenue, Kelso, WA 98626 Phone (360)577-7222 Fax (360)636-1068 www.alsglobal.com

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ALS ENVIRONMENTAL

Analytical Report

ALS ENVIRONMENTAL

QA/QC Report

Service Request: K1609659 **Pate Collected:** N/A **Date Received:** N/A **Date Extracted:** 9/22/2016 **Date Analyzed:** 10/3/2016

Triplicate Summary Lipids, Total

Sample Name: Batch QC Units: PERCENT Lab Code: K1609528-016 TRP Basis: Wet Weight Test Notes:

Duplicate Triplicate Percent Relative Prep Analysis Sample Sample Sample Standard Result Analyte Method Method MRL Result Result Result Average Deviation Notes Lipids, Total **EPA 3541 NOAA** 0.03 1.1 1.2 1.3 1.2 10

ALS ENVIRONMENTAL

QA/QC Report

Client: MWH Global, Inc **Service Request:** K1609659 **Project:** Metals in Sediment and Tissue 2016/10509465.2000 **Date Collected:** NA **Project:** Metals in Sediment and Tissue 2016/10509465.2000
Matrix: Tissue

Date Received: NA Date Extracted: 9/22/2016 **Date Analyzed:** 10/3/2016

Laboratory Control Sample Lipids, Total

Sample Name: KWG1607862-3 LCS Units: PERCENT

Test Notes:

Basis: Wet Weight

Metals

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ALS Group USA, Corp. dba ALS Environmental Analytical Report

Client: MWH Global, Inc **Service Request:** K1609659 **Project:** Metals in Sediment and Tissue 2016/10509465.2000 **Service Request: Service Request:** K1609659 **Project:** Metals in Sediment and Tissue 2016/10509465.2000 **Project:** Metals in Sediment and Tissue 2016/10509465.2000 **Sample Matrix:** Animal tissue **Date Received: 08/19/16**

Mercury, Total

Prep Method: METHOD Units: ng/g Analysis Method: 1631E Basis: Wet Test Notes:

ALS Group USA, Corp. dba ALS Environmental Analytical Report

Client: MWH Global, Inc **Service Request:** K1609659 **Project:** Metals in Sediment and Tissue 2016/10509465.2000 **Date Collected:** 08/11/16 Metals in Sediment and Tissue 2016/10509465.2000 **Sample Matrix:** Animal tissue **Date Received: 08/19/16**

Mercury, Total

ALS Group USA, Corp. dba ALS Environmental QA/QC Report

ALS Group USA, Corp. dba ALS Environmental

QA/QC Report

ALS Group USA, Corp. dba ALS Environmental

QA/QC Report

ALS Group USA, Corp. dba ALS Environmental QA/QC Report

Sample Name: MWH-007 **Lab Code: K1609659-001**

Sample Name: MWH-009 **Lab Code: K1609659-002**

Sample Name: MWH-011 **Lab Code: K1609659-003**

Sample Name: MWH-012 **Lab Code: K1609659-004**

Sample Name: MWH-014 **Lab Code: K1609659-005**

Sample Name: MWH-016 **Lab Code: K1609659-006**

Sample Name: MWH-017 **Lab Code: K1609659-007**

Sample Name: MWH-029 **Lab Code: K1609659-008**

Sample Name: MWH-030 **Lab Code: K1609659-009**

Sample Name: MWH-069 **Lab Code: K1609659-010**

Sample Name: MWH-074 **Lab Code: K1609659-011**

Sample Name: MWH-075 **Lab Code: K1609659-012**

Sample Name: MWH-013 **Lab Code: K1609659-013**

Sample Name: Method Blank **Lab Code: KQ1610520-01**

Sample Name: MWH-007 **Lab Code: K1609659-001**

Sample Name: MWH-009 **Lab Code: K1609659-002**

Sample Name: MWH-011 **Lab Code: K1609659-003**

Sample Name: MWH-012 **Lab Code: K1609659-004**

Sample Name: MWH-014 **Lab Code: K1609659-005**

Sample Name: MWH-016 **Lab Code: K1609659-006**

Sample Name: MWH-017 **Lab Code: K1609659-007**

Sample Name: MWH-029 **Lab Code: K1609659-008**

Sample Name: MWH-030 **Lab Code: K1609659-009**

Sample Name: MWH-069 **Lab Code: K1609659-010**

Sample Name: MWH-074 **Lab Code: K1609659-011**

Sample Name: MWH-075 **Lab Code: K1609659-012**

Sample Name: MWH-013 **Lab Code: K1609659-013**

Sample Name: Method Blank **Lab Code: KQ1610520-01**

INITIAL AND CONTINUING CALIBRATION VERIFICATION

 Concentration Units: ug/L

Client: MWH Global, Inc

Service Request: K1609659

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

ICV Source: CCV Source: Inorganic Ventures ALS MIXED

INITIAL AND CONTINUING CALIBRATION VERIFICATION

Client: MWH Global, Inc

Service Request: K1609659

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

 Concentration Units: ug/L

INITIAL AND CONTINUING CALIBRATION VERIFICATION

Client: MWH Global, Inc

Service Request: K1609659

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

ICV Source: CCV Source: Inorganic Ventures ALS MIXED

INITIAL AND CONTINUING CALIBRATION VERIFICATION

Client: MWH Global, Inc

Service Request: K1609659

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

ICV Source: CCV Source: Inorganic Ventures ALS MIXED

ALS Group USA, Corp.

dba ALS Environmental

- 2a - LOW LEVEL INITIAL CALIBRATION AND LOW LEVEL CONTINUING CALIBRATION VERIFICATION Metals

ALS Group USA, Corp.

dba ALS Environmental

- 2a - LOW LEVEL INITIAL CALIBRATION AND LOW LEVEL CONTINUING CALIBRATION VERIFICATION Metals

ALS Group USA, Corp.

dba ALS Environmental

Metals

LOW LEVEL INITIAL CALIBRATION AND LOW LEVEL CONTINUING CALIBRATION VERIFICATION

 - 3 -

BLANKS

Client: MWH Global, Inc

Service Request: K1609659

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

Preparation Blank Matrix (soil/water): WATER

Preparation Blank Concentration Units (ug/L or mg/kg): ug/L

 - 3 -

BLANKS

Client: MWH Global, Inc

Service Request: K1609659

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

Preparation Blank Matrix (soil/water): WATER

Preparation Blank Concentration Units (ug/L or mg/kg): ug/L

 - 3 -

BLANKS

Client: MWH Global, Inc

Service Request: K1609659

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

Preparation Blank Matrix (soil/water): WATER

Preparation Blank Concentration Units (ug/L or mg/kg): ug/L

 - 4 -

ICP INTERFERENCE CHECK SAMPLE

Client: MWH Global, Inc

Service Request: K1609659

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

ICP ID Number: K-ICP-AES-04 1CS Source:

Inorganic Ventures

Concentration Units): ug/L

 - 4 -

ICP INTERFERENCE CHECK SAMPLE

Client: MWH Global, Inc

Service Request: K1609659

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

ICP ID Number: K-ICP-MS-04 1CS Source:

Inorganic Ventures

Concentration Units): ug/L

- 5A - SPIKE SAMPLE RECOVERY Metals

Sample Name: MWH-011S Lab Code: K1609659-003S

- 5B - POST SPIKE SAMPLE RECOVERY Metals

Sample Name: MWH-011A Lab Code: K1609659-003A

- 6 - Metals

DUPLICATES

- 7 - LABORATORY CONTROL SAMPLE Metals

Client: MWH Global, Inc

Service Request: K1609659

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

Aqueous LCS Source: **ALS MIXED**

Solid LCS Source: ERA D065540

ALS Group USA, Corp. dba ALS Environmental QA/QC Report

Total Metals

Sample Name: Standard Reference Material Units: mg/Kg (ppm)

Lab Code: K1609659-SRM1 Basis: Dry Lab Code: K1609659-SRM1 Basis: Dry
Test Notes: Dorm-4 Solids = 94.5% Dorm-4 Solids $= 94.5\%$

Source: N.R.C.C. Dorm-4

ALS Group USA, Corp. dba ALS Environmental QA/QC Report

Service Request: K1609659 **Pate Collected:** NA **Date Received: NA Date Extracted:** 09/01/16 **Date Analyzed:** 09/07,27/16

Standard Reference Material Summary Total Metals

Sample Name: Standard Reference Material
Lab Code: $K1609659-SRM2$ Lab Code: K1609659-SRM2
Test Notes: Tort-3 Solids = 99 Tort-3 Solids $= 99.1\%$

Source: N.R.C.C. Tort-3

- 9 -

ICP SERIAL DILUTIONS

Client: MWH Global, Inc **Service Request:** K1609659

Project No.: 10509465.2000 **Units:** UG/L

Project Name: Metals in Sediment and Tissue 20

Sample Name: MWH-011L Lab Code: K1609659-003L

dba ALS Environmental

Metals

- 10 -

DETECTION LIMITS

Client: MWH Global, Inc

Service Request: K1609659

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

ICP/ICP-MS ID #:

GFAA ID #: AA ID #:

dba ALS Environmental

Metals

- 10 -

DETECTION LIMITS

Client: MWH Global, Inc

Service Request: K1609659

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

ICP/ICP-MS ID #: K-ICP-MS-04

GFAA ID #: AA ID #:

- 11A -

ICP INTERELEMENT CORRECTION FACTORS

Client: MWH Global, Inc

Service Request: K1609659

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

- 11A -

ICP INTERELEMENT CORRECTION FACTORS

Client: MWH Global, Inc

Service Request: K1609659

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

- 11B -

ICP INTERELEMENT CORRECTION FACTORS

Client: MWH Global, Inc

Service Request: K1609659

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

- 11B -

ICP INTERELEMENT CORRECTION FACTORS

Client: MWH Global, Inc

Service Request: K1609659

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

- 11B -

ICP INTERELEMENT CORRECTION FACTORS

Client: MWH Global, Inc

Service Request: K1609659

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

- 11B -

ICP INTERELEMENT CORRECTION FACTORS

Client: MWH Global, Inc

Service Request: K1609659

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

- 11B -

ICP INTERELEMENT CORRECTION FACTORS

Client: MWH Global, Inc

Service Request: K1609659

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

- 11B -

ICP INTERELEMENT CORRECTION FACTORS

Client: MWH Global, Inc

Service Request: K1609659

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

-12- ICP LINEAR RANGES (QUARTERLY) Metals

Client: MWH Global, Inc

Service Request: K1609659

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

Analyte Integ. Time (Sec.) Concentration (ug/L) Method ICP ID Number: K-ICP-AES-04 Calcium 450000 15.000 6010C Iron 360000 15.000 6010C Magnesium 15.000 90000 6010C Manganese 15.000 180000 6010C Potassium | 15.000 | 450000 | 6010C **Sodium 450000 15.000 6010C**

-12- ICP LINEAR RANGES (QUARTERLY) Metals

Client: MWH Global, Inc

Service Request: K1609659

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

Analyte Integ. Time (Sec.) Concentration (ug/L) Method ICP ID Number: K-ICP-MS-04 Aluminum 45.000 18000 6020A Antimony 45.000 900 6020A Arsenic 45.000 3000 6020A Barium 3000 45.000 6020A Beryllium 900 45.000 6020A Cadmium 3000 45.000 6020A Chromium 3000 45.000 6020A Cobalt 3000 45.000 6020A Copper 3000 45.000 6020A Lead 3000 45.000 6020A Nickel 45.000 3000 6020A Selenium 3000 45.000 6020A Silver | 45.000 | 900 | 6020A **Thallium 3000 45.000 6020A Vanadium 45.000 3000 6020A Zinc 3000 45.000 6020A**
-13- PREPARATION LOG Metals

Client: MWH Global, Inc

Service Request: K1609659

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

Method: P

-13- PREPARATION LOG Metals

Client: MWH Global, Inc

Service Request: K1609659

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

Method: MS

Metals - 14 -

ANALYSIS RUN LOG Service Request: K1609659

Client: MWH Global, Inc

Run Number: 092716AICP04

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

Metals - 14 -

ANALYSIS RUN LOG Service Request: K1609659

Client: MWH Global, Inc

Run Number: 092716AICP04

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

Metals - 14 -

ANALYSIS RUN LOG Service Request: K1609659

Client: MWH Global, Inc

Run Number: 090716a1

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

Metals - 14 -

ANALYSIS RUN LOG Service Request: K1609659

Client: MWH Global, Inc

Run Number: 090716a1

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

Metals - 14 -

ANALYSIS RUN LOG Service Request: K1609659

Client: MWH Global, Inc

Run Number: 090716a1

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

Metals - 14 -

ANALYSIS RUN LOG Service Request: K1609659

Client: MWH Global, Inc

Run Number: 090716a1

Project No.: 10509465.2000

Project Name: Metals in Sediment and Tissue 20

15-IN

ICP-MS INTERNAL STANDARDS RELATIVE INTENSITY SUMMARY

15-IN

ICP-MS INTERNAL STANDARDS RELATIVE INTENSITY SUMMARY

K1609659-007 MWH-017 2048 109 94 90 94 103 94

15-IN

ICP-MS INTERNAL STANDARDS RELATIVE INTENSITY SUMMARY

K1609659-012 MWH-075 2113 108 94 91 93 103 93 K1609659-013 MWH-013 2117 106 93 89 93 100 91

CCV8 CCV 2127 104 92 90 94 97 91

CCB8 CCB 2137 104 93 90 93 99 92 ICS-A2 ICSA 2141 99 84 82 82 89 85 ICS-AB2 ICSAB 2146 98 83 80 81 90 84

ZZZZZZ ZZZZZZ 2122

ZZZZZZ ZZZZZZ 2132

 FORM XV-IN

Page 94 of 538

15-IN

ICP-MS INTERNAL STANDARDS RELATIVE INTENSITY SUMMARY

Lab Name: ALS Group USA, Corp.

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Lab Code: <u>ALSK ____</u> Case No.: ______________ NRAS No.: ____________ SDG NO.: <u>K1609659</u>

10509465.2000 Contract: 20509465.2000

ICP-MS Instrument ID: K-ICP-MS-04 Start Date: End Date: 09/07/2016 09/07/2016

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15-IN

ICP-MS INTERNAL STANDARDS RELATIVE INTENSITY SUMMARY

Lab Name: ALS Group USA, Corp.

Lab Code: <u>ALSK ____</u> Case No.: ______________ NRAS No.: ____________ SDG NO.: <u>K1609659</u>

10509465.2000 Contract: 10509465.2000

ICP-MS Instrument ID: K-ICP-MS-04 Start Date: End Date: 09/07/2016 09/07/2016

15-IN

ICP-MS INTERNAL STANDARDS RELATIVE INTENSITY SUMMARY

Lab Code: <u>ALSK ____</u> Case No.: ______________ NRAS No.: ____________ SDG NO.: <u>K1609659</u> Lab Name: ALS Group USA, Corp.

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10509465.2000 Contract: 10509465.2000

ICP-MS Instrument ID: K-ICP-MS-04 Start Date: End Date: 09/07/2016 09/07/2016

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Raw Data

ALS Environmental—Kelso Laboratory 1317 South 13th Avenue, Kelso, WA 98626 Phone (360)577-7222 Fax (360)636-1068 www.alsglobal.com

RIGHT SOLUTIONS | RIGHT PARTNER

Page 98 of 538

Total Solids

ALS Environmental—Kelso Laboratory 1317 South 13th Avenue, Kelso, WA 98626 Phone (360)577-7222 Fax (360)636-1068 www.alsglobal.com

RIGHT SOLUTIONS | RIGHT PARTNER

Page 99 of 538

Service Request Number(s):

Blank Data

 9659

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R\:ICP\misc\digforms\TfSSUE Aliquot Bench Sheets1

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Date:

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R\:ICP\misc\digforms\TISSUE Aliquot Bench Sheets1

Service Request Number(s):

K1609659

Analysis for: Lipids

ALIQUOT DATA

<u>dlwitten Meger Winn</u> Reviewed:

ALS Environmental, Inc.

Service Request #: Analysis For:

Freeze Dried Solids

R:ICP\MISC\DIGFORMS\FISH TISSUE ALIQUOT BENCH SHEETS

ALS Inc. Service Request Number(s):
K *to q 46 5 9*

Analysis for:

Analyst:

Reviewed:

R\:ICP\misc\digforms\TISSUE Aliquot Bench Sheets1

 $8/24/16$

Date: \mathcal{V}_2 \mathcal{U}_4

Date:

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ALS Inc.

Service Request Number(s):

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KI60 9557 KI609659

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Service Request Number(s):

Blank Data

 14659

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Benchsheet

Comments: Data entered 8-29-16 L.J. Reviewed 8/29/16 A.C.

LIPIDS

ALS Environmental—Kelso Laboratory 1317 South 13th Avenue, Kelso, WA 98626 Phone (360)577-7222 Fax (360)636-1068 www.alsglobal.com

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Page 107 of 538

Lipids Raw Benchsheet

Reviewed By: Valla

Reviewed By: Collar

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ALS Environmental—Kelso Laboratory 1317 South 13th Avenue, Kelso, WA 98626 Phone (360)577-7222 Fax (360)636-1068 www.alsglobal.com

RIGHT SOLUTIONS | RIGHT PARTNER

Page 110 of 538

Service Request #: K1607170, K1609659

 MS/MSD with $#$: StarLims Run #:

1631 Tissue Data Review Form

Comments

Batch Information Report

Batch Number: Method Number: EPA 1631E

Project Number(s): Date Analyzed: 9/7/16

Instrument ID: K-AFS-04 **Analyst Name: Brian Sheldon**

Analyst Comments:

PMT: 525 OFFSET: 3,039 NOISE: 36

Run Reporl

Batch Number: Method Number: EPA 1631E

Project Number(s):
 Project Number(s):
 Instrument ID: K-AFS-04
 EXECUTE:
 Date Analyst Name: Brian S

Analyst Name: Brian Sheldon

Page 1 of 5 (Run Report)

Run Report

Batch Number: Method Number: EPA 1631E

Project Number(s): **Date Analyzed: 9/7/16**

Instrument ID: K-AFS-04 **Analyst Name:** Brian Sheldon

Page 2 of 5 (Run Report)

Run Reporl

Batch Number: Method Number: EPA 1631E

Project Number(s):

Instrument ID: K-AFS-04
 Project Number(s):
 Date Analyzed: 9/7/16
 Analyst Name: Brian S

Analyst Name: Brian Sheldon

Page 3 of 5 (Run Report)

Run Report

Batch Number: Method Number: EPA 1631E

Project Number(s): Instrument ID: K-AFS-04 Date Analyzed: 9/7/16 Analyst Name: Brian Sheldon

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Page 4 of 5 (Run Report)

Run Report

Batch Number: Method Number: EPA 1631E

Project Number(s): Instrument ID: K-AFS-04

Date Analyzed: 9/7/16 Analyst Name: Brian Sheldon

Analyst Comments:

PMT: 525 OFFSET: 3,039 NOISE: 36 $\ddot{}$

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QA Summary Report

Batch Number: Method Number: EPA 1631E

Project Number(s): Instrument ID: K-AFS-04 Date Analyzed: 9/7/16 Analyst Name: Brian Sheldon

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QA Summary Report

Batch Number: Method Number: EPA 1631E

Project Number(s): Instrument ID: K-AFS-04

Date Analyzed: 9/7/16 Analyst Name: Brian Sheldon

Page 2 of 5 (QA Summary Report)
QA Summary Report

Batch Number: Method Number: EPA 1631E

Project Number(s): Instrument ID: K-AFS-04

Date Analyzed: 9/7/16 Analyst Name: Brian Sheldon

QA Summary Report

Batch Number: Method Number: EPA 1631E

Project Number(s): Instrument ID: K-AFS-04

Date Analyzed: 9/7/16 Analyst Name: Brian Sheldon

QA Summary Report

Batch Number: Method Number: EPA 1631E

Project Number(s): Instrument ID: K-AFS-04

Date Analyzed: 9/7/16 Analyst Name: Brian Sheldon

QA Comments:

Page 5 of 5 (QA Summary Report)

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Sample Results Summary Report

Batch Number: Method Number: EPA 1631E

Project Number(s): Instrument ID: K-AFS-04 Date Analyzed: 9/7/16 **Analyst Name:** Brian Sheldon

Batch Number: Method Number: EPA 1631E

Project Number(s):
 Project Number(s):
 Instrument ID: K-AFS-04 **Analyzed:** 9/7/16 **Analyst Name:** Brian S

Analyst Name: Brian Sheldon

Page 1 of 5 (Run Information Report)

Batch Number: Method Number: EPA 1631E

Project Number(s):

Instrument ID: K-AFS-04 **Example 20 and Server Analyzed: 2016**

Analyst Name: Brian S

Analyst Name: Brian Sheldon

Page 2 of 5 (Run Information Report)

Batch Number: Method Number: EPA 1631E

Project Number(s):
 Project Number(s):
 Instrument ID: K-AFS-04
 Date Analyst Name: Brian S

Analyst Name: Brian Sheldon

Page 3 of 5 (Run Information Report)

Batch Number: Method Number: EPA 1631E

Project Number(s): Project Number(s): Date Analyzed: 9/7/16

Instrument ID: K-AFS-04 Analyst Name: Brian Sheldon

Page 4 of 5 (Run Information Report)

Batch Number: Method Number: EPA 1631E

Project Number{s):

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Date Analyzed: 9/7/16 Analyst Name: Brian Sheldon

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Page 5 of 5 (Run Information Report)

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Tissue Dry Wt. MRL and MDL Calculations:

Standard MRL = 1.0 Standard MDL= 0.06

Standard Dilution = 1

Standard Sample Mass = 0.40

Weight & **Dilution Adjuste,**

Conversion from dry weight to wet weight:

Standard MRL = 1.0 Standard $MDL = 0.06$ Standard Dilution $=$ 1

Standard Sample Mass = 0.40

Dry Percent Wet Sample I.D. Weight Solids Weight Dilution MRL MDL K 1609659-001 0.398 17.6 **2.261** 5 **0.88 0.05 K** 1609659-002 0.194 6.6 **2.953** 5 **0.68 0.04 K** 1609659-003 0.404 17.0 **2.376** 5 **0.84 0.05 K** 1609659-004 0.406 16.5 **2.461** 5 **0.81 0.05** K1609659-005 0.402 14.1 **2.851** 5 **0.70 0.04 K** 1609659-006 0.411 14.4 **2.854** 5 **0.70 0.04 K** 1609659-007 0.403 15.8 **2.551** 5 **0.78 0.05 K** 1609659-008 0.410 16.4 **2.500** 50 **8.00 0.48** K1609659-009 0.405 19.3 **2.098** 5 **0.95 0.06** K1609659-010 0.352 18.0 **1.956** 50 **10.23 0.61** K1609659-011 0.417 10.7 **3.897** 5 **0.51 0.03** K1609659-012 0.411 15.7 **2.618** 5 **0.76 0.05** K 1609659-013 0.201 15.3 **1.314** 5 **1.52 0.09 #DIV/0I** 1 **#DIV/0I #DIV/0I #DIV/0I** 1 **#DIV/0I #DIV/0I #DIV/0I** 1 **#DIV/0I #DIV/0I #DIV/0I** 1 **#DIV/0I #DIV/0I #DIV/0I** 1 **#DIV/0I #D!V/0I #DIV/0I** 1 **#DIV/0I #DIV/0I #DIV/0I** 1 **#DIV/0I #DIV/0I** Method Blank 0.400 15.000 **2.667** 1 **0.2 0.01**

Weight & **Dilution Adjuste,**

Batch Number: Method Number: EPA 1631E

Date Analyzed: 9/7/16 Analyst Name: Brian Sheldon

Page 1 of 5 (Peak Report)

Batch Number: Method Number: EPA 1631E

Project Number(s): Instrument ID: K-AFS-04 Date Analyzed: 9/7/16 Analyst Name: Brian Sheldon

Page 2 of 5 (Peak Report)

Batch Number: Method Number: EPA 1631E

Project Number(s): Instrument ID: K-AFS-04 Date Analyzed: 9/7/16 Analyst Name: Brian Sheldon

Page 3 of 5 (Peak Report)

Batch Number: Method Number: EPA 1631E

Project Number(s): Instrument ID: K-AFS-04 **Date Analyzed:** 9/7/16 **Analyst Name:** Brian Sheldon

Page 4 of 5 (Peak Report)

Mercury Guru ver **Page©19952 UF3**8 rooks Rand Inc

Batch Number: Method Number: EPA 1631E

Project Number(s): Instrument ID: K-AFS-04 **Date Analyzed:** 9/7/16 **Analyst Name:** Brian Sheldon

Batch Number: Method Number: EPA 1631E

Project Number(s): Instrument ID: K-AFS-04

Date Analyzed: 9/7/16 Analyst Name: Brian Sheldon

Page 1 of 9 (Peak Report)

Mercury Guru ver ආයුගුම 1995-80 ජලාලිrooks Rand Inc

Batch Number: Method Number: EPA 1631E

Project Number(s): Instrument ID: K-AFS-04 Date Analyzed: 9/7/16 Analyst Name: Brian Sheldon

Page 2 of 9 (Peak Report)

Mercury Guru ver p $\frac{266}{138}$ - 1385 $\frac{27}{38}$ Brooks Rand Inc

Batch Number: Method Number: EPA 1631E

Project Number(s): Instrument ID: K-AFS-04 Date Analyzed: 9/7/16 **Analyst Name:** Brian Sheldon

Page 3 of 9 (Peak Report)

Mercury Guru ver 426ුලි©ි 1399 & P61348Brooks Rand Inc

Batch Number: Method Number: EPA 1631E

Project Number(s): Instrument ID: K-AFS-04 Date Analyzed: 9/7/16 Analyst Name: Brian Sheldon

Page 4 of 9 (Peak Report)

Batch Number: Method Number: EPA 1631E

Project Number(s): Instrument ID: K-AFS-04 Date Analyzed: 9/7/16 Analyst Name: Brian Sheldon

Page 5 of 9 (Peak Report)

Batch Number: Method Number: EPA 1631E

Page 6 of 9 (Peak Report)

Batch Number: Method Number: EPA 1631E

Page 7 of 9 (Peak Report)

Mercury Guru ver ආයුඹු© 1995-30 ff9§rooks Rand Inc

Batch Number: Method Number: EPA 1631E

Project Number(s): Instrument ID: K-AFS-04

Date Analyzed: 9/7/16 **Analyst Name:** Brian Sheldon

Page 8 of 9 (Peak Report)

Mercury Guru ver ආZුලු©14953059Brooks Rand Inc

Batch Number: Method Number: EPA 1631E

Project Number(s): Instrument ID: K-AFS-04 Date Analyzed: 9/7/16 Analyst Name: Brian Sheldon

Preparation Information Benchsheet

19 Total Samples consisting of 13 Client Samples, 2 Client QC Samples, 4 Batch QC Samples associated with the current Prep Run.

Spiking Solutions

* Denotes volume of mixed stock standard.

Conversion from dry weight to wet weight:

Service Request # KILOO9659 Instrument ID# K-ICP-AES-04

ICP-OES 6010 Data Review Form

 $\sqrt{2}$ No

Data Review Form

KQ1610520-01 MB - **Metals T**

MB Evaluation

Evaluation
6010C/Metals T - Ca3933 - Result: 9.19 MRL: 2 S2frrCQ25 7 20X 601 QC/Metals T - Mg2795 - Result: 1.54 MRL: 0.5 **K1609659-003DUP** - **Metals T**

```
ou9659-003DUP - Metais 1<br>DUP RPD<br>6010CMotols T. Es2500, PPD: 22 Limit: 20 CDYTIC Uster IO ClGest / MOD tanco Qenerols
  P RPD<br>6010C/Metals T – Fe2599 – RPD: 23 Limit: 20 part(wlate 1a d1gest/ran-harrager@ov's
```


Page 1 of 1

Sample Name: BLK Acquired: 9/27/2016 8:21:31 Type: Cal Method: 2016B-ICP04(v18) Mode: IR Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: INT. STD. ICP15-100-C

Sample Name: BLK Acquired: 9/27/2016 8:21:31 Type: Cal Method: 2016B-ICP04(v18) Mode: IR Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: INT. STD. ICP15-100-C

Sample Name: STD A Acquired: 9/27/2016 8:24:00 Type: Cal Method: 20168-ICP04(v18) Mode: IR Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: ICP15-68-F

Sample Name: STD B Acquired: 9/27/2016 8:26:16 Type: Cal Method: 2016B-ICP04(v18) Mode: IR Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: ICP15-98-B

#2 1854.2 37213. 3746.0

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Sample Name: KQ1610520-01 Acquired: 9/27/2016 9:19:54 Type: Unk Method: 2016B-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: EM 092716A K1609659-MB

Sample Name: KQ1610520-01 Acquired: 9/27/2016 9:19:54 Type: Unk Method: 2016B-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: EM 092716A K1609659-MB

Sample Name: KQ1610520-01 Acquired: 9/27/2016 9:24:01 Type:Unk Method: 2016B-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: EM 092716A K1609659-MB REPOUR

Sample Name: KQ1610520-01 Acquired: 9/27/2016 9:24:01 Type: Unk Method: 2016B-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: EM 092716A K1609659-MB REPOUR

Sample Name: KQ1610520-02 Acquired: 9/27/2016 9:26:30 **Type:Unk Method: 2016B-ICP04(v18) Mode: CONC** Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: **Comment: EM 092716A K1609659-LCSW**

Sample Name: KQ1610520-02 Acquired: 9/27/2016 9:26:30 Type: Unk Method: 2016B-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: EM 092716A K1609659-LCSW

Sample Name: KQ1610520-03 Acquired: 9/27/2016 9:28:50 Type:Unk Method: 2016B-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: EM 092716A K1609659-SRM1 DORM-4

Sample Name: KQ1610520-03 Acquired: 9/27/2016 9:28:50 Type: Unk Method: 2016B-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: EM 092716A K1609659-SRM1 DORM-4

Sample Name: KQ1610520-04 Acquired: 9/27/2016 9:31:14 Type:Unk Method: 2016B-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: EM 092716A K1609659-SRM2 TORT-3

Sample Name: KQ1610520-04 Acquired: 9/27/2016 9:31:14 Type: Unk Method: 2016B-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: EM 092716A K1609659-SRM2 TORT-3

Sample Name: K1609659-003 Acquired: 9/27/2016 9:34:37 Type:Unk Method: 2016B-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: EM 092716A

#1 1.543 -.0030 61.99 #2 1.545 .0004

Sample Name: K1609659-003 Acquired: 9/27/2016 9:34:37 Type: Unk Method: 20168-ICP04(v18) User: admin Dilution: 1 Comment: EM 092716A Mode: CONC Corr. Factor: 1.000000 Test Type: Sample Type:

Sample Name: K1609659-003L Acquired: 9/27/2016 9:36:59 Type:Unk Method: 20168-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 5 Test Type: Sample Type: Comment: EM 092716A 1/5

Sample Name: K1609659-003D Acquired: 9/27/2016 9:39:25 Type:Unk Method: 2016B-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: EM 092716A

Sample Name: K1609659-003D Acquired: 9/27/2016 9:39:25 Type: Unk Method: 2016B-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: EM 092716A

Sample Name: K1609659-003S Acquired: 9/27/2016 9:41:48 Type: Unk Method: 20168-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: EM 092716A

Sample Name: K1609659-003S Acquired: 9/27/2016 9:41:48 Type: Unk Method: 2016B-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: EM 092716A

Sample Name: K1609659-001 Acquired: 9/27/2016 9:44:08 Type:Unk Method: 2016B-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: EM 092716A

#1 1.550 .0022 59.54 1.552

Sample Name: K1609659-001 Acquired: 9/27/2016 9:44:08 Type: Unk Method: 2016B-ICP04(v18) User: admin Dilution: 1 Comment: EM 092716A Mode: CONC Corr. Factor: 1.000000 Test Type: Sample Type:

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Sample Name: K1609659-002 Acquired: 9/27/2016 9:53:55 Type:Unk Method: 2016B-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: EM 092716A

#1 2.311 .0017 60.60 2.305 -.0045

Sample Name: K1609659-002 Acquired: 9/27/2016 9:53:55 Type: Unk Method: 2016B-ICP04(v18) User: admin Dilution: 1 Comment: EM 092716A Mode: CONC Corr. Factor: 1.000000 Test Type: Sample Type:

Sample Name: K1609659-004 Acquired: 9/27/2016 9:56:19 Type:Unk Method: 2016B-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: EM 092716A

#2 1.820 -.0031 61.07

Sample Name: K1609659-004 Acquired: 9/27/2016 9:56:19 Type: Unk Method: 2016B-ICP04(v18) User: admin Dilution: 1 Comment: EM 092716A Mode: CONC Corr. Factor: 1.000000 Test Type: Sample Type:

Sample Name: K1609659-005 Acquired: 9/27/2016 9:58:42 Type:Unk Method: 2016B-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: EM 092716A

#2 1.288 -.0019 60.43

Sample Name: K1609659-005 Acquired: 9/27/2016 9:58:42 Type: Unk Method: 2016B-ICP04(v18) User: admin Dilution: 1 Comment: EM 092716A Mode: CONC Corr. Factor: 1.000000 Test Type: Sample Type:

Sample Name: K1609659-006 Acquired: 9/27/2016 10:01:04 Type:Unk Method: 2016B-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: EM 092716A

Sample Name: K1609659-006 Method: 2016B-ICP04(v18) User: admin Dilution: 1 Comment: EM 092716A Acquired: 9/27/2016 10:01:04 Type: Unk Mode: CONG Test Type: Corr. Factor: 1.000000 Sample Type:

Sample Name: K1609659-007 Acquired: 9/27/2016 10:03:27 Type:Unk Method: 2016B-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: EM 092716A

#2 1.501 .0008 65.82

Sample Name: K1609659-007 Acquired: 9/27/2016 10:03:27 Type: Unk Method: 2016B-ICP04(v18) User: admin Dilution: 1 Comment: EM 092716A Mode: CONC Corr. Factor: 1.000000 Test Type: Sample Type:

Sample Name: K1609659-008 Acquired: 9/27/2016 10:05:50 Type: Unk Method: 20168-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: EM 092716A

Sample Name: K1609659-008 Acquired: 9/27/2016 10:05:50 Type: Unk Method: 20168-ICP04(v18) User: admin Dilution: 1 Comment: EM 092716A Mode: CONC Corr. Factor: 1.000000 Test Type: Sample Type:

Sample Name: K1609659-009 Acquired: 9/27/2016 10:08:13 Type:Unk Method: 2016B-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: EM 092716A

Sample Name: K1609659-009 Acquired: 9/27/2016 10:08:13 Type: Unk Method: 2016B-ICP04(v18) User: admin Dilution: 1 Comment: EM 092716A Mode: CONC Corr. Factor: 1.000000 Test Type: Sample Type:

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Sample Name: K1609659-010 Acquired: 9/27/2016 10:10:40 Type:Unk Method: 2016B-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: EM 092716A

Sample Name: K1609659-010 Acquired: 9/27/2016 10:10:40 Type: Unk Method: 2016B-ICP04(v18) User: admin Dilution: 1 Comment: EM 092716A Mode: CONC Corr. Factor: 1.000000 Test Type: Sample Type:

Sample Name: K1609659-011 Acquired: 9/27/2016 10:13:00 Type:Unk Method: 2016B-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: EM 092716A

#2 1.540 .0005 60.61

Sample Name: K1609659-011 Acquired: 9/27/2016 10:13:00 Type: Unk Method: 2016B-ICP04(v18) User: admin Dilution: 1 Comment: EM 092716A Mode: CONC Corr. Factor: 1.000000 Test Type: Sample Type:

Sample Name: K1609659-012 Acquired: 9/27/2016 10:15:20 Type:Unk Method: 2016B-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: EM 092716A

Sample Name: K1609659-012 Acquired: 9/27/2016 10:15:20 Type: Unk Method: 2016B-ICP04(v18) User: admin Dilution: 1 Comment: EM 092716A Mode: CONC Corr. Factor: 1.000000 Test Type: Sample Type:

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Sample Name: K1609659-013 Acquired: 9/27/2016 10:24:51 Type:Unk Method: 2016B-ICP04(v18) Mode: CONC Corr. Factor: 1.000000 User: admin Dilution: 1 Test Type: Sample Type: Comment: EM 092716A

#2 3.819 .0007 65.59

Sample Name: K1609659-013 Acquired: 9/27/2016 10:24:51 Type: Unk Method: 2016B-ICP04(v18) User: admin Dilution: 1 Comment: EM 092716A Mode: CONC Corr. Factor: 1.000000 Test Type: Sample Type:

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6020A ICP-MS Data Review Form

Comments: Report Mn by 6010. *MB* = Cu - 0,3ppb, sam plas 10x.

Primary Review By: JP6 Secondary Review By: ______ *rt.._;)-1,,,*

Date: *1/o/n* Date: 1/1/16

Data Review Form

KQ1610520-01 MB - **Metals** T

MB Evaluation

6020A/Metals T - Cu-KED3 - Result: 0.1541205 MRL: 0.1 \int *C*, \int *Ples* 10x. **K1609659-003SDL** - **Metals** T **Serial Dillution** 6020NMetals T - As-KED2 - Recovery: 11 Limit: 90 - *Pos.* **f** *-::p,IH-* **pc,sf~'>. K1609659-00JMS** - **Metals T MS Recovery**

6020NMetalsT-Sb-KED1-Recovery:71 Limits:75-125 *-L..C.51.r'_, S/l-M1* **¹/J.S"~ Sp1***'j..-c...* **,~H-<-~.**

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Dataset Report

User Name: ALKLS.ALKLSXP373 Computer Name: ALKLSXP373 Dataset File Path: C:\NexlONData\DataSet\090716A\ Report Date/Time: Thursday, September 08, 2016 14:41:14

The Dataset

C:\NexIONData\DataSet\090716A\K1608469-C:\NexIONData\DataSet\090716A\K1609288-C:\NexlONData\DataSet\090716A\K1609288-C:\NexlONData\DataSet\090716A\K1609288-C:\NexIONData\DataSet\090716A\K1609288-C:\NexlONData\DataSet\090716A\K1609288-C:\NexlONData\DataSet\090716A\K1609288-C:\NexlONData\DataSet\090716A\K1609288-C:\NexlONData\DataSet\090716A\CCV.059 C:\NexIONData\DataSet\090716A\CCB.060 C:\NexlONData\DataSet\090716A\LLCCVT.06 C:\NexlONData\DataSet\090716A\K1609288-C:\NexIONData\DataSet\090716A\K1609288-C:\NexIONData\DataSet\090716A\K1609288-C:\NexIONData\DataSet\090716A\K1609288-C:\NexIONData\DataSet\090716A\K1609288-C:\NexlONData\DataSet\090716A\K1609288-C:\NexlONData\DataSet\090716A\K1609288-C:\NexlONData\DataSet\090716A\KQ161052 C:\NexlONData\DataSet\090716A\KQ161052 C:\NexlONData\DataSet\090716A\KQ161052\ C:\NexlONData\DataSet\090716A\CCV.072 C:\NexlONData\DataSet\090716A\CCB.073 C:\NexlONData\DataSet\090716A\K1609659-C:\NexlONData\DataSet\090716A\K1609659-C:\NexlONData\DataSet\090716A\K1609659-C:\NexlONData\DataSet\090716A\K1609659-C:\NexlONData\DataSet\090716A\K1609659-C:\NexlONData\DataSet\090716A\K1609659-C:\Nex!ONData\DataSet\090716A\K1609659-C:\NexlONData\DataSet\090716A\KQ161052 C:\NexlONData\DataSet\090716A\MO STD.0 C:\NexIONData\DataSet\090716A\K1609659-C:\NexlONData\DataSet\090716A\CCV.084 C:\NexlONData\DataSet\090716A\CCV.085 C:\NexlONData\DataSet\090716A\CCB.086 C:\NexlONData\DataSet\090716A\LLCCVT.0! C:\NexlONData\DataSet\090716A\K1609659-C:\NexlONData\DataSet\090716A\K1609659-C:\NexlONData\DataSet\090716A\K1609659-C:\NexlONData\DataSet\090716A\K1609659-C:\NexlONData\DataSet\090716A\K1609659-C:\NexlONData\DataSet\090716A\K1609659-C:\NexlONData\DataSet\090716A\K1609659-C:\NexlONData\DataSet\090716A\K1609659-C:\NexlONData\DataSet\090716A\K1609659-C:\NexlONData\DataSet\090716A\KQ161064\ C:\NexlONData\DataSet\090716A\CCV.098 C:\NexlONData\DataSet\090716A\CCV.099 C:\NexlONData\DataSet\090716A\CCB.100 C:\NexIONData\DataSet\090716A\ICSA.101 C:\NexlONData\DataSet\090716A\ICSAB.102 C:\NexlONData\DataSet\090716A\MO STD.1(C:\NexlONData\DataSet\090716A\KQ161064\ C:\NexlONData\DataSet\090716A\KQ161064 C:\NexlONData\DataSet\090716A\KQ161064: C:\NexlONData\DataSet\090716A\K1609289-C:\NexlONData\DataSet\090716A\K1609289-C:\NexlONData\DataSet\090716A\K1609289-C:\NexlONData\DataSet\090716A\K1609289-

C:\NexlONData\DataSet\090716A\CCV.111 C:\NexIONData\DataSet\090716A\CCV.112 C:\NexIONData\DataSet\090716A\CCB.113 C:\NexlONData\DataSet\090716A\LLCCVT.1'
− C:\NexlONData\DataSet\090716A\K1609289-5 3 22:10 C:\NexlONData\DataSet\090716A\K1609289-
1 C:\NexlONData\DataSet\090716A\K1609289-5 C:\NexlONData\DataSet\090716A\K1609289-
5 C:\NexlONData\DataSet\090716A\K1609289-5 C:\NexlONData\DataSet\090716A\K1609289-
5 C:\NexlONData\DataSet\090716A\K1609289-C:\NexlONData\DataSet\090716A\K1609289-
52:03:03:03:03:03:03:03:03:04 C:\NexlONData\DataSet\090716A\K1609289-5
5 C:\NexlONData\DataSet\090716A\K1609289-
5 C:\NexlONData\DataSet\090716A\K1609289-5 C:\NexlONData\DataSet\090716A\K1609289-
5 C:\NexlONData\DataSet\090716A\K1609289-C:\NexlONData\DataSet\090716A\K1609289-5 C:\NexlONData\DataSet\090716A\K1609289-
C:\NexlONData\DataSet\090716A\K1609289-C:\NexlONData\DataSet\090716A\K1609289-5 C:\NexlONData\DataSet\090716A\K1609289-C:\NexlONData\DataSet\090716A\CCV.125 C:\NexlONData\DataSet\090716A\CCB.126
C:\NexlONData\DataSet\090716A\K1609289 C:\NexlONData\DataSet\090716A\K1609289-5 C:\NexlONData\DataSet\090716A\K1609289-
C:\NexlONData\DataSet\090716A\K1609289-5 C:\NexlONData\DataSet\090716A\K1609289-
C:\NexlONData\DataSet\090716A\K1609289-C:\NexlONData\DataSet\090716A\K1609289-5 C:\NexlONData\DataSet\090716A\K1609289-K1609289-016 00: 11 :45 Thu 08-Sep-16 Sample 5 C:\NexlONData\DataSet\090716A \K1609289- C:\NexIONData\DataSet\090716A\K1609289-5 C:\NexlONData\DataSet\090716A\K1609289-K1609289-019 00:26:16 Thu 08-Sep-16 Sample 5 C:\NexlONData\DataSet\090716A \K1609289- C:\NexlONData\DataSet\090716A\K1609289-C:\NexlONData\DataSet\090716A\CCV.137 C:\NexlONData\DataSet\090716A\CCV.138 C:\NexIONData\DataSet\090716A\CCB.139 C:\NexlONData\DataSet\090716A\LLCCVT.14 C:\NexlONData\DataSet\090716A\LLCCVT.14 500 C:\NexlONData\DataSet\090716A\K1608469-500 C:\NexlONData\DataSet\090716A\K1608469-C:\NexlONData\DataSet\090716A\CCV.144 C:\NexIONData\DataSet\090716A\CCV.145 C:\NexlONData\DataSet\090716A\CCB.146 C:\NexlONData\DataSet\090716A\LLCCVT.14 C:\NexlONData\DataSet\090716A\LLCCVT.14

SmartTune Wizard - Details

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Optimization Details 
smartTune file: c:\NexIONData\wizard\smartTune\CAS smartTune Full FAST.swz 
Optimization Status 
Start Time: 9/7/2016 9:21:00 AM 
Mass calibration and Resolution 
        Optimization settings: 
                Method: C:\NexIONData\Method\CAS Tuning.mth. 
                Masscal File: c:\NexIONData\Masscal\Default.tun 
                Iterations: 6 
                Target accuracy(+/- amu): 0.1 for Mass cal. and 0.1 for Resolution 
                Peak height (%) for Res. Opt.: 5
        Optimization Results: 
        Initial Try
                Target/Obtained mass (7.016/7.025), Target/Obtained resolution (0.7/0.699) 
                Target/Obtained mass (9.0122/9.025), Target/Obtained resolution (0.7/0.690) 
                Target/Obtained mass (23.985/24.075), Target/Obtained resolution (0.7/0.696) 
                Target/Obtained mass (58.9332/58.975), Target/Obtained resolution (0.7/0.687) 
                Target/Obtained mass (114.904/114.925), Target/Obtained resolution (0.7/0.689) 
                Target/Obtained mass (139.905/139.925), Target/Obtained resolution (0.7/0.699) 
                Target/Obtained mass (207.977/207.975), Target/Obtained resolution (0.7/0.683) 
                Target/Obtained mass (208.98/208.975), Target/Obtained resolution (0.7/0.705) 
                Target/Obtained mass (238.05/238.025), Target/Obtained resolution (0.7/0.690) 
[Passed] Optimum value(s): N/A
```
Daily Performance Report

Sample ID: Daily Performance Check Sample Date/Time: Wednesday, September 07, 2016 09:25:33 Sample Description: Method File: C:\NexlONData\Method\CAS Daily Performance.mth Dataset File: C:\NexlONData\Dataset\Default\Daily Performance Check.3639 MassCal File: C:\NexlONData\MassCal\Default.tun Conditions File: C:\NexlONData\Conditions\Default.dac Dual Detector Mode: Pulse Acq. Dead Time (ns): 35 Current Dead Time (ns): 35 Torch Z position (mm): 0.00

Summary

Replicates

LABWORKS - Summary Report

Sample ID: Blank Sample Date/Time: Wednesday, September 07, 2016 13:37:34 Sample Description: Autosampler Position: 1 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\Blank.003 User Name: JOB Batch ID:

Sample Unit

ppb ppb

Concentration Results

Sample ID: Standard 1 Sample Date/Time: Wednesday, September 07, 2016 13:42:26 Sample Description: Autosampler Position: 2 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\Standard 1.004 User Name: JOB Batch ID:

Sample ID: ICV Sample Date/Time: Wednesday, September 07, 2016 13:47:16 Sample Description: Autosampler Position: 3 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\ICV.005 User Name: JDB Batch ID:

QC Calculated Values

Sample ID: CCV Sample Date/Time: Wednesday, September 07, 2016 13:52:08 Sample Description: Autosampler Position: 2 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\CCV.006 User Name: JDB Batch ID:

QC Calculated Values

Sample ID: ICB Sample Date/Time: Wednesday, September 07, 2016 13:56:59 Sample Description: Autosampler Position: 1 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\ICB.007 User Name: JOB Batch ID:

Sample ID: CCB Sample Date/Time: Wednesday, September 07, 2016 14:01 :49 Sample Description: Autosampler Position: 1 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\CCB.008 User Name: JDB Batch ID:

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LABWORKS - Summary Report

Sample ID: LLICVT Sample Date/Time: Wednesday, September 07, 2016 14:06:40 Sample Description: Autosampler Position: 4 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\LLICVT.009 User Name: JDB Batch ID:

QC Calculated Values

Sample ID: ICSA Sample Date/Time: Wednesday, September 07, 2016 14:11 :31 Sample Description: Autosampler Position: 5 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\ICSA.010 User Name: JDB Batch ID:

QC Calculated Values

Sample ID: ICSAB Sample Date/Time: Wednesday, September 07, 2016 14:16:21 Sample Description: Autosampler Position: 6 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\ICSAB.011 User Name: JDB Batch ID:

Sample ID: MO STD Sample Date/Time: Wednesday, September 07, 2016 14:21:11 Sample Description: Autosampler Position: 7 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\MO STD.012 User Name: JOB Batch ID:

QC Calculated Values IS Symbol Analyte Mass QC Std % Recovery IS % Recovery Spike % RDuplicate Rel. % Difference $\sqrt{ }$ Li-KED1 6 Be-KED1 9 Al-KED1 27 V-KED3 51 Cr-KED3 52 Cr-KED3 53 Co-KED3 59 Ni-KED3 60 Ni-KED3 62 Cu-KED3 63 Zn-KED3 64 Cu-KED3 65 Zn-KED3 66 L> Ge-KED3 72 Ge-KED2 72 As-KED2 75 Se-KED2 77 Se-KED2 78 Se-KED2 82 Mo-KED2 95 Mo-KED2 97 Mo-KED2 98 Cd-KED2111 Cd-KED2114 L> ln-KED2 115 $\sqrt{ }$ Rh-KED2 103 Ag-KED2 107 | Ag-KED2 109
|-
| In-KED1 115 In-KED1 115 Sb-KED1 121 Sb-KED1 123 Ba-KED1 135 | Ba-KED1 137
|- Lu-KED1 175 Lu-KED1 175 I TI-KED1 203 I TI-KED1 205 Pb-KED1 208 Mn-STD1 55 $\vert \rangle$ Ge-STD 72 **QC Out of Limits** Measurement Type Analyte 94 97 100 98 94 92 97 96 Mass Out of Limits Message Dilution % Difference

Sample ID: KQ1610171-01 Sample Date/Time: Wednesday, September 07, 2016 14:26:03 Sample Description: 5 Autosampler Position: 301 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\KQ1610171-01.013 User Name: JDB Batch ID:

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Sample ID: KQ1610171-03 Sample Date/Time: Wednesday, September 07, 2016 14:30:54 Sample Description: 5 Autosampler Position: 302 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\KQ1610171-03.014 User Name: JDB Batch ID:

Sample ID: KQ1610171-04 Sample Date/Time: Wednesday, September 07, 2016 14:35:44 Sample Description: 5 Autosampler Position: 303 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\KQ1610171-04.015 User Name: JDB Batch ID:

Sample ID: K1608059-001 Sample Date/Time: Wednesday, September 07, 2016 14:40:34 Sample Description: 5 Autosampler Position: 304 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1608059-001.016 User Name: JDB Batch ID:

Sample ID: K1608059-002 Sample Date/Time: Wednesday, September 07, 2016 14:45:25 Sample Description: 5 Autosampler Position: 305 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1608059-002.017 User Name: JDB Batch ID:

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Sample ID: K1608059-003 Sample Date/Time: Wednesday, September 07, 2016 14:50:15 Sample Description: 5 Autosampler Position: 306 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1608059-003.018 User Name: JDB Batch ID:

Sample ID: K1608059-004 Sample Date/Time: Wednesday, September 07, 2016 14:55:06 Sample Description: 5 Autosampler Position: 307 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1608059-004.019 User Name: JDB Batch ID:

Sample ID: CCV Sample Date/Time: Wednesday, September 07, 2016 14:59:59 Sample Description: Autosampler Position: 2 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\CCV.020 User Name: JDB Batch ID:

QC Calculated Values IS Symbol Analyte Mass QC Std % Recovery IS % Recovery Spike % ROuplicate Rel. % Difference I> Li-KED1 6 109 Be-KED1 9 97
Al-KED1 27 100 Al-KED1 27 100
V-KED3 51 102 V-KED3 51 102
Cr-KED3 52 103 Cr-KED3 52 103
Cr-KED3 53 106 Cr-KED3 53 106
Co-KED3 59 104 Co-KED3 59 104
Ni-KED3 60 103 Ni-KED3 60 103
Ni-KED3 62 104 Ni-KED3 62 104
Cu-KED3 63 104 Cu-KED3 63 104

Zn-KED3 64 97 Zn-KED3 64 97
Cu-KED3 65 102 Cu-KED3 65 102

Zn-KED3 66 96 $Zn-KED3$ 66 L> Ge-KED3 72
L> Ge-KED2 72 98 98 I> Ge-KED2 72 102 As-KED2 75 100
Se-KED2 77 97 Se-KED2 77 97

Se-KED2 78 96 Se-KED2 78 96

Se-KED2 82 95 Se-KED2 82 95
Mo-KED2 95 100 00-KED2 95 100
Mo-KED2 97 100 Mo-KED2 97 100
Mo-KED2 98 100 Mo-KED2 98 100
Cd-KED2111 98 Cd-KED2111 98
Cd-KED2114 99 Cd-KED2114 L> ln-KED2 115 101 I> Rh-KED2103 101 Ag-KED2 107 98
Ag-KED2 109 98 L Ag-KED2109 98 In-KED1 115 103
Sb-KED1 121 103 98 Sb-KED1 121 98
I Sb-KED1 123 97 I Sb-KED1123 97 I Ba-KED1135 98 L Ba-KED1 137
|-
| Lu-KED1 175 Lu-KED1 175
TI-KED1 203 100 100 TI-KED1 203 100
TI-KED1 205 101 TI-KED1 205 101
Pb-KED1 208 100 Pb-KED1 208 100
Mn-STD1 55 106 r Mn-STD1 55
- Ge-STD 72 Dilution % Difference

L> Ge-STD 72 100 **QC Out of Limits**

Measurement Type Analyte Mass Out of Limits Message

Sample ID: CCB Sample Date/Time: Wednesday, September 07, 2016 15:04:49 Sample Description: Autosampler Position: 1 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\CCB.021 User Name: JDB Batch ID:

Sample ID: K1608059-005 Sample Date/Time: Wednesday, September 07, 2016 15:09:42 Sample Description: 5 Autosampler Position: 308 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1608059-005.022 User Name: JDS Batch ID:

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Sample ID: K1608059-005D Sample Date/Time: Wednesday, September 07, 2016 15:14:33 Sample Description: 5 Autosampler Position: 309 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A \K1608059-005D.023 User Name: JDB Batch ID:

Sample ID: K1608059-005L Sample Date/Time: Wednesday, September 07, 2016 15:19:23 Sample Description: 25 Autosampler Position: 310 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1608059-005L.024 User Name: JDB Batch ID:

Sample ID: K1608059-005A Sample Date/Time: Wednesday, September 07, 2016 15:24:14 Sample Description: 5 +50ppb +1 0ppb Ag Autosampler Position: 311 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1608059-005A.025 User Name: JDB Batch ID:

Sample ID: K1608059-005S Sample Date/Time: Wednesday, September 07, 2016 15:29:03 Sample Description: 5 Autosampler Position: 312 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1608059-005S.026 User Name: JDB Batch ID:

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Sample ID: KQ1610171-02 Sample Date/Time: Wednesday, September 07, 2016 15:33:53 Sample Description: 5 Autosampler Position: 313 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\KQ1610171-02.027 User Name: JDB Batch ID:

Sample ID: MO STD Sample Date/Time: Wednesday, September 07, 2016 15:38:45 Sample Description: Autosampler Position: 7 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\MO STD.028 User Name: JDB Batch ID:

Sample ID: K1608059-006 Sample Date/Time: Wednesday, September 07, 2016 15:43:37 Sample Description: 5 Autosampler Position: 314 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1608059-006.029 User Name: JOB Batch ID:

Sample ID: K1608059-007 Sample Date/Time: Wednesday, September 07, 2016 15:48:27 Sample Description: 5 Autosampler Position: 315 Number of Replicates: 3 Dataset File: C:\NexIONData\DataSet\090716A\K1608059-007.030 User Name: JDB Batch ID:

Sample ID: K1608059-008 Sample Date/Time: Wednesday, September 07, 2016 15:53:17 Sample Description: 5 Autosampler Position: 316 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1608059-008.031 User Name: JDB Batch ID:

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Sample ID: CCV Sample Date/Time: Wednesday, September 07, 2016 15:58:09 Sample Description: Autosampler Position: 2 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\CCV.032

User Name: JDB

Batch ID:

Sample ID: CCV Sample Date/Time: Wednesday, September 07, 2016 16:02:39 Sample Description: Autosampler Position: 2 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\CCV.033 User Name: JDB Batch ID:

Sample ID: CCB Sample Date/Time: Wednesday, September 07, 2016 16:07:30 Sample Description: Autosampler Position: 1 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\CCB.034 User Name: JDB Batch ID:

Sample ID: LLCCVT Sample Date/Time: Wednesday, September 07, 2016 16:12:21 Sample Description: Autosampler Position: 4 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\LLCCVT.035 User Name: JDB Batch ID:

Concentration Results Mass Meas. lntens. Mean Meas. lntens. RSD Conc. Mean Conc. RSD Sample Unit Analyte | Li-KED1 6 1085076.5 1.6
| Be-KED 157.3 31.8 ppb 0.04679 34.2 ppb I 157.3 31.8 $\frac{1}{2}$ AI-KED1 $\sqrt{27}$ 132191.9 3.4 4.28558 2.1 ppb r V-KED3 5 171.3 8.4 0.42259 10.1 ppb Cr-KED3 52 249.7 9.0

Cr-KED3 53 28.0 32.7 0.43281 8.6 ppb Cr -KED3 53 28.0 32.7
 Co -KED3 59 61.0 12.8 0.42014 35.1 ppb 0.45×12.8

Ni-KED3 60 333.3 9.6 0.03987 17.1 ppb Ni-KED3 60 333.3 9.6
Ni-KED3 62 28.7 35.8 11.2 ppb 0.42361 0.43297 38.1 ppb | Ni-KED3 62 \ 28.7 35.8 Cu-KED3 63 254.7 3.2
2n-KED3 64 184.2 14.0 0.19420 5.1 ppb | Zn-KED3 64 1842 14.0 1.16101 15.7 ppb Cu-KED3 65 5.1
2n-KED3 66 5 102.0 5.1 0.20742 3.6 ppb 0.91929 Zn-KED3 66 102.0 13.7
Ge-KED3 72 27471.7 1.6 16.0 ppb L> Ge-KED3 72 27471.7 1.6 ppb | Ge-KED2 72 144422.8 1.4
| As-KED2 75 256.7 3.9 ppb As-KED2 75 256.7 3.9

Se-KED2 77 12.7 32.9 1.04569 3.2 ppb Se-KED2 77 12.7

Se-KED2 78 91.1 41.6 ppb 1.20356 5.4 Se-KED2 78 91.1
Se-KED2 82 5.1 2.21248 10.7 ppb Se-KED2 82 5.1
Mo-KED2 95 142.7 21.5 1.80486 4.6 ppb Mo-KED2 95 142.7
Mo-KED2 97 108.7 14.2 0.10056 14.0 ppb Mo-KED2 97 108.7
Mo-KED2 98 235.4 14.9 0.11440 16.0 ppb Mo-KED2 98 235.4 16.3
Cd-KED2 111 30.3 6.9 0.09750 15.5 ppb 01-20 111 30.3 6.9
12.0 Cd-KED2 114 71.7 12.0 04204 6.7 ppb 12.0 Cd-KED2 114 71.7 12.0
In-KED2 115 44058.7 1.3 0.04786 11.7 ppb $|$ > In-KED2 ppb I> Rh-KED2 103 241411.0 1.6 ppb 8.4 ppb I Ag-KED2 107 226.7 9.2 0.03985 0.04105 L Ag-KED2 109 225.7 7.2 6.8 ppb \lceil > In-KED1 ppb Pub-Kedia 1287.4 5.4
I Sb-Kedia 123 5.4 5.4 5.4 5.4 0.10550 À7 ppb Pala Sb-KED1 123 1025.5 5.4
Ba-KED1 135 407.3 5.8 0.11102 7.3 pb I Ba-KED1 135 407.3 5.8 0.10928 5.3 dgp L Ba-KED1 137 672.0 10.8 0.10939 10.3 \mathcal{Z} ppb **Property** I> Lu-KED1 175 598356.2 4.7 ppb TI-KED1 203 460.7 4.7
TI-KED1 205 1071.4 21.8 0.04310 10.4 ppb I TI-KED1 205 1071.4 21.8 0.04437 29.1 ppb Pb-KED1 208 1536.7 31.0
Mn-STD1 55 6045.9 1.7 0.05306 36.4 ppb r Mn-STD1 55 6045.9 1.7 0.11794 2.3 ppb L> Ge-STD 72 1695171.5 3.4 ppb

Sample ID: LLCCVT Sample Date/Time: Wednesday, September 07, 2016 16:16:52 Sample Description: Autosampler Position: 4 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\LLCCVT.036 User Name: JDB Batch ID:

Sample ID: KQ1610557-01 Sample Date/Time: Wednesday, September 07, 2016 16:32:26 Sample Description: 5 Autosampler Position: 317 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\KQ1610557-01.037 User Name: JDB Batch ID:

Sample ID: KQ1610557-02 Sample Date/Time: Wednesday, September 07, 2016 16:37:16 Sample Description: 5 Autosampler Position: 318 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\KQ1610557-02.038 User Name: JOB Batch ID:

Sample ID: T1601476-001 Sample Date/Time: Wednesday, September 07, 2016 16:42:06 Sample Description: 5 Autosampler Position: 319 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\T1601476-001.039 User Name: JOB Batch ID:

Sample ID: T1601476-001D Sample Date/Time: Wednesday, September 07, 2016 16:46:56 Sample Description: 5 Autosampler Position: 320 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\T1601476-001 D.040 User Name: JDB Batch ID:

Sample ID: T1601476-001L Sample Date/Time: Wednesday, September 07, 2016 16:51 :47 Sample Description: 25 Autosampler Position: 321 Number of Replicates: 3 Dataset File: C:\NexIONData\DataSet\090716A\T1601476-001L.041 User Name: JOB Batch ID:

Sample ID: T1601476-001A Sample Date/Time: Wednesday, September 07, 2016 16:56:37 Sample Description: 5 +50ppb +1 0ppb Ag Autosampler Position: 322 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\T1601476-001A.042 User Name: JDB Batch ID:

Concentration Results

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QC Calculated Values IS Symbol Analyte Mass QC Std % Recovery IS % Recovery Spike % RDuplicate Rel. % Difference

Li-KED1 6 107 $\sqrt{ }$ Li-KED1 6 Be-KED1 9 AI-KED1 27 **V-KED3 51** Cr-KED3 52 Cr-KED3 53 Co-KED3 59 Ni-KED3 60 Ni-KED3 62 Cu-KED3 63 Zn-KED3 64 Cu-KED3 65 Zn-KED3 66 L> Ge-KED3 72 102 $\sqrt{ }$ Se-KED2 72 103 As-KED2 75 Se-KED2 77 Se-KED2 78 Se-KED2 82 Mo-KED2 95 Mo-KED2 97 Mo-KED2 98 Cd-KED2 111 Cd-KED2114 L> ln-KED2 115 100 $\sqrt{ }$ Rh-KED2 103 100 Ag-KED2 107 Ag-KED2 109 $\sqrt{ }$ $\sqrt{$ I Sb-KED1 121 Sb-KED1 123 **Ba-KED1 135** L Ba-KED1 137 [> Lu-KED1 175 100 I TI-KED1 203 TI-KED1 205 Pb-KED1 208 Mn-STD1 55 L> Ge-STD 72 107 **QC Out of Limits** Measurement Type Analyte Mass Out of Limits Message Dilution % Difference

Sample ID: T1601476-001S Sample Date/Time: Wednesday, September 07, 2016 17:01 :28 Sample Description: 5 Autosampler Position: 323 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\T1601476-001S.043 User Name: JDB Batch ID:

Sample ID: KQ1610485-01 Sample Date/Time: Wednesday, September 07, 2016 17:06:18 Sample Description: 5 Autosampler Position: 324 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\KQ1610485-01.044 User Name: JDB Batch ID:

Sample ID: KQ1610485-02 Sample Date/Time: Wednesday, September 07, 2016 17:11 :08 Sample Description: 5 Autosampler Position: 325 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\KQ1610485-02.045 User Name: JDB Batch ID:

QC Calculated Values

Sample ID: KQ1610485-03 Sample Date/Time: Wednesday, September 07, 2016 17:15:58 Sample Description: 5 Autosampler Position: 326 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\KQ1610485-03.046 User Name: JDB Batch ID:

QC Calculated Values IS Symbol Analyte Mass QC Std % Recovery IS % Recovery Spike % RDuplicate Rel. % Difference $\sqrt{ }$ Li-KED1 6 I Be-KED1 9 | Al-KED1 27
| V-KED3 51 V-KED3 51 Cr-KED3 52 Cr-KED3 53 Co-KED3 59 Ni-KED3 60 Ni-KED3 62 Cu-KED3 63 Zn-KED3 64 Cu-KED3 65 | Zn-KED3 66 L> Ge-KED3 72 Ge-KED2 72 I As-KED2 75 Se-KED2 77 I Se-KED2 78 L Se-KED2 82 r Mo-KED2 95 Mo-KED2 97 I Mo-KED2 98 | Cd-KED2111 Cd-KED2114 L> ln-KED2 115 $\sqrt{ }$ Rh-KED2 103 I Ag-KED2107 2 Ag-KED2 109
Sh-KED1 115 -In-KED1 115 Sb-KED1 121 Sb-KED1 123 | Ba-KED1 135 | Ba-KED1 137
|- Lu-KED1 175 Lu-KED1 175 I TI-KED1 203 I TI-KED1 205 L Pb-KED1208 r Mn-STD1 55
- Ge-STD 72 Ge-STD 72 **QC Out of Limits** 107 95 99 97 96 100 105 102 Dilution % Difference

Measurement Type Analyte

Mass Out of Limits Message

Sample ID: CCV Sample Date/Time: Wednesday, September 07, 2016 17:20:50 Sample Description: Autosampler Position: 2 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\CCV.047 User Name: JDB Batch ID:

QC Calculated Values IS Symbol Analyte Mass QC Std % Recovery IS % Recovery Spike % RUuplicate Rel. % Difference 1 > Li-KED1 6 107 I Be-KED1 9 96 L AI-KED1 27 [V-KED3 51 97 Cr-KED3 52 98 Cr-KED3 53 97 Co-KED3 59 98 Ni-KED3 60 97 Ni-KED3 62 104 Cu-KED3 63 97 Zn-KED3 64 97 Cu-KED3 65 98 Zn-KED3 66 95 L> Ge-KED3 72 98 [> Ge-KED2 72 97 As-KED2 75 101 I Se-KED2 77 87 | Se-KED2 78 99 L Se-KED2 82 92 [Mo-KED2 95 102 Mo-KED2 97 101 | Mo-KED2 98 101 Cd-KED2 111 101 Cd-KED2 114 101 L> ln-KED2 115 96 \lceil > Rh-KED2103 99 I Ag-KED2 107 99 L Ag-KED2 109 98 $\sqrt{ }$ ln-KED1 115 101 Sb-KED1 121 98 Sb-KED1 123 98 Ba-KED1 135 96 L Ba-KED1137 94 [> Lu-KED1 175 103 I TI-KED1 203 100 I TI-KED1 205 99 L Pb-KED1 208 99 [Mn-STD1 55 96 L> Ge-STD 72 103 **QC Out of Limits** Measurement Type QC Std 2 Analyte Se-KED2 Mass 77 Out of Limits Message Out of Control Dilution % Difference

Sample ID: CCB Sample Date/Time: Wednesday, September 07, 2016 17:25:41 Sample Description: Autosampler Position: 1 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\CCB.048 User Name: JDB Batch ID:

Sample ID: KQ1610485-04 Sample Date/Time: Wednesday, September 07, 2016 17:30:33 Sample Description: 5 Autosampler Position: 327 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\KQ1610485-04.049 User Name: JDB Batch ID:

Sample ID: K1608469-001 Sample Date/Time: Wednesday, September 07, 2016 17:35:23 Sample Description: 5 Autosampler Position: 328 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1608469-001.050 User Name: JDB Batch ID:

Sample ID: K1608469-002 Sample Date/Time: Wednesday, September 07, 2016 17:40:14 Sample Description: 5 Autosampler Position: 329 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1608469-002.051 User Name: JDB Batch ID:

Sample ID: K1609288-001 Sample Date/Time: Wednesday, September 07, 2016 17:45:04 Sample Description: 5 Autosampler Position: 330 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609288-001.052 User Name: JDS Batch ID:

Sample ID: K1609288-002 Sample Date/Time: Wednesday, September 07, 2016 17:49:54 Sample Description: 5 Autosampler Position: 331 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609288-002.053 User Name: JDB Batch ID:

Sample ID: K1609288-003 Sample Date/Time: Wednesday, September 07, 2016 17:54:44 Sample Description: 5 Autosampler Position: 332 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609288-003.054 User Name: JDB Batch ID:

Sample ID: K1609288-004 Sample Date/Time: Wednesday, September 07, 2016 17:59:35 Sample Description: 5 Autosampler Position: 333 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609288-004.055 User Name: JDB Batch ID:

QC Calculated Values

Sample ID: K1609288-005 Sample Date/Time: Wednesday, September 07, 2016 18:04:25 Sample Description: 5 Autosampler Position: 334 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609288-005.056 User Name: JDB Batch ID:

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Sample ID: K1609288-005D Sample Date/Time: Wednesday, September 07, 2016 18:09:15 Sample Description: 5 Autosampler Position: 335 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609288-005D.057 User Name: JDB Batch ID:

Sample ID: K1609288-005L Sample Date/Time: Wednesday, September 07, 2016 18:14:05 Sample Description: 25 Autosampler Position: 336 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609288-005L.058 User Name: JDB Batch ID:

Sample ID: CCV Sample Date/Time: Wednesday, September 07, 2016 18:18:57 Sample Description: Autosampler Position: 2 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\CCV.059 User Name: JDB Batch ID:

QC Calculated Values

Sample ID: CCB Sample Date/Time: Wednesday, September 07, 2016 18:23:48 Sample Description: Autosampler Position: 1 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\CCB.060 User Name: JDB Batch ID:

Sample ID: LLCCVT Sample Date/Time: Wednesday, September 07, 2016 18:28:39 Sample Description: Autosampler Position: 4 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\LLCCVT.061 User Name: JDB Batch ID:

Sample ID: K1609288-005A Sample Date/Time: Wednesday, September 07, 2016 18:33:30 Sample Description: 5 +50ppb +10ppb Ag Autosampler Position: 337 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609288-005A.062 User Name: JDB Batch ID:

Sample ID: K1609288-005S Sample Date/Time: Wednesday, September 07, 2016 18:38:20 Sample Description: 5 Autosampler Position: 338 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A \K1609288-005S .063 User Name: JDB Batch ID:

Sample ID: K1609288-006 Sample Date/Time: Wednesday, September 07, 2016 18:43:10 Sample Description: 5 Autosampler Position: 339 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609288-006.064 User Name: JDB Batch ID:

Sample ID: K1609288-007 Sample DatefTime: Wednesday, September 07, 2016 18:48:01 Sample Description: 5 Autosampler Position: 340 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609288-007.065 User Name: JDB Batch ID:

Sample ID: K1609288-008 Sample Date/Time: Wednesday, September 07, 2016 18:52:51 Sample Description: 5 Autosampler Position: 341 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609288-008.066 User Name: JDB Batch ID:

QC Calculated Values IS Symbol Analyte Mass QC Std % Recovery IS % Recovery Spike % ROuplicate Rel. % Difference \lceil $>$ Li-KED1 6 109 **I** Be-KED1 9 L AI-KED1 27 V-KED3 51 Cr-KED3 52 Cr-KED3 53 Co-KED3 59 Ni-KED3 60 Ni-KED3 62 Cu-KED3 63 Zn-KED3 64 Cu-KED3 65 Zn-KED3 66 \vert > Ge-KED3 72 88 \lceil > Ge-KED2 72 92 As-KED2 75 Se-KED2 77 Se-KED2 78 Se-KED2 82 Mo-KED2 95 Mo-KED2 97 Mo-KED2 98 Cd-KED2111 Cd-KED2 114 L> ln-KED2 115 93 $\sqrt{ }$ Rh-KED2103 92 | Ag-KED2 107 L Ag-KED2 109 I> ln-KED1 115 98 Sb-KED1 121 Sb-KED1 123 **Ba-KED1 135** L Ba-KED1 137 $\sqrt{2}$ Lu-KED1 175 101 I TI-KED1 203 I TI-KED1 205 Pb-KED1 208 Mn-STD1 55 $\vert \rangle$ Ge-STD 72 90 **QC Out of Limits** Dilution % Difference

Measurement Type Analyte Mass

Out of Limits Message

Sample ID: K1609288-009 Sample Date/Time: Wednesday, September 07, 2016 18:57:41 Sample Description: 5 Autosampler Position: 342 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609288-009.067 User Name: JDB Batch ID:

Sample ID: K1609288-010 Sample Date/Time: Wednesday, September 07, 2016 19:02:32 Sample Description: 5 Autosampler Position: 343 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609288-010.068 User Name: JDB Batch ID:

Sample ID: KQ1610520-01 Sample Date/Time: Wednesday, September 07, 2016 19:07:22 Sample Description: 5 Autosampler Position: 344 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\KQ1610520-01.069 User Name: JDB Batch ID:

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LABW~RKS - Summ~ort

Sample ID: KQ1610520-03 *C;lJ{V* \ **0** Sample Date/Time: Wednesday, September 07, 2016 19:12:13 Sample Description: 5 Autosampler Position: 345 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\KQ1610520-03.070 User Name: JDB Batch ID:

LABW~ - Summary Report

Sample ID: KQ1610520-04 GALIN L TORT. Sample Date/Time: Wednesday, September 07, 2016 19:17:03 Sample Description: 5 Autosampler Position: 346 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\KQ1610520-04.071 User Name: JDB Batch ID:

Sample ID: CCV Sample Date/Time: Wednesday, September 07, 2016 19:21 :56 Sample Description: Autosampler Position: 2 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\CCV.072 User Name: JDB Batch ID:

QC Calculated Values

Sample ID: CCB Sample Date/Time: Wednesday, September 07, 2016 19:26:47 Sample Description: Autosampler Position: 1 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\CCB.073 User Name: JDB Batch ID:

Sample ID: K1609659-001 Sample Date/Time: Wednesday, September 07, 2016 19:31:40 Sample Description: 5 Autosampler Position: 347 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609659-001.074 User Name: JDB Batch ID:

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LABWORKS - Summary Report

Sample ID: K1609659-002 Sample Date/Time: Wednesday, September 07, 2016 19:36:30 Sample Description: 5 Autosampler Position: 348 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609659-002.075 User Name: JDB Batch ID:

Sample ID: K1609659-003 Sample Date/Time: Wednesday, September 07, 2016 19:41:20 Sample Description: 5 Autosampler Position: 349 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609659-003.076 User Name: JDB Batch ID:

Sample ID: K1609659-003D Sample Date/Time: Wednesday, September 07, 2016 19:46:11 Sample Description: 5 Autosampler Position: 350 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609659-003D.077 User Name: JDB Batch ID:

QC Calculated Values

Sample ID: K1609659-003L Sample Date/Time: Wednesday, September 07, 2016 19:51:01 Sample Description: 25 Autosampler Position: 351 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609659-003L.078 User Name: JDB Batch ID:

QC Calculated Values

Sample ID: K1609659-003A Sample Date/Time: Wednesday, September 07, 2016 19:55:52 Sample Description: 5 +50ppb +10ppb Ag Autosampler Position: 352 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609659-003A.079 User Name: JDB Batch ID:

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Sample ID: K1609659-003S Sample Date/Time: Wednesday, September 07, 2016 20:00:41 Sample Description: 5 Autosampler Position: 353 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609659-003S.080 User Name: JOB Batch ID:

Sample ID: KQ1610520-02 Sample Date/Time: Wednesday, September 07, 2016 20:05:31 Sample Description: 5 Autosampler Position: 354 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\KQ1610520-02.081 User Name: JDB Batch ID:

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LABWORKS - Summary Report

Sample ID: MO STD Sample Date/Time: Wednesday, September 07, 2016 20:10:23 Sample Description: Autosampler Position: 7 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\MO STD.082 User Name: JDB Batch ID:

Sample ID: K1609659-004 Sample Date/Time: Wednesday, September 07, 2016 20:15:15 Sample Description: 5 Autosampler Position: 355 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609659-004.083 User Name: JDB Batch ID:

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Sample ID: CCV Sample Date/Time: Wednesday, September 07, 2016 20:20:07 Sample Description: Autosampler Position: 2 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\CCV.084 User Name: JDB Batch ID:

Concentration Results

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Sample ID: CCV Sample Date/Time: Wednesday, September 07, 2016 20:24:38 Sample Description: Autosampler Position: 2 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\CCV.085 User Name: JDB Batch ID:

QC Calculated Values

Sample ID: CCB Sample Date/Time: Wednesday, September 07, 2016 20:29:29 Sample Description: Autosampler Position: 1 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\CCB.086 User Name: JDB Batch ID:

Sample ID: LLCCVT Sample Date/Time: Wednesday, September 07, 2016 20:34:20 Sample Description: Autosampler Position: 4 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\LLCCVT.087 User Name: JDB Batch ID:

QC Calculated Values

Sample ID: K1609659-005 Sample Date/Time: Wednesday, September 07, 2016 20:39:11 Sample Description: 5 Autosampler Position: 356 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609659-005.088 User Name: JDB Batch ID:

Sample ID: K1609659-006 Sample Date/Time: Wednesday, September 07, 2016 20:44:02 Sample Description: 5 Autosampler Position: 357 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609659-006.089 User Name: JDB Batch ID:

Sample ID: K1609659-007 Sample Date/Time: Wednesday, September 07, 2016 20:48:52 Sample Description: 5 Autosampler Position: 358 Number of Replicates: 3 Dataset File: **C:\NexlON** Data\DataSet\090716A \K1609659-007 .090 User Name: JDB Batch ID:

Sample ID: K1609659-008 Sample Date/Time: Wednesday, September 07, 2016 20:53:42 Sample Description: 5 Autosampler Position: 359 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609659-008.091 User Name: JDB Batch ID:

Sample ID: K1609659-009 Sample Date/Time: Wednesday, September 07, 2016 20:58:33 Sample Description: 5 Autosampler Position: 360 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609659-009.092 User Name: JDB Batch ID:

Sample ID: K1609659-010 Sample Date/Time: Wednesday, September 07, 2016 21 :03:25 Sample Description: 5 Autosampler Position: 101 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A \K1609659-010.093 User Name: JDB Batch ID:

QC Calculated Values IS Symbol Analyte Mass QC Std % Recovery IS % Recovery Spike % R.Ouplicate Rel. % Difference $\begin{bmatrix} > & \text{Li-KED1} & 6 \end{bmatrix}$ 106 Be-KED1 9 AI-KED1 27 V-KED3 51 Cr-KED3 52 Cr-KED3 53 Co-KED3 59 \ Ni-KED3 60 \ Ni-KED3 62 Cu-KED3 63 Zn-KED3 64 Cu-KED3 65 Zn-KED3 66 \vert ₂ Ge-KED3 72 90 \lceil Se-KED2 72 93 As-KED2 75 Se-KED2 77 Se-KED2 78 Se-KED2 82 Mo-KED2 95 Mo-KED2 97 \ Mo-KED2 98 \ Cd-KED2111 \ Cd-KED2114 \vert > ln-KED2 115 93 I> Rh-KED2103 94 \ Ag-KED2107 Ag-KED2 109 $\sqrt{ }$ ln-KED1 115 100 Sb-KED1 121 Sb-KED1 123 **Ba-KED1 135** Ba-KED1 137 [> Lu-KED1 175 102 TI-KED1 203 TI-KED1 205 Pb-KED1 208 Mn-STD1 55 L> Ge-STD 72 95 **QC Out of Limits** Measurement Type **Analyte** Mass Out of Limits Message Dilution % Difference

Sample ID: K1609659-011 Sample Date/Time: Wednesday, September 07, 2016 21 :08:16 Sample Description: 5 Autosampler Position: 102 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609659-011.094 User Name: JDB Batch ID:

QC Calculated Values

Sample ID: K1609659-012 Sample Date/Time: Wednesday, September 07, 2016 21 :13:06 Sample Description: 5 Autosampler Position: 103 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609659-012.095 User Name: JDB Batch ID:

Page 2

Sample ID: K1609659-013 Sample Date/Time: Wednesday, September 07, 2016 21:17:57 Sample Description: 5 Autosampler Position: 104 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609659-013.096 User Name: JDB Batch ID:

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Sample ID: KQ1610643-01 Sample Date/Time: Wednesday, September 07, 2016 21 :22:48 Sample Description: 5 Autosampler Position: 105 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\KQ1610643-01.097 User Name: JDB Batch ID:

Sample ID: CCV Sample Date/Time: Wednesday, September 07, 2016 21:27:40 Sample Description: Autosampler Position: 2 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\CCV.098 User Name: JDB Batch ID:

QC Calculated Values IS Symbol Analyte Mass QC Std % Recovery IS % Recovery Spike % RUuplicate Rel. % Difference $\begin{bmatrix} > & \text{Li-KED1} & 6 \end{bmatrix}$ 104 **IBe-KED1 9 96** Al-KED1 27 103 V-KED3 51 99 Cr-KED3 52 101 Cr-KED3 53 104 Co-KED3 59 103 Ni-KED3 60 103 Ni-KED3 62 103 Cu-KED3 63 103 Zn-KED3 64 99 Cu-KED3 65 103 Zn-KED3 66 98 L> Ge-KED3 72 90 \lceil > Ge-KED2 72 92 As-KED2 75 100 I Se-KED2 77 97 **Se-KED2 78 97** Se-KED2 82 87 Mo-KED2 95 103 Mo-KED2 97 103 Mo-KED2 98 104 Cd-KED2111 102 Cd-KED2114 103 L> ln-KED2 115 91 [> Rh-KED2 103 93 Ag-KED2 107 99 L Ag-KED2 109 99 \lceil > ln-KED1 115 97 Sb-KED1 121 102 Sb-KED1 123 101 Ba-KED1 135 94 L Ba-KED1 137 94 \lceil > Lu-KED1 175 99 TI-KED1 203 99 TI-KED1 205 98 Pb-KED1 208 98 Mn-STD1 55 116 $\lfloor > \hspace{1.5cm} \text{Ge-STD} \hspace{.1cm} \text{72} \hspace{2.5cm} \text{93}$ **QC Out of Limits** Dilution % Difference

Sample ID: CCV Sample Date/Time: Wednesday, September 07, 2016 21:32:10 Sample Description: Autosampler Position: 2 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\CCV.099 User Name: JDB Batch ID: Concentration Results Analyte Meas. lntens. Mean Meas. lntens. RSD Conc. Mean Conc. RSD Sample Unit r> Li-KED1 1218381.5 6.0 ppb Be-KED1 9 85669.7 3.7
Al-KED1 27 858882.2 3.9 23.46917 2.4 ppb $\frac{1}{2}$ AI-KED1 27 3.9 24.94270 **2.2** ppb $\begin{array}{|c|c|c|c|c|}\n\hline\n\end{array}$ V-KED3 51 $\begin{array}{|c|c|c|c|}\n\hline\n\end{array}$ 9690.6 2.7 25.37063 2.3 ppb Cr-KED3 52 13737.6 1.9
Cr-KED3 53 1703.4 3.0 25.32775 2.7 ppb 27.16771 2.6 ppb Cr-KED3 53 \1703.4 3.0
Co-KED3 59 32131.1 2.3 25.98054 2.7 ppb 1.32 Co-KED3 59 2.32 Co-KED3 59 2.3
Ni-KED3 60 1875 9.7 3.6 4.4 ppb Ni-KED3 60 18759.7 3.6
Ni-KED3 62 1619.4 2.4 25.18129 I Ni-KED3 62 2.4 26.00415 3.1 ppb Cu-KED3 63 30123.7 0.6

Zn-KED3 64 3655.8 3.4 25.57273 0.9 ppb 24.54347 4.1 ppb Zn-KED3 64 3655.8 3.4
Cu-KED3 65 15591.5 1.5 | Cu-KED3 65 15591.5 \ 1.5 25.33097 **2.2** ppb I Zn-KED3 66 **2432.2** 2.1 23.89151 2.8 ppb L> Ge-KED3 72 26323.9 0.7 ppb r> Ge-KED2 72 137667.6 1.7 ppb 25.09292 1.3 ppb As-KED2 75 5457.4
Se-KED2 77 189.3 Se-KED2 77 189.3
Se-KED2 78 634.4 6.4 23.29428 5.0 ppb Se-KED2 78 634.4
Se-KED2 82 285.2 እ5 24.03529 1.0 ppb Se-KED2 82 285.2
Mo-KED2 95 34311.2 14.4 22.87822 12.0 ppb Mo-KED2 95 34311.2
Mo-KED2 97 22694.7 0.9 25.62486 1.4 ppb 25.61532 1.1 ppb I Mo-KED2 97 22694.7 1.9 0.4 Mo-KED2 98 58249.0 2.0
1.1 14826.7 1.1 1.1 25.38517 1.2 ppb Cd-KED2 111 14826.7 1.1
Cd-KED2 114 35030.0 1.4 26.20714 1.0 ppb 25.89164 0.6 ppb I Cd-KED2 114 35030.0 1.4 L> ln-KED2 115 42536.6 1.9 ppb ppb $\sqrt{5}$ Rh-KED2 103 237796.3 24.43479 I Ag-KED2 107 131098.3 2.1 1.3 ppb 24.42926 L Ag-KED2 109 126148.9 1.2 0.6 ppb 74726.0 6.3
312750.3 2.9 ppb I Sb-KED1 121 312750.3 2.9 24.73233 3.5 ppb I Sb-KED1 123 239596.9 3.4 24.85480 4.6 ppb 2β I Ba-KED1 135 91469.0 4.7 23.27992 ppb L Ba-KED1 137 151017.6 **4.2** 22.85839 2.2 pb dad r> Lu-KED1 175 635945.0 5.8 3.2 I TI-KED1 203 248291.9 3.9 24.70802 ppb 3.0 568182.1 2.9
744198.9 3.0 24.62778 ppb 24.64476 2.8 ppb Pb-KED1 208 744198.9 3.0
Mn-STD1 55 1043205.9 1.1 27.87234 3.2 ppb 1.1 Mn-STD1 55 1043205.9 1.1
Ge-STD 72 1620737.2 2.2 ppb L> Ge-STD 72 1620737.2 **2.2**

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QC Calculated Values

Sample ID: CCB Sample Date/Time: Wednesday, September 07, 2016 21:37:01 Sample Description: Autosampler Position: 1 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\CCB.100 User Name: JDB Batch ID:

Sample ID: ICSA Sample Date/Time: Wednesday, September 07, 2016 21 :41 :52 Sample Description: Autosampler Position: 5 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\ICSA.101 User Name: JDB Batch ID:

Sample ID: ICSAB Sample Date/Time: Wednesday, September 07, 2016 21:46:42 Sample Description: Autosampler Position: 6 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\ICSAB.102 User Name: JDB Batch ID:

Sample ID: MO STD Sample Date/Time: Wednesday, September 07, 2016 21:51:32 Sample Description: Autosampler Position: 7 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\MO STD.103 User Name: JDB Batch ID:

Concentration Results

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Sample ID: KQ1610643-02 Sample Date/Time: Wednesday, September 07, 2016 21 :56:24 Sample Description: 5 Autosampler Position: 106 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\KQ1610643-02.104 User Name: JDB Batch ID:

IS Symbol Analyte Mass QC Std % Recovery IS % Recovery Spike % RDuplicate Rel. % Difference $\sqrt{ }$ Li-KED1 6 Be-KED1 9 Al-KED1 27 V-KED3 51 Cr-KED3 52 Cr-KED3 53 Co-KED3 59 Ni-KED3 60 Ni-KED3 62 Cu-KED3 63 Zn-KED3 64 Cu-KED3 65 Zn-KED3 66 L> Ge-KED3 72 $\sqrt{ }$ Ge-KED2 72 As-KED2 75 Se-KED2 77 Se-KED2 78 Se-KED2 82 Mo-KED2 95 Mo-KED2 97 Mo-KED2 98 Cd-KED2111 Cd-KED2114 L> ln-KED2 115 $\sqrt{ }$ Rh-KED2103 | Ag-KED2 107
| Ag-KED2 109 | Ag-KED2 109
|- In-KED1 115 In-KED1 115 Sb-KED1 121 Sb-KED1 123 Ba-KED1 135 L Ba-KED1 137
-
- Lu-KED1 175 Lu-KED1 175 I TI-KED1 203 TI-KED1 205 Pb-KED1 208 106 91 92 90 93 103 103 Dilution % Difference

 \vert Ge-STD 72 **QC Out of Limits**

Mn-STD1 55

QC Calculated Values

Measurement Type

Analyte Mass

Out of Limits Message

103

Sample ID: KQ1610643-03 Sample Date/Time: Wednesday, September 07, 2016 22:01 :15 Sample Description: 5 Autosampler Position: 107 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\KQ1610643-03.105 User Name: JDB Batch ID:

QC Calculated Values IS Symbol Analyte Mass QC Std % Recovery IS % Recovery Spike % RUuplicate Rel. % Difference \lceil > Li-KED1 6 99 Be-KED1 9 AI-KED1 27 V-KED3 51 Cr-KED3 52 Cr-KED3 53 Co-KED3 59 Ni-KED3 60 Ni-KED3 62 Cu-KED3 63 Zn-KED3 64 Cu-KED3 65 Zn-KED3 66 L> Ge-KED3 72 86 \lceil > Ge-KED2 72 90 As-KED2 75 Se-KED2 77 Se-KED2 78 Se-KED2 82 Mo-KED2 95 Mo-KED2 97 Mo-KED2 98 Cd-KED2111 Cd-KED2114 L> ln-KED2 115 89 $\sqrt{1}$ Rh-KED2103 89 Ag-KED2 107 L Ag-KED2 109 \lceil > ln-KED1 115 94 Sb-KED1 121 Sb-KED1 123 **Ba-KED1 135** | Ba-KED1 137 \lceil > Lu-KED1 175 97 I TI-KED1 203 I TI-KED1 205 Pb-KED1 208 Mn-STD1 55 \vert Se-STD 72 94 **QC Out of Limits** Measurement Type Analyte Mass Out of Limits Message Dilution % Difference

Sample ID: KQ1610643-04 Sample Date/Time: Wednesday, September 07, 2016 22:06:05 Sample Description: 5 Autosampler Position: 108 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\KQ1610643-04.106 User Name: JDB Batch ID:

QC Calculated Values

Sample ID: K1609289-001 Sample Date/Time: Wednesday, September 07, 2016 22:10:56 Sample Description: 5 Autosampler Position: 109 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609289-001.107 User Name: JDB Batch ID:

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Sample ID: K1609289-001D Sample Date/Time: Wednesday, September 07, 2016 22:15:46 Sample Description: 5 Autosampler Position: 110 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609289-001D.108 User Name: JDB Batch ID:

QC Calculated Values

Sample ID: K1609289-001L Sample Date/Time: Wednesday, September 07, 2016 22:20:36 Sample Description: 25 Autosampler Position: 111 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609289-001L.109 User Name: JDB Batch ID:

Sample ID: K1609289-001A Sample Date/Time: Wednesday, September 07, 2016 22:25:27 Sample Description: 5 +50ppb +10ppb Ag Autosampler Position: 112 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609289-001A.110 User Name: JDB Batch ID:

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Sample ID: CCV Sample Date/Time: Wednesday, September 07, 2016 22:30:18 Sample Description: Autosampler Position: 2 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\CCV.111 User Name: JOB Batch ID:

 $\overline{}$ $V_{\mathcal{B}}% (\varepsilon)$ $\frac{1}{3}$

QC Calculated Values

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Sample ID: CCV le Date!Time: Wednesday, September 07, 2016 22:34:48 Sample Description: Autosambler Position: 2 Number of Replicates: 3 Dataset File: C\\NexIONData\DataSet\090716A\CCV.112 Batch ID:

QC Calculated Values

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Sample ID: CCB Sample Date/Time: Wednesday, September 07, 2016 22:39:39 Sample Description: Autosampler Position: 1 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\CCB.113 User Name: JDB Batch ID:

Sample ID: LLCCVT Sample Date/Time: Wednesday, September 07, 2016 22:44:30 Sample Description: Autosampler Position: 4 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\LLCCVT.114 User Name: JDB Batch ID:

QC Calculated Values IS Symbol Analyte Mass QC Std % Recovery IS % Recovery Spike % RDuplicate Rel. % Difference 1 > Li-KED1 6 108 Be-KED1 9 101
Al-KED1 27 102 AI-KED1 27 102
V-KED3 51 101 V-KED3 51 101
Cr-KED3 52 106 Cr-KED3 52 Cr-KED3 53 115
Co-KED3 59 106 Co-KED3 59 Ni-KED3 60 103 Ni-KED3 62 71 Cu-KED3 63 108 Zn-KED3 64 112
Cu-KED3 65 111 Cu-KED3 65 Zn-KED3 66 102 L> Ge-KED3 72 87 1 > Ge-KED2 72 91 As-KED2 75 I Se-KED2 77 97 I Se-KED2 78 97 | Se-KED2 82 84 1 Mo-KED2 95 98 I Mo-KED2 97 111 Mo-KED2 98 100
Cd-KED2 111 96 Cd-KED2111 Cd-KED2 114 104 L> ln-KED2 115 91 \lceil > Rh-KED2103 94 I Ag-KED2 107 99 \lfloor Ag-KED2 109 1 > ln-KED1 115 99 Sb-KED1 121 97
Sb-KED1 123 98 Sb-KED1 123 98
Ba-KED1 135 97 I Ba-KED1 135 97 \lfloor Ba-KED1 137 \lceil > Lu-KED1 175 98 I TI-KED1 203 91 TI-KED1 205 99 L Pb-KED1 208 107 Mn-STD1 55 129 $\vert \rangle$ Ge-STD 72 92 Dilution % Difference

QC Out of Limits

Measurement Type Analyte Mass

Out of Limits Message

Sample ID: K1609289-001S Sample Date/Time: Wednesday, September 07, 2016 22:49:21 Sample Description: 5 Autosampler Position: 113 Number of Replicates: 3 Dataset File: C:\NexIONData\DataSet\090716A\K1609289-001S.115 User Name: JOB Batch ID:

Sample ID: K1609289-002 Sample Date/Time: Wednesday, September 07, 2016 22:54:12 Sample Description: 5 Autosampler Position: 114 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609289-002.116 User Name: JDB Batch ID:

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Sample ID: K1609289-003 Sample Date/Time: Wednesday, September 07, 2016 22:59:02 Sample Description: 5 Autosampler Position: 115 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609289-003.117 User Name: JDB Batch ID:

Sample ID: K1609289-004 Sample Date/Time: Wednesday, September 07, 2016 23:03:52 Sample Description: 5 Autosampler Position: 116 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609289-004.118 User Name: JDB Batch ID:

Sample ID: K1609289-005 Sample Date/Time: Wednesday, September 07, 2016 23:08:43 Sample Description: 5 Autosampler Position: 117 Number of Replicates: 3 Dataset File: **C:\NexlON** Data\DataSet\090716A \K 1609289-005.119 User Name: JDB Batch ID:

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Sample ID: K1609289-006 Sample Date/Time: Wednesday, September 07, 2016 23:13:33 Sample Description: 5 Autosampler Position: 118 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609289-006.120 User Name: JDB Batch ID:

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Sample ID: K1609289-007 Sample Date/Time: Wednesday, September 07, 2016 23:18:24 Sample Description: 5 Autosampler Position: 119 Number of Replicates: 3 Dataset File: **C:\NexlON** Data\DataSet\090716A \K1609289-007 **.121** User Name: JDB Batch ID:

Sample ID: K1609289-008 Sample Date/Time: Wednesday, September 07, 2016 23:23:15 Sample Description: 5 Autosampler Position: 120 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609289-008.122 User Name: JDB Batch ID:

Sample ID: K1609289-009 Sample Date/Time: Wednesday, September 07, 2016 23:28:05 Sample Description: 5 Autosampler Position: 121 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609289-009.123 User Name: JDB Batch ID:

Sample ID: K1609289-010 Sample Date/Time: Wednesday, September 07, 2016 23:32:55 Sample Description: 5 Autosampler Position: 122 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609289-010.124 User Name: JDB Batch ID:

Sample ID: CCV Sample Date/Time: Wednesday, September 07, 2016 23:37:49 Sample Description: Autosampler Position: 2 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\CCV.125 User Name: JDB Batch ID:

QC Calculated Values

Sample ID: CCB Sample Date/Time: Wednesday, September 07, 2016 23:42:39 Sample Description: Autosampler Position: 1 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\CCB.126 User Name: JOB Batch ID:

Sample ID: K1609289-011 Sample Date/Time: Wednesday, September 07, 2016 23:47:32 Sample Description: 5 Autosampler Position: 123 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609289-011.127 User Name: JDB Batch ID:

Sample ID: K1609289-012 Sample Date/Time: Wednesday, September 07, 2016 23:52:23 Sample Description: 5 Autosampler Position: 124 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609289-012.128 User Name: JOB Batch ID:

Sample ID: K1609289-013 Sample Date/Time: Wednesday, September 07, 2016 23:57:13 Sample Description: 5 Autosampler Position: 125 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609289-013.129 User Name: JDB Batch ID:

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Sample ID: K1609289-014 Sample Date/Time: Thursday, September 08, 2016 00:02:04 Sample Description: 5 Autosampler Position: 126 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609289-014.130 User Name: JOB Batch ID:

Sample ID: K1609289-015 Sample Date/Time: Thursday, September 08, 2016 00:06:55 Sample Description: 5 Autosampler Position: 127 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609289-015.131 User Name: JOB Batch ID:

Sample ID: K1609289-016 Sample Date/Time: Thursday, September 08, 2016 00:11 :45 Sample Description: 5 Autosampler Position: 128 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609289-016.132 User Name: JOB Batch ID:

Sample ID: K1609289-017 Sample Date/Time: Thursday, September 08, 2016 00:16:35 Sample Description: 5 Autosampler Position: 129 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609289-017.133 User Name: JOB Batch ID:

Sample ID: K1609289-018 Sample Date/Time: Thursday, September 08, 2016 00:21 :26 Sample Description: 5 Autosampler Position: 130 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609289-018.134 User Name: JDB Batch ID:

Sample ID: K1609289-019 Sample Date/Time: Thursday, September 08, 2016 00:26:16 Sample Description: 5 Autosampler Position: 131 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609289-019.135 User Name: JOB Batch ID:

Sample ID: K1609289-020 Sample Date/Time: Thursday, September 08, 2016 00:31:07 Sample Description: 5 Autosampler Position: 132 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\K1609289-020.136 User Name: JDB Batch ID:

Sample ID: CCV Sample Date/Time: Thursday, September 08, 2016 00:35:59 Sample Description: Autosampler Position: 2 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\CCV.137 User Name: JDB Batch ID: Concentration Results Analyte Mass Meas. Intens. Mean Meas. Intens. RSD Conc. Mean Conc. RSD Sample Unit $\sqrt{5}$ Li-KED1 6 \ 1194728.9 3.5 ppb Be-KED1 9 85002.2 0.9
Al-KED1 27 834479.1 1.0 23.74239 3.2 ppb $\frac{1}{2}$ AI-KED1 27 \834479.1 1.0 24.70725 2.5 ppb $\begin{bmatrix} 1.4 & 1.4 \\ 1.4 & 1.4 \\ 1.4 & 1.4 \end{bmatrix}$ 25.59690 0.6 ppb 12836.1 0.4 0.4 0.4 0.4 0.4 0.4 cr-KED3 53 0.4 0.54 0.4 26.27209 2.5 ppb 1542.1 5.6 53 53 53 53
Co-KED3 59 29684.X 5.9 27.32437 7.4 ppb Co-KED3 59 29684. 1.9
Ni-KED3 60 17811.8 1.8 26.63998 1.6 ppb Ni-KED3 60
Ni-KED3 62 26.54395 3.7 ppb Ni-KED3 62 1541.4 1.7
Cu-KED3 63 28178.4 2.3 27.48597 4.3 ppb Cu-KED3 63 28178.4 2.3
2n-KED3 64 3416.5 0.9 26.55711 3.1 ppb Zn-KED3 64 3416.5 0.9
Cu-KED3 65 14551.4 0.8 25.46204 2.5 ppb Cu-KED3 65 14551.4 0.8
2n-KED3 66 2304.5 1.6 26.24881 3.3 ppb Zn-KED3 66
Ge-KED3 72 25.13600 3.6 ppb 2.6 L> Ge-KED3 72 23720.0 ppb $\sqrt{5}$ Ge-KED2 አወ ppb As-KED2 75 5200.9
Se-KED2 77 196.7 $2.\overline{7}$ 25.12303 4.6 ppb Se-KED2 77 196.7
Se-KED2 78 615.0 9.4 25.43843 9.2 ppb Se-KED2 78 615.0 5.7 24.46866 3.9 ppb L Se-KED2 82 271.9 12.2 22.96420 13.4 ppb r Mo-KED2 95 32468.9 2.2 25.67592 0.8 ppb Mo-KED2 97 21515.5 1.7
Mo-KED2 98 55957.5 0.9 25.6 778 0.3 ppb Mo-KED2 98 55957.5 0.9
Cd-KED2 111 14126.3 0.6 25.72868 1.2 ppb Cd-KED2 111 14126.3 0.6
Cd-KED2 114 33379.2 1.2 25.33692 1.4 ppb 0d-KED2 114 33379.2 1.2
In-KED2 115 40321.5 1.8 25.42387 0.9 ppb \vert > In-KED2 ppb I> Rh-KED2 103 226050.0 2.3 ppb Ag-KED2 107 124597.7 1.0 24.43627 ppb L Ag-KED2 109 119834.9 1.4 24.41558 1.0 ppb In-KED1 pop P Sb-KED1 121 304550.5 1.5
Sb-KED1 123 231673.2 1.8 24.76991 1.8 ppb Sb-KED1 24.71366 2.4 ppb Ba-KED1 135 86881.6 3.5 22.74960 1.2 ppb L Ba-KED1 137 143391.5 1.6 22.33372 1.7 ppb I> Lu-KED1 175 614983.5 1.5 ppb 5 I TI-KED1 203 237968.6 1.5 24.46521 1.0 ppb $\frac{1}{a}$ 205 549728.4 1.3
208 722094.0 1.0 24.61636 2.2 ppb Pb-KED1 208 722094.0 1.0
Mn-STD1 55 990863.2 7.8 24.70339 1.3 ppb r Mn-STD1 55 990863.2 7.8 29.87609 5.3 ppb L> Ge-STD 72 1434780.8 3.9 ppb

QC Calculated Values

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Sample ID: CCV Sample Date/Time: Thursday, September 08, 2016 00:40:29 Sample Description: Autosampler Position: 2 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\CCV.138 User Name: JOB Batch ID:

QC Calculated Values

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LABWORKS - Summary Report

Sample ID: CCB Sample Date/Time: Thursday, September 08, 2016 00:45:20 Sample Description: Autosampler Position: 1 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\CCB.139 User Name: JDB Batch ID:

Sample ID: LLCCVT Sample Date/Time: Thursday, September 08, 2016 00:50:11 Sample Description: Autosampler Position: 4 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\LLCCVT.140 User Name: JDB Batch ID:

QC Calculated Values

Sample ID: LLCCVT Sample Date/Time: Thursday, September 08, 2016 00:54:41 Sample Description: Autosampler Position: 4 Number of Replicates: 3 Dataset File: C:\NexlONData\DataSet\090716A\LLCCVT.141 User Name: JDB Batch ID:

QC Calculated Values

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Fish Tissue Metals Concentration Laboratory Report

Jupiter Environmental Laboratories, Inc. www.jupiterlabs.com clientservices@jupiterlabs.com 150 S. Old Dixie Highway Jupiter, FL 33458 Phone: (561)575-0030 Fax: (561)575-4118

October 2, 2015

Caleb Zurstadt USDA Forest Service Region 4 Utah Acquis Support Center 2222 West 2300 South Salt Lake City, UT 84119

RE: LOG# 1542993 Project ID: Whole Body Fish Tissue COC# 542993

Dear Caleb Zurstadt:

Enclosed are the analytical results for sample(s) received by the laboratory on Tuesday, August 18, 2015. Results reported herein conform to the most current NELAC standards, where applicable, unless indicated by * in the body of the report. The enclosed Chain of Custody is a component of this package and should be retained with the package and incorporated therein.

Results for all solid matrices are reported in dry weight unless otherwise noted. Results for all liquid matrices are reported as received in the laboratory unless otherwise noted. Results relate only to the samples received. Should insufficient sample be provided to the laboratory to meet the method and NELAC Matrix Duplicate and Matrix Spike requirements, then the data will be analyzed, evaluated and reported using all other available quality control measures.

Samples are disposed of after 30 days of their receipt by the laboratory unless extended storage is requested in writing. The laboratory maintains the right to charge storage fees for archived samples. This report will be archived for 5 years after which time it will be destroyed without further notice, unless prior arrangements have been made.

Certain analyses are subcontracted to outside NELAC certified laboratories, please see the Project Summary section of this report for NELAC certification numbers of laboratories used. A Statement of Qualifiers is available upon request.

If you have any questions concerning this report, please feel free to contact me.

Sincerely,

as Trill

Melissa Mills for Kacia Baldwin V.P. of Operations

Report ID: 1542993 - 1496225 Page 1 of 47 10/2/2015

CERTIFICATE OF ANALYSIS FDOH# E86546

SAMPLE ANALYTE COUNT

Workorder: 1542993 Project ID: Whole Body Fish Tissue

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CERTIFICATE OF ANALYSIS FDOH# E86546

SAMPLE ANALYTE COUNT

Workorder: 1542993 Project ID: Whole Body Fish Tissue

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CERTIFICATE OF ANALYSIS FDOH# E86546

SAMPLE SUMMARY

Workorder: 1542993 Project ID: Whole Body Fish Tissue

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CERTIFICATE OF ANALYSIS FDOH# E86546

ANALYTICAL RESULTS

Workorder: 1542993

Project ID: Whole Body Fish Tissue

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ANALYTICAL RESULTS

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ANALYTICAL RESULTS

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ANALYTICAL RESULTS

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ANALYTICAL RESULTS

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ANALYTICAL RESULTS

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ANALYTICAL RESULTS QUALIFIERS

Workorder: 1542993 Project ID: Whole Body Fish Tissue

PARAMETER QUALIFIERS

J4 MS/MSD recovery exceeded control limits due to matrix interference. LCS/LCSD recovery was within acceptable range.

PROJECT COMMENTS

1542993 A reported value of U indicates that the compound was analyzed for but not detected above the MDL. A value flagged with an "i" flag indicates that the reported value is between the laboratory method detection limit and the practical quantitation limit.

All samples are reported "as received" as wet weight.

Report ID: 1542993 - 1496225 Page 35 of 47 10/2/2015

CERTIFICATE OF ANALYSIS FDOH# E86546

QUALITY CONTROL DATA

Workorder: 1542993

Project ID: Whole Body Fish Tissue

METHOD BLANK: 84770

LABORATORY CONTROL SAMPLE & LCSD:

84772

Report ID: 1542993 - 1496225 Page 36 of 47 10/2/2015

CERTIFICATE OF ANALYSIS FDOH# E86546

QUALITY CONTROL DATA

Workorder: 1542993

Project ID: Whole Body Fish Tissue

MATRIX SPIKE SAMPLE: 84774

Original: 1542993024

SAMPLE DUPLICATE: 84773

Original: 1542993024

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CERTIFICATE OF ANALYSIS FDOH# E86546

Workorder: 1542993

Project ID: Whole Body Fish Tissue

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CERTIFICATE OF ANALYSIS FDOH# E86546

Workorder: 1542993

Project ID: Whole Body Fish Tissue

LABORATORY CONTROL SAMPLE & LCSD: 84779 84780

Iron mg/Kg U 4.4

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CERTIFICATE OF ANALYSIS FDOH# E86546

Workorder: 1542993

Project ID: Whole Body Fish Tissue

MATRIX SPIKE SAMPLE: 84782

Original: 1542993030

SAMPLE DUPLICATE: 84781

Original: 1542993030

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CERTIFICATE OF ANALYSIS FDOH# E86546

Workorder: 1542993

Project ID: Whole Body Fish Tissue

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CERTIFICATE OF ANALYSIS FDOH# E86546

Workorder: 1542993

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Workorder: 1542993

Project ID: Whole Body Fish Tissue

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QUALITY CONTROL DATA QUALIFIERS

Workorder: 1542993 Project ID: Whole Body Fish Tissue

QUALITY CONTROL PARAMETER QUALIFIERS

- J4 MS/MSD recovery exceeded control limits due to matrix interference. LCS/LCSD recovery was within acceptable range.
- P1 RPD value not applicable for sample concentrations less than 5 times the PQL.

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CERTIFICATE OF ANALYSIS FDOH# E86546

QUALITY CONTROL DATA CROSS REFERENCE TABLE

Workorder: 1542993 Project ID: Whole Body Fish Tissue

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> **CERTIFICATE OF ANALYSIS FDOH# E86546**

QUALITY CONTROL DATA CROSS REFERENCE TABLE

Workorder: 1542993 Project ID: Whole Body Fish Tissue

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CERTIFICATE OF ANALYSIS FDOH# E86546

QUALITY CONTROL DATA CROSS REFERENCE TABLE

Workorder: 1542993 Project ID: Whole Body Fish Tissue

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CERTIFICATE OF ANALYSIS FDOH# E86546

Jupiter Environminial Librarian II ...

Sampling Site Address

Attn: Caleb Zurs +
Project What Buty Fish

QA/QC level with report

SFWMD Q ADaPT Q

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 \mathcal{Q}

Page_t of 3

3112993

www.jupiterlabs.com
150 S. Old Dixie Highway, Jupiter, FL 33458
575-0030 · (888) 287-3218 · clientservices@jupiterlabs.com

J.E.L. Log # 1542993

 $P.0.$ #

Quote $#$

Floring -AM Laugraturies: Inc.

Page 2 of 3

Page 3 of 3

 542993

Login Checklist

Cooler Unpacked/Checked by: ___________________Date: 8/18/15

JEL LOG#: I *-S-l(.2_\'1 ¹*

Cooler Check

Note: if the temperature of a cooler is above 6C or an evidence seal is damaged then identify the bottles in the affected cooler(s) on the sample discrepancy form . *Write tracking number only if waybill copy cannot be placed in the folder

Condition of Containers:

ORDER FOR SUPPLIES OR SERVICES SCHEDULE - CONTINUATION

PAGE NO

 \overline{a} I.

isi Prescribed by GSA FAR (40 CFR) 5)

Sara Ouly

L.. 'l

£ s

The table below has the analysis that the contract bid was based on (i.e, the analysis we need run). And according the Kacia this is the method you use for trace metal analysis:

A unique 3rd generation collision/reaction cell is utilized in all 7700 Series instruments to remove spectral interferences that might otherwise bias results. Helium (He) mode is already established as the *only* reliable cell method for complex and variable samples, because it filters out *all* polyatomics, even unidentified ones.

received these 8.18 **@** around 930 an n UPS

Hope this helps gets us moving forward. Sorry about this, I'm pretty tapped with dealing with forest fire priorities.

From: Sara Ouly [mailto:souly@jupiterlabs.com] **Sent:** Wednesday, August 19, 2015 6:58 AM **To:** Zurstadt, Caleb F -FS **Cc:** kbaldwin@jupiterlabs.com **Subject:** RE: Award for Whole Body Fish Tissue Aralysis (AG-84N8-D-15-0093)

We can't run the samples without the chain of custody. This is because it gives us the information that we currently do not possess, such as what tests you wish to run, what the order and the name of the samples are, and when they were collected. Without this written approval by the client to proceed in the manner specified, we can only hold the samples. I can fill out the COC for you, but you must tell me in writing all of the information needed so that I have documentation of consent and then I can send you a copy so that you may make sure I have everything correct.

1

ORDER FOR SUPPLIES OR SERVICES

PAGE NO

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https://www.gsaadvantage.gov/advantage/buyer/buyer_review_quote_print.do

7/23/2015

The Krassel Ranger District, located in McCall, Idaho has a need for Lab Testing on fish samples. All testing must meet EPA Standardized Guidelines. The testing includes:

Assuming an award is made prior to July 31, 2015; all analysis requirements are due no later than September 30, 2015.

Offerors shall provide descriptive information on Analysis Methods and Report:

EPA 1631, Appendix - Total Mercury in Tissue

EPA 1638, Modified w/DRC - Trace Metals in Tissue

Description of details provided in report on analysis results

The government intends to evaluate all offers to determine overall Best Value to the Government. The government will review technical components of the proposed sampling methods, delivery and price. Technical components and delivery, when combined are of greater value when compared to price.

Appendix 5: Genetic Studies Results

Appendix 5: Genetic Studies Results

Stibnite Gold Project Midas Gold Idaho, Inc.

April 2017

Prepared by

MWH Americas, Inc. 727 E. Riverpark Lane, Suite 150 Boise, Idaho 83706

2015 Fish Tissue Genetics Study Human: ALProject: ID bulltrout ID & genotype disagree

Sfo alleles

Missing data

2016 eDNA Results

Appendix 6: Summary of Fish Counts and Basic Habitat Information from Snorkel Surveys

Appendix 6: Summary of Fish Counts and Basic Habitat Information from Snorkel Surveys

Stibnite Gold Project Midas Gold Idaho, Inc.

April 2017

Prepared by

MWH Americas, Inc. 727 E. Riverpark Lane, Suite 150 Boise, Idaho 83706

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ARTICLE

Effects of Artificial Lighting at Night on Predator Density and Salmonid Predation

Thomas Reid Nelson,[*](https://orcid.org/0000-0002-7960-2084) Cyril J. Michel, Meagan P. Gary, Brendan M. Lehman, and Nicholas J. Demetras

University of California Santa Cruz, Institute of Marine Sciences, affiliated with the National Oceanic and Atmospheric Administration, National Marine Fisheries Service, Southwest Fisheries Science Center, 110 McAllister Way, Santa Cruz, California 95060, USA

Jeremy J. Hammen and Michael J. Horn

U.S. Bureau of Reclamation, Fisheries and Wildlife Resources Group, USBR 86‐68290, Denver, Colorado 80225, USA

Abstract

Predation of juvenile salmonids within California's Sacramento–San Joaquin Delta (the Delta) has been identified as a contributing factor to low survival during out-migration through the system. Artificial lighting at night (ALAN) may contribute to increased levels of salmonid predation by attracting predators and prey, increasing predator reaction distance, and boosting foraging success. To assess ALAN effects on predator (piscivorous fishes) density and the relative predation risk of Chinook Salmon Oncorhynchus tshawytscha smolts in the Delta, we preformed field-based experiments with introduced ALAN. We used adaptive resolution imaging sonar cameras to generate predator density estimates in light and dark treatments throughout nightly experiments at 30-min intervals. We simultaneously deployed predation event recorders to estimate the impact of ALAN intensity (lux) on relative predation risk of Chinook Salmon smolts. Early in the night (1–3 h past sunset), predator density and relative predation risk of smolts were unrelated to ALAN. However, late in the night (3–5 h past sunset), ALAN presence increased predator density, and the relative predation risk of juvenile salmonids increased with increasing lux. Predation risk was also positively related to predator density, and increased late-night predator density under ALAN, coupled with late-night foraging benefits of ALAN, likely contributed to the lux–risk relationship. The exact mechanism behind this discrepancy between early- and late-night trends is unknown and could be a result of our experimental design or the predator community sampled here. However, if these temporal trends prove robust to future investigations, late-night lighting reduction campaigns during out-migration could maximize the human benefits of ALAN while minimizing the negative impacts on salmonids. Overall, our findings align with others and suggest that ALAN increases juvenile salmonid predation. Although many questions remain unanswered, it appears that reducing artificial illumination is a practical management strategy to reduce predation.

The diel light cycle is a driving force behind animal behavior and ecological interactions (Navara and Nelson 2007; Hölker et al. 2010a, 2010b). However, abundant and increasing artificial lighting at night (ALAN) disrupts ecological processes across a wide range of taxa (Gaston et al. 2013, 2014a; Zapata et al. 2019). Artificial lighting at night affects animal migration and orientation, foraging and predation, reproduction, and even human health (Navara and Nelson 2007). Although ALAN impacts in terrestrial ecosystems are apparent, ALAN also affects

^{*}Corresponding author: [thomas.nelson@noaa.gov](mailto:)

Corrections added on 27 February 2021, after first online publication

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aquatic environments, given that many urban areas are near coastlines, estuaries, and freshwater shorelines (Davies et al. 2014; Jechow and Hölker 2019; Zapata et al. 2019). The presence of ALAN in aquatic environments attracts prey species and piscivorous fishes (Becker et al. 2011; Lehman et al. 2019), increases the foraging efficiency and predatory behavior of fishes (Bolton et al. 2017), and increases the reaction distance and prey consumption rate of fishes (Vogel and Beauchamp 1999; Mazur and Beauchamp 2003, 2006). Therefore, ALAN in the aquatic environment may increase piscivorous fish predation rates and negatively impact prey survival.

Juvenile salmonids face a gauntlet of potential predators as they migrate from freshwater to marine environments (Poe et al. 1991; Rieman et al. 1991; Osterback et al. 2013). Predation risk is exacerbated in clear waters (Gregory and Levings 1998) and out-migrating salmonids employ nocturnal migrations as one strategy to minimize this risk (Chapman et al. 2013; Clark et al. 2016; Furey et al. 2016). However, in many ecosystems, salmonid out-migration traverses anthropogenically altered habitat and urban centers (Michel et al. 2013, 2015; Schroeder et al. 2015), likely leading to increased ALAN exposure (Jechow and Hölker 2019; Zapata et al. 2019). Given that ALAN aggregates and slows out-migrating salmon, attracts predators (including fishes, birds, and mammals), and increases piscivore consumption of salmonids (Yurk and Trites 2000; Tabor et al. 2004; Celedonia et al. 2011; Tabor et al. 2017), increased migratory ALAN exposure likely increases juvenile salmonid predation risk and mortality.

The Sacramento–San Joaquin River Delta (the Delta) is a heavily modified tidal freshwater system consisting of large, interconnected waterways that drain the Central Valley of California (Monsen et al. 2007; Lehman et al. 2019). The Delta provides water for irrigation and municipalities across a large portion of the state and habitat for threatened and endangered fish species (Mount and Twiss 2005; Williams 2006). Four runs of Chinook Salmon *Oncorhynchus tshawytscha* (winter, spring, fall, and late fall) and steelhead *O. mykiss* must pass through the Delta during both juvenile and spawning migrations (Williams 2006). Populations of these salmonids have drastically declined over the past century, and poor juvenile survival is a contributing factor (Yoshiyama et al. 1998; Williams 2006; Lindley et al. 2009). Current survival estimates of out-migrating juvenile Chinook Salmon, or smolts, through the Delta are as low as 5% (Buchanan et al. 2013, 2018) and are likely affected by predation from nonnative fishes (Grossman 2016; Buchanan et al. 2018; Michel et al. 2018).

Management strategies that minimize out-migrant mortality are needed to rebuild salmonid populations. Decreasing the interactions of juvenile salmonids and their predators may be a way to lower predation-induced mortality. Removing anthropogenic contact and aggregation

points between predators and prey is likely one way to decrease these interactions (Lehman et al. 2019). Major metropolitan areas located throughout the Delta (e.g., Sacramento, Stockton, and Antioch, California), and elsewhere along rivers of the California Central Valley, likely produce an abundance of ALAN, and ALAN has been identified as a prevalent predator–prey contact point (Lehman et al. 2019). Therefore, we investigated whether ALAN affected piscivorous fish (predator) density and the relative predation risk of Chinook Salmon smolts in the Delta. This study was an important first step in determining if ALAN removal or intensity reduction may be a viable management strategy to help lower predation mortality in the Delta and other waterways.

METHODS

Study system.— Prior to ocean entry, all out-migrating Central Valley Chinook Salmon pass through the heavily modified Delta and San Francisco Estuary (Nichols et al. 1986), typically in the spring (Williams 2006). During outmigration, nocturnal movements are preferred in all river reaches except the estuary (Chapman et al. 2013), and migration speed is slowest in the Delta (Michel et al. 2013). Salmon released in the Sacramento River take, on average, 12.4 d to reach the Delta and 13.2 d to migrate from the Delta to the ocean (Michel et al. 2013). Most Delta land is below water level, waterways are channelized and leveed, and riprap covers 73% of mainstream shorelines (Mount and Twiss 2005; Lehman et al. 2019). Invasive fishes and vegetation are prevalent in the Delta (Underwood et al. 2006; Brown et al. 2007) and may contribute to poor survival of out-migrating salmonids (Grossman 2016; Michel et al. 2018; Zeug et al., in press).

Delta piscivores consist of fishes, birds, and mammals. Piscivorous fishes that may consume juvenile salmonids are dominated by invasive species, including Striped Bass *Morone saxatilis*, Largemouth Bass *Micropterus salmoides*, other black basses *Micropterus* spp., sunfish (*Lepomis* spp. and *Pomoxis* spp.), and catfish (Ictaluridae) (Grossman 2016; Michel et al. 2018). The Sacramento Pikeminnow *Ptychocheilus grandis* is the only significant native piscivorous fish in this system (Brown and Moyle 1981). Avian and mammalian piscivory within the Delta is poorly studied (Grossman 2016). However, avian salmonid predation occurs in San Francisco Bay (Evans et al. 2011; Adrean et al. 2012; Riensche et al. 2012), and piscivorous birds that occur in the Delta (e.g., terns, gulls, cormorants, and herons) depredate salmonids elsewhere (Osterback et al. 2013; Evans et al. 2016; Sherker 2020). Mammalian (e.g., river otters *Lutra canadensis*, harbor seals *Phoca vitulina*, and sea lions) depredation of juvenile salmonids also occurs in other ecosystems (Dolloff 1993; Yurk and Trites 2000; Chasco et al. 2017) and is possible in the Delta.

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Experimental design.— We conducted our study during spring 2019 (April 22, 2019 to June 7, 2019) at five sites in the western Sacramento–San Joaquin Delta (Figure 1). For each of the first two study weeks, predation experiments occurred on four consecutive nights within a single site (L3, L4; Figure 1). We randomly chose the light location within experimental reaches (upstream or downstream) without replacement to ensure that both locations were equally represented each week. After the first 2 weeks, we found an insignificant effect of light location on relative predation risk with a Cox proportional hazards model $(P = 0.69; \text{ Cox } 1972)$. Therefore, we only visited new experimental sites on two consecutive nights, and we randomly selected the single light location before experiments began. Experiments occurred at three more sites (L6, L7, L8; Figure 1) using this design; however, predation at these sites was almost nonexistent (mean 2% predation). Therefore, we revisited the second site (L4) over two more weeks and employed the two-night experiment once each week.

We delineated our experimental reaches within each site with floating lines, measuring 200 m in length (alongshore) and 25–50 m in width (perpendicular to shore). All reaches had riprapped shorelines, and we defined within-reach treatments as the 100-m upstream and downstream sections of the reach. We mounted the artificial light source on a 4-m pole and placed it at the waterline halfway within the upstream (50 m into reach) or downstream (150 m into reach) portion of the reach, depending on the experimental treatment. The light source was two LED floodlights, which emitted 20,000 lx, and it was oriented parallel with the waterline (Figure 2).

To assess the relationship of ALAN intensity with the relative predation risk of juvenile salmon, we used predation event recorders (PERs; Demetras et al. 2016). Each PER was an independent drifting GPS-enabled platform baited with a tethered hatchery-origin live Chinook Salmon smolt (mean $TL = 87$ mm; $SE = 0.01$) at 1 m depth. To tether each smolt, we looped fluorocarbon fishing line through their mouth and operculum. We attached this tether to a magnet that initiated a timer when it was pulled by a predation event (event time). To determine the exact location and time of predation events, we subtracted event time from deployment end time. On average, we had 82 (SE = 3.6) PER deployments each night that typically began 1 h after sunset and continued for 4 h. To ensure that PERs traversed both light and dark treatments, we deployed PERs at the reach end where tidal flow carried them through the entire reach and spread them along the width of this end. When PERS reached the opposite end of the 200-m reach, they were collected and returned to the starting point. Before redeployment, we ensured that each PER had an active tethered smolt in good condition and used smolts two to three times when no predation occurred.

FIGURE 1. Experimental sites within the Sacramento–San Joaquin Delta. The top left panel is the state of California and Central Valley rivers. Sacramento (SAC), Stockton (ST), and San Francisco (SF) are major cities along salmonid out-migration routes denoted by gray diamonds. The lower left panel is the extent of the Sacramento–San Joaquin Delta, and the main panel shows the location of experimental sites.

FIGURE 2. Schematic of an experimental study reach. We positioned floating lines parallel to the shoreline to ensure that predation event recorders (cylinders) drifted through water that our introduced LED light source (yellow rectangles) illuminated. Largemouth Bass and Striped Bass represent potential predators. Adaptive resolution imaging sonar (ARIS) cameras and the artificial light source are depicted at their respective positions within the experimental design.

Predation event recorders do not provide an estimation of absolute predation rates as tethered prey are less able to evade capture. However, PERs provide a cost-effective method to investigate drivers of predation and predator response that would be difficult to detect otherwise (Demetras et al. 2016; Michel et al. 2020a, 2020b). Our PERs were slightly modified versions of those described in detail in Demetras et al. (2016). Specifically, we constructed PERs with 5.08-cm-diameter clear PVC pipe with the majority of components (GPS, timer, reed switch) contained within the PVC housing and sealed with a rubber end cap. Given the PERs' ability to capture precise predation locations and times, we were able to associate ALAN intensity (lux) with each observed predation event.

To compare predator density among light and dark treatments, we deployed ARIS (adaptive resolution imaging sonar; Sound Metrics) cameras in experimental reaches at 50 and 150 m (Figure 2). We positioned ARIS cameras at approximately 2 m depth, with one ARIS located directly offshore of the light source (light treatment) and the other at a distance of 100 m from the light source (dark treatment). Cameras continuously recorded from the start of PER deployments to the end, with a viewing window of $2-10$ m, a -1° pitch, and a lens heading perpendicular to the reach length. In addition to among-treatment comparisons, ARIS cameras provided predator density data that we incorporated into our predation risk models.

Light surveys.— After all PERs were retrieved, we surveyed surface light intensity (lux) within experimental reaches from a motorized vessel using an optometer (International Light Technologies; ILT2400). We performed survey transects at the inside floating line, the middle of the reach, and the outside line. We measured light attenuation with depth directly parallel to the light source at the nearshore and offshore limits of the study site. Starting at the surface, we lowered the optometer at 0.5-m intervals until the bottom or 4 m was reached and held it for 1 min at each unique depth to record a mean lux value. To account for variation with distance from the light source and inherent variability among nights, we standardized lux at depth by dividing the value at each depth by the surface value from each cast. This value was light attenuation (*At*), and we fit the following exponential decay equation:

$$
At = e^{[-(kd * depth + kt * turb)]},
$$

where *kd* (attenuation with depth) and *kt* (attenuation with turbidity [*turb*]) were fit with both coefficients or only *kd*. The model of best fit had a corrected Akaike information criterion (AIC_c; -467.772) within 2 units of the model with the lowest AIC_c (-469.411) and included the fewest parameters, *kd* (1.279) and depth.

We interpolated lux across experimental reaches using the "autoKrige" function (Hiemstra et al. 2009) in the automap package for R and assigned each PER GPS position a lux value at the water surface and at depth. This function generated an exponential variogram (we fixed this model type within the function) from each survey and used weighted least squares to select the bestfitting values of nugget, range, and sill. Using this model, we interpolated water surface lux over a 500,000-cell grid, which resulted in smooth, fine-scale lux values across experimental reaches for each night (Figure 3). We assigned interpolated surface lux values to each PER GPS position and used the above attenuation model to predict lux at 1 m depth—the approximate depth where tethered smolts

FIGURE 3. Interpolated lux values for site L4 on sampling night 8 (L4_8). Predation proportion rasters are shown for late-night (\geq 180 min past sunset) and all-night data sets. We produced predation proportion rasters by generating kernel densities of all predation event recorder (PER) predation events and dividing these by kernel densities of all PER GPS positions.

drifted. If any PERs drifted outside of the survey bounds, we assigned the lux value from the nearest grid cell. Our lux meter malfunctioned during the second night of sampling at L7, so this night was removed from PER predation risk analysis.

Processing and reduction of ARIS data.— To postprocess ARIS footage, we used Echoview version 10.2. This software removed background data and excess noise, and it identified all fish or fish-like objects (targets) \geq 200 mm in the ARIS footage (Boswell et al. 2008). We then manually reviewed each fish-like object and removed all nonfish before analysis. Although both ARIS cameras were deployed with similar settings each night, frame rate differed between and within cameras, given inherent processing speed differences between computers throughout sampling nights. To account for differing frame rate and instances of Echoview assigning multiple unique identifiers to the same fish (double counting), we used fish density instead of fish counts for analysis. To calculate fish density, we exported total beam sampling volume and the total number of fish pings in 30-min increments from each ARIS (light, dark ALAN treatments) on each experimental night. We then divided the number of fish pings by beam volume within the corresponding 30-min time frame to obtain fish per $m³$ for each 30 min of sampling within each treatment on a given night. The number of fish pings and beam volume within a given time frame are inherently tied to frame rate. For example, ARIS A has a frame rate of 4 frames/s and ARIS B has a frame rate of 8 frames/s, and each ARIS samples 4 m^3 each frame. Assuming that one fish is continually present on both ARIS A and B for an entire 30-min window, ARIS A would sample 28,800 $m³$ and have 7,200 fish pings in 30 min, and ARIS B would sample $57,600 \text{ m}^3$ with 14,400 fish pings. The resulting 30-min fish density for both hypothetical ARIS cameras would be 0.25 fish/m³ per 30 min. The ARIS in the light treatment malfunctioned during the sixth night of sampling at L4 and density data were not recorded for this night; therefore, it was censored from all statistical analyses. Although we did not identify large fish species, it is likely that these fishes are mainly piscivores (see Discussion), and fish density will be referred to hereafter as predator density.

Statistical analysis.— To determine if ALAN altered predator density and if predator density was related to time of night, we analyzed ARIS data with a generalized linear mixed-effects model (GLMM) using the lme4 package in R (Bates et al. 2015; R Core Team 2019). We fit the GLMM using the Gamma family and a log link, given that the data distribution of 30-min predator density was nonnormal and nonnegative. To remove the few zeros in the data set $(n = 13; 5.75\%$ of data) and allow the Gamma model to run, we added 1×10^{-9} to all density measurements before GLMM analysis. We included ALAN treatment, minutes past sunset (30-min bin increments 1–8), and the interaction of ALAN and minutes past sunset as independent variables in the GLMM. To account for potential differences in baseline predator density among sampling nights and the fact that each ALAN treatment was resampled throughout each night, we included a random effect of sampling night in the GLMM. To investigate the interaction of minutes past sunset and ALAN on predator density, we split the data into early (1–3 h past sunset) and late-night (3–5 h past sunset) subsets and ran GLMMs without the interaction term.

To assess whether the relative predation risk of Chinook Salmon smolts was related to lux, time of night, and predator density, we evaluated Cox proportional hazards models with the R function "coxph," in the survival package (Cox 1972; Therneau 2015). We included lux, minutes past sunset, mean 30-min predator density among light and dark treatments throughout experimental nights, and the interaction of lux and minutes past sunset as independent variables. To investigate the interaction of minutes past sunset and lux, we split the data into early (1–3 h past sunset) and late-night (3–5 h past sunset) subsets and ran Cox models without the interaction term. We checked proportional hazards assumptions of Cox models using the "cox.zph" function in R, within the survival package (Therneau 2015). Although lux at depth was predicted from surface lux, we evaluated another set of Cox models using this variable instead of surface lux to demonstrate how predation risk responded to changes in lux at depth. To investigate the relative effect on predation risk of each variable, we also ran the Cox models using scaled (independent) variables. Neither of these approaches changed overall model fits or significance; however, they provided further insight that we might have missed in original model output.

RESULTS

We found that predator density was positively related to ALAN presence late in the night. When we analyzed all density data, the effects of minutes past sunset and ALAN presence on predator density were nonsignificant; however, a significant interaction between these two variables was present (Table 1). In the early-night model, neither minutes past sunset nor ALAN presence had a significant relationship with predator density. While in the late-night model, predator density was predicted to increase by a factor of 3.96 in the presence of ALAN (Figure 4A), and no relationship of minutes past sunset was detected (Table 1). This difference in the ALAN effect between early- and late-night models demonstrates why the interaction of ALAN and minutes past sunset was significant in the overall model.

Relative predation risk of Chinook Salmon smolts increased with artificial illumination intensity (lux) late in the night and was positively related to predator density in all models (Figures 3, 4B, and 5). In the full-night model, relative predation risk had a positive relationship with predator density, a negative relationship with minutes past sunset, and a positive relationship with the interaction of minutes past sunset and lux. However, no overall relationship of relative predation risk with lux was detected (Table 2). Neither minutes past sunset nor lux was significantly related to predation risk in the early-night model, but the positive relationship with predator density remained. In the late-night model, relative predation risk was predicted to increase by factors of 1.030 (scaled factor = 1.217) and 1.502×10^{19} (scaled factor = 1.987) for each unit increase in lux and predator density, respectively, and no significant effect of minutes past sunset was detected (Table 2; Figure 5). For each unit increase of lux at 1 m depth in the late-night model, the raw factor change was greater (1.111). However, this increase was

TABLE 1. Results of predator density generalized linear mixed models for all data, early-night (1–3 h past sunset), and late night-data subsets (3–5 h past sunset). Coefficient estimates of the presence of ALAN, minutes past sunset (Min), and their interaction are given. Exponentiated estimates (*e^{est}*), standard error (SE), *t*, and *P*-values are also reported for each parameter.

	Estimate	e^{est}	SE		P
		All data			
ALAN	-0.310	0.734	0.448	-0.691	0.490
Min	-0.015	0.985	0.060	-0.254	0.799
$ALAN \times Min$	0.212	1.237	0.090	2.358	0.018
		Early data			
ALAN	0.207	1.230	0.280	0.740	0.459
Min	0.022	1.022	0.108	0.201	0.841
		Late data			
ALAN	1.376	3.957	0.297	4.627	< 0.001
Min	0.108	1.114	0.140	0.771	0.441

FIGURE 4. Mean large fish (\geq 200 mm TL, likely predators) density (large fish/m³ per 30 min [\pm 1 SE]) (A) across each time bin within both light and dark treatments, and (B) across each experimental night $(\pm 1 \text{ SE})$ within both light and dark treatments. The estimated nightly mean PER predation risk (filled circles) is shown from the overall Cox model. Continuous lines under experimental nights represent consecutive experimental nights at a given site.

because lux at depth only varied from 0 to 20, while lux at the surface varied from 0 to 72 (Table 2). Similar to the predator density models, the significant interaction of lux and minutes past sunset in the overall model was the result of a significant effect of lux late in the night with no effect early in the night on relative predation risk.

DISCUSSION

With continuing human development along rivers, estuaries, and coastlines, it will be increasingly important to consider the impacts of ALAN on aquatic organisms and ecosystems (Davies et al. 2014; Jechow and Hölker 2019; Zapata et al. 2019). Elsewhere, ALAN has attracted

young salmonids and their predators, resulting in elevated predation rates (Tabor et al. 2004, 2017). Similarly, introduced ALAN in the Delta increased predator density and relative predation risk of Chinook Salmon smolts; however, these effects were only detected 3 h past sunset and later. These results indicate that ALAN reduction may decrease predation rates and mortality of out-migrating salmonids in the Delta and are an important first step in assessing ALAN impacts along out-migration routes in the California Central Valley and elsewhere.

As with any field experiment, there were notable environmental limitations and assumptions associated with our study. First and foremost, our metric of relative

FIGURE 5. The nightly mean percentage $(\pm 1 \text{ SE})$ of PER observations when predation events occurred (Pred %) binned across increasing lux values (top row) for early-night (1–3 h after sunset; column 1) and late-night (3–5 h past sunset; column 2) data subsets. Predicted relative predation risk of increasing lux (row 2) and large fish (≥200 mm, likely predators) density (row 3). Predictions (solid black line) above the horizontal dashed line (risk = 1) indicate increased relative predation risk and below the dashed line represent decreased risk, with 95% confidence intervals in gray. The mean observation of each variable within each model is where the solid black line crosses the dashed line.

predation risk focused only on predators and did not account for any prey effects, given that we used tethered smolts. Nonetheless, the PER technique has been used to investigate environmental drivers of predation (Demetras et al. 2016; Michel et al. 2020a, 2020b), and the drifting PER tether likely mimics natural prey behavior better than traditional fixed-tethering experiments. Given this tethering limitation and the focus on predators, future research should determine how ALAN affects free swimming out-migrating smolts. Detailed analysis of past, current, and future smolt telemetry data comparing migration speed, timing, and mortality in illuminated and dark river reaches (e.g., Celedonia et al. 2011) is one way to elucidate these ALAN effects. Field experiments where free-swimming acoustically tagged smolts are released under artificially illuminated and dark conditions coupled with a high-resolution telemetry array could also be used to determine if ALAN affects smolt movement and migration. Artificial illumination may reduce migration speed and survival because salmon out-migration is predominately nocturnal (Chapman et al. 2013; Clark et al. 2016; Furey et al. 2016) and ALAN attracts and slows juvenile salmonids (Tabor et al. 2004, 2017; Celedonia et al. 2011; Riley et al. 2013). This attraction may bring salmonids closer to shore, exposing them to mammalian (e.g., river otters; Dolloff 1993) and avian piscivores (e.g., herons;

TABLE 2. Results of Cox proportional hazard models for all data, early-night (1–3 h past sunset), and late-night (3–5 h past sunset) data subsets. Predictor variables include light intensity (Lux) or lux at depth (Lux [Depth]), minutes past sunset (Min), and 30-min predator (fish ≥200 mm) density (Pred Den). The coefficient (*coef*) column indicates the direction (positive versus negative) of the effect of each predictor variable on relative predation risk and coefficient standard error (SE [*coef*]) is reported. The *e coef* columns report the relative predation risk change for each unit increase in predictor variables of both raw and scaled variables. The *z* column is the Wald statistic value, which evaluates whether the coefficient of a given variable is significantly different from 0, and the significance value is indicated in the *P* column. The *P-*value for the "cox.zph" test (which checked proportional hazards assumptions of Cox models) is also reported in the *p_zph* column.

	coef	raw e^{coef}	SE (coef)	scaled e^{coef}	Z	\boldsymbol{P}	p_{zph}
			All data				
Lux	-0.061	0.940	0.041	0.670	-1.494	0.135	0.782
Lux (Depth)	-0.220	0.803	0.147	0.670	-1.494	0.135	0.782
Min	-0.004	0.996	0.001	0.472	-3.123	0.002	0.227
Pred Den	42.3099	2.369×10^{18}	4.228	1.843	10.006	< 0.001	0.590
$Lux \times Min$	0.0004	1.0004	0.0002	1.625	2.228	0.026	0.696
			Early data				
Lux	-0.006	0.994	0.022	0.964	-0.267	0.789	0.879
Lux (Depth)	-0.021	0.979	0.078	0.964	-0.267	0.789	0.879
Min	-0.006	0.994	0.004	0.429	-1.723	0.085	0.490
Pred Den	41.111	7.150×10^{17}	6.501	1.708	6.323	< 0.001	0.591
			Late data				
Lux	0.029	1.030	0.009	1.217	3.017	0.003	0.770
Lux (Depth)	0.105	1.111	0.035	1.217	3.017	0.003	0.770
Min	-0.006	0.994	0.003	0.252	-1.902	0.057	0.281
Pred Den	44.156	1.502×10^{19}	5.979	1.987	7.385	< 0.001	0.404

Sherker 2020), as well as benthic (e.g., sculpin and catfish) and vegetation-associated (e.g., Largemouth Bass) piscivorous fishes (Michel et al. 2018, 2020a). If ALAN sufficiently slows out-migration, increased water temperatures, low river flows, and prolonged exposure to predators along the out-migration corridor could also reduce salmonid survival (Henderson et al. 2019; Michel et al. 2020a). Although our study found a response of predators to ALAN, future investigations focused on prey responses will help elucidate if ALAN reduction is a viable restoration strategy.

While we were not able to identify the large fishes used for predator density calculations to species, it is likely that many were piscivores, given the positive relationship between density and relative predation risk. Furthermore, we conducted opportunistic hook-and-line sampling on a few occasions prior to nightly experiments, and Largemouth Bass and Striped Bass were caught and released at site L4. These species are common predators of salmonid smolts in the Delta (Sabal et al. 2016; Michel et al. 2018) and were likely a major component of the predator community in this study. Because this study was conducted at night, we could not use traditional video recordings to identify PER predators. Coupling underwater infrared light sources with traditional video (Mazur and Beauchamp 2003) may aid in nighttime predator community identification without introducing additional light in the visible spectrum (Jordan and Howe 2007; Horodysky et al. 2010; Mitchem et al. 2018). However, the sampling area of traditional video would likely be less than it was for our ARIS cameras, and traditional video would be problematic in turbid waters. Combining ARIS with infrared-assisted video analysis could be advantageous for lowlight predator density and community analysis. Our predator density metric also did not differentiate between individual fish remaining under the light source or new individuals being attracted to the light, distinctions that are potentially important but are pooled in our analysis. Future work should investigate differences in fish behavior in dark and illuminated treatments to determine how ALAN may change the behavioral response of fishes.

Predator density was greater in ALAN treatments and relative predation risk of Chinook Salmon smolts increased with increasing ALAN intensity (lux) 3 h past sunset and later. However, no ALAN impacts on density or predation risk were detected early in the night. These discrepancies may be driven by ambient light levels and the relationship of predation success with light. Foraging efficiency and predation success of piscivorous fishes increases with increasing light levels, but fish can also successfully forage in low light (McMahon and Holanov 1995; Mazur and Beauchamp 2003, 2006). Mazur and Beauchamp (2003) suggested that salmonids have a foraging threshold around 1 lx, with success continually increasing until it asymptotes around 20 lx. However, for Largemouth Bass, this threshold could be as low as ambient moonlight (0.003 lx; McMahon and Holanov 1995). Early in experimental nights, ambient light levels may have been sufficient for successful foraging and depredation throughout experimental reaches, resulting in similar predator densities and predation risk throughout the reach. After 3 h past sunset, an ambient light foraging threshold may have been reached where predation success was diminished in unilluminated reach sections. The additional light provided by ALAN likely allowed for continued unimpeded predation success in illuminated portions of the reach. This increased success likely led to greater predator density in ALAN treatments and increased predation risk with increasing lux late in the night. Therefore, the overall decrease in predation risk with elapsed minutes past sunset may be explained by decreasing successful foraging area with time.

The late-night relationship of elevated predation risk with increasing lux was also likely driven by increased predator density in ALAN treatments, given that predation risk was positively correlated with predator density in all models. The increased foraging success of predators in ALAN would have also contributed to the elevated predation risk with lux. However, decoupling these two drivers was not possible in our experiment. To elucidate foraging thresholds and quantify the effects of ambient illumination on predation risk, future studies should deploy a light meter throughout the night to monitor background illumination.

The lack of ALAN effects on predation early in the night could also be a result of the predator community we sampled and our experimental lights. As mentioned above, predation risk in our study was higher earlier in the night, but ALAN did not affect density or predation risk during this time. Assuming that the majority of our predators were roving channel-associated fishes (e.g., Striped Bass), it may have taken time for them to cue into the light, which we established each night. If our light was a permanent fixture, perhaps fishes and other piscivores would have established nocturnal feeding habits under this fixture early in the night. For example, harbor seals that fed on out-migrating salmonids under an illuminated bridge were most abundant 1–1.5 h after nightfall and decreased as the night progressed (Yurk and Trites 2000). It is also possible that our experimental ALAN attracted prey fishes, which in turn attracted predators (Becker et al. 2011), but this mechanism took time to establish. Future studies could replicate our experiment at existing artificial illumination sources to see if the delayed ALAN effect exists at established light sources across variable predator communities. Recording ARIS footage continuously at illuminated and unilluminated sites could also determine how light impacts prey and predator density on a 24-h cycle (e.g., Becker et al. 2011).

Another pertinent direction for future research that would benefit ALAN management is to determine a minimum lux value that does not impact fishes. Our experiment did not empirically test variable lux intensity; however, we did sample a large range of lux values that may provide insight into predator response. If we fit a penalized spline to the lux predictor in our late-night Cox model, predation risk does not increase until after 8 lx is reached. A similar trend can be seen in the top right panel of Figure 2, with greater increases in predation percentage after 10 lx. However, we did not use this spline in our final model, given that the linear fit was much better $(\Delta AIC_c =$ 12.33), so this value should be interpreted with caution. Additionally, this minimum value may only be relevant to the system and predator community studied here. Previous work has suggested that ALAN intensities should remain as low as possible $(0.1 lx)$ to mitigate impacts on salmonids during out-migration (Tabor et al. 2004, 2017). To determine a minimum lux management goal, future experimental tests of ALAN intensity with variable treatments among experimental nights are needed.

The discrepancies between early- and late-night ALAN effects on predator density and predation risk could prove useful for ALAN management. For humans, artificial illumination is one of the building blocks of modern society and many social, recreational, and economic benefits result from ALAN (Doll et al. 2006; Hölker et al. 2010a; Gaston et al. 2014b). However, human activity is still highest during the day and early hours of the night, decreasing as the night progresses (Monsivais et al. 2017; Martín-Olalla 2018; Bhattacharya and Kaski 2019). Our results suggest that ALAN effects on predator density and predation may be minimal early in the night, but this trend could be driven by our experimental light design and predator community. If these temporal trends prove robust to future investigations, late-night lighting reduction campaigns during out-migration could maximize the human benefits of artificial illumination while minimizing the negative impacts on out-migrating salmonids. Overall, our findings align with others and suggest that ALAN increases juvenile salmonid predation (Tabor et al. 2004, 2017). Although many questions remain unanswered, it appears that reducing artificial illumination is a practical management strategy to reduce predation on out-migrating salmonids.

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ORIGINAL ARTICLE

Vadose Zone Journal

Heat transport from atmosphere through the subsurface to drinking-water supply pipes

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Abstract

Drinking-water quality in supply pipe networks can be negatively affected by high temperatures during hot summer months due to detrimental bacteria encountering ideal conditions for growth. Thus, water suppliers are interested in estimating the temperature in their distribution networks. We investigate both experimentally and by numerical simulation the heat and water transport from ground surface into the subsurface, (i.e., above drinking-water pipes). We consider the meteorological forcing functions by a sophisticated approach to model the boundary conditions for the heat balance at the soil–atmosphere interface. From August to December 2020, soil temperatures and soil moisture were measured dependent on soil type, land-use cover, and weather data at a pilot site, constructed specifically for this purpose at the University of Stuttgart with polyethylene and cast-iron pipes installed under typical in situ conditions. We included this interface condition at the atmosphere–subsurface boundary into an integrated non-isothermal, variably saturated (Richards') the numerical simulator DuMu^x 3. This allowed, after calibration, to match measured soil temperatures with $\pm 2^{\circ}$ C accuracy. The land-use cover influenced the soil temperature in 1.5 m more than the soil material used for back-filling the trench above the pipe.

1 INTRODUCTION

In recent years, unusually warm water in drinking-water supply pipes has been observed and suspected to be caused by high air temperatures during the summers (Osmancevic & Hüsam, [2021\)](#page-1252-0). Such conditions pose a threat to drinkingwater quality as elevated temperatures can lead to increased and undesired microbial activity and, hence, to deteriorated drinking-water quality. It is feared that this situation will worsen with climate change. Therefore, detailed knowledge on the influence of soil and recharge temperatures is needed

for being able to estimate the temperature in the water of distribution networks.

Generally, higher soil temperatures over the last decades are confirmed (e.g., Chen et al., [2021\)](#page-1251-0) and have various effects on microbiological and chemical processes as well as on vegetation periods that vary with time. Therefore, there are several approaches to determine soil temperatures with simple models, validated for special regions. Sharma et al. [\(2010\)](#page-1252-0) were predicting soil temperatures up to 0.5 m, based on air temperature, using linear regression in Southern Mexico. Horton Brian [\(2011\)](#page-1251-0) worked with a coefficient model, based

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on rainfall and air temperature, covering entire Australia. Jungqvist et al. [\(2014\)](#page-1251-0) were interested in Swedish forests and worked with climate models. Rankinen et al. [\(2004\)](#page-1252-0) proposed a model, working with regional soil and climate parameters, predicting soil temperatures between 20 and 50 cm, interested on snow-covered soils.

Various codes to calculate water saturations are available, which are all based on solving Richards' equation, with different priorities. Simunek and Bradford [\(2008\)](#page-1252-0) give an overview of various simulation codes available, namely HYDRUS, MODFLOW-SURFACT, STOMP, SWAP, TOUGH2, and VS2DI. HYDRUS is commonly used for variably saturated non-isothermal conditions, for example, Steenpass et al. [\(2010\)](#page-1252-0). Guo et al. (2017) use PCSiWaPro to predict seepage and concentration of pollutants. The tool WASIM is often used to describe catchments, as for example in Burkina Faso (Idrissou et al., [2020\)](#page-1251-0), where a successful validation was performed.

There are many topics, related to soil temperature and saturation, such as drainage in cyclic loading (Komolvilas & Kikumoto, [2017\)](#page-1252-0) or the influence of groundwater temperature on river temperature (Kurylyk et al., [2015\)](#page-1252-0). Kelleners et al. [\(2016\)](#page-1252-0) worked on water flow in variably saturated soils, taking into account snow. Herrada et al. [\(2014\)](#page-1251-0) proposed a model to predict infiltration rates. Corona and Ge [\(2022\)](#page-1251-0) were focusing on high intensity precipitation events. Tran et al. [\(2016\)](#page-1252-0) showed that taking heat transport into account yields better results for an inversion scheme to define soil parameters. Sandor and Fodor [\(2012\)](#page-1252-0) demonstrated the relevance of a good vegetation model to quantify root uptake. Wessolek et al. [\(2022\)](#page-1252-0) focused on heat transport and worked with earth cables as heat sources.

Literature shows various approaches to describe the temperature evolution in soils forced from the atmosphere together with water flow. One approach is to reconstruct the soil thermal field from a single measurement, which was developed and validated by Wang [\(2012\)](#page-1252-0). Another approach, one that we are also following, is to solve the energy balance at the surface, which needs expressions for evapotranspiration and sensible heat flow. This is, for example, done in Saito et al. [\(2006\)](#page-1252-0) within the simulator HYDRUS, where evaporation is set as boundary condition (Bittelli et al., [2008\)](#page-1251-0) and evaporation is part of the solution.

This study uses an approach similar to HYDRUS, calculating sensible heat flow based on the Penman–Monteith equation (Allen et al., [1998\)](#page-1251-0) by using adequate measurements of meteorological data. The evaporation is then calculated from the latent heat flow, using the Bowen ratio, as described in Section [2.2.6.](#page-1241-0) The goal of this study is to implement this kind of boundary condition in an open-access code, to calibrate against longer time series, and finally to obtain time series of the temperature of groundwater recharge which in

Core Ideas

- ∙ We measure and model the temperature of groundwater recharge.
- ∙ We solve the heat and water balance at the atmosphere–subsurface interface.
- ∙ Data: time series of hydrometeorological parameters, hydraulic, and thermal material properties.
- ∙ Goal: estimate temperatures relevant for buried drinking-water supply pipes.

this case improve the management of drinking-water networks during changing climatic conditions.

We modeled the influence of the meteorological forcing functions, the soil structure, and the land-use cover on temperatures between ground surface and the subsurface. At a pilot site, we focus with our analysis up to a depth of 1.5 m, that is, a typical laying depth for drinking-water pipes in Germany. While applied to drinking-water supply pipes, this work is of general relevance as it describes and models the temperature of groundwater recharge driven by meteorological forcing functions. This temperature distribution and its evolution with climate change are critical for resilient urban infrastructure.

A pilot site at the University of Stuttgart has been constructed, where drinking-water pipes of PE and cast iron were installed at a depth of 1.50 m over a horizontal stretch of 15 m. The trench was back-filled with two different materials: gravelly material typical for conditions when pipes are buried below streets and the naturally occurring silty clay. Two different types of land cover, natural vegetation and asphalt, have been placed on the surface. The subsurface has been instrumented with 64 temperature sensors, 8 soil moisture sensors, and detailed hydrometeorological observations are available from the neighboring University of Stuttgart's weather station. In addition to the measurements, we adapted and employed a numerical simulator for estimating both soil temperatures and moisture contents. The study aimed at including the incorporation of the meteorological forcing and the variable saturation conditions.

Section [2](#page-1238-0) describes the test site and the numerical simulation methods, starting with a description of the pilot site. Furthermore, this section introduces the numerical model and the novel approach chosen to implement the conditions at the air–subsurface interface. The results section (Section [3\)](#page-1243-0) presents the comparison between measured data and numerical simulations with a calibrated set of parameters. The results are discussed in Section [4,](#page-1249-0) followed by the conclusions (Section [5\)](#page-1250-0).

FIGURE 1 Setup of the pilot site.

2 MATERIALS AND METHODS

This section provides a description of the pilot site, then the numerical model is introduced. A particular emphasis is put on the adaption of the model to be capable of including the meteorological forcing functions in the heat and mass balances.

2.1 The experimental pilot site

Two pipes typical for drinking-water supply in terms of dimension (inner diameter: 150 mm) and material (PE and cast iron) were laid into a 1.50 m deep, 15 m long, and \approx 1 m wide trench (Figure $S1$). The pipe was set on a slight angle into beach sand with a horizontal length of 12 m, Figure 1 shows the dimensions. At both ends of the pipes, vertical extensions above ground surface were constructed. The trench was backfilled with two types of porous media, (a) a silty clay, the naturally occurring material at the site and (b) gravelly material typical for street construction ("KFT gravel" and "sieved broken gravel"). The pipes were filled with drinking water. On the ground surface, two \approx 12 cm thick layers of different landuse were established, (a) the naturally occurring vegetation ("grass") and (b) asphalt. Thus, in total, four combinations of porous media and land-use were replicated, the stations referenced by Roman numerals. Each station was insulated against thermal influences from the neighboring sites.

The trench was installed in direct vicinity to [\(https://lhg-](https://lhg-902.iws.uni-stuttgart.de)[902.iws.uni-stuttgart.de\)](https://lhg-902.iws.uni-stuttgart.de) the weather station of the University of Stuttgart, where the following hydro-meteorological variables are measured:

- ∙ long wave radiation incoming
- ∙ short wave radiation incoming
- ∙ air temperature in 2 m above ground
- ∙ wind velocity in 2 m above ground
- ∙ relative humidity in 2m above ground
- ∙ precipitation intensity

At the site, a stationary phreatic water table at a depth of ~5.1 m below ground surface has been observed, sitting on top of a local low-conductive layer ("Arietenkalk"). Atmospheric pressure was set constant to a value of 101,325 Pa, as its influence on vapor pressure was considered negligible.

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The variables soil temperature and soil moisture were observed with 64 sensors (24 temperature sensors in the porous media, 16 temperature sensors in each pipe, 8 soil moisture sensors) with the aim to monitor the evolution of temperature and soil moisture from the ground surface toward the water pipes. Temperatures were measured using PT1000 sensors, soil moistures using "Teros 10" sensors from the metergroup. Time series of all variables were recorded in 5 min intervals, backed up, and stored in a hdf5 file using python (Virtanen et al., [2020\)](#page-1252-0) for data processing.

The observed data will be compared to numerical simulation results obtained by including meteorological forcing functions at the ground surface on both heat and water balances.

2.2 The numerical model

2.2.1 Balance equations

The modeling of heat transport in a partially water saturated porous medium requires solving the coupled balances for water and heat. Our system of interest is located in the variably saturated zone, where the water balance can be formulated in terms of the Richards' equation, which can be considered as a simplified version of the multiphase flow equation under the assumption of an infinitely mobile gas phase. Richards' equation combines the multiphase version of Darcy's law with the continuity equation for water, thus taking into account the effects of capillary pressure and relative permeability:

$$
\frac{\partial \phi S_{\rm w} \rho_{\rm w}}{\partial t} - \nabla \left\{ \rho_{\rm w} \frac{k_{\rm rw}}{\mu_{\rm w}} \mathbf{K} \left(\nabla p_{\rm w} - \rho_{\rm w} \mathbf{g} \right) \right\} = q_{\rm w} , \qquad (1)
$$

where S is the saturation, ϕ the porosity, ρ the density, μ the dynamic viscosity, k_r , the relative permeability as a function of saturation, p the pressure, q_w the source term for water, and K the hydraulic conductivity tensor. The index w denotes water. The Richards' model assumes that the pressure of the gas phase is constant, that is, in our setting at atmospheric conditions, 1 bar. The pressure of the water phase depends on the water saturation via the capillary-pressure-saturation relationship as will be detailed further below.

Assuming thermal equilibrium between solid and fluids, the heat balance coupled to the Richards' equation can be written as follows:

$$
\phi \frac{\partial (\rho_{\rm w} u_{\rm w} S_{\rm w})}{\partial t} + (1 - \phi) \frac{\partial (\rho_{\rm s} c_{\rm s} T)}{\partial t} + \nabla \cdot (\rho_{\rm w} h_{\rm w} \mathbf{v}_{\rm w} - \lambda \nabla T)
$$

= $q_{\rm h}.$ (2)

Here, u_w is the specific internal energy of the water phase, c_s is the specific heat capacity of the solid, *ℎ* the specific enthalpy, λ represents the averaged heat conductivity of the fluid-filled porous medium, v_w is the velocity of the water phase, which is obtained from Darcy's law as it is already implicitly inserted in Equation (1), T is the temperature, and q_h the source term for heat.

2.2.2 Numerical simulation platform

For numerical implementation, we were working with DuMux 3 (Koch et al., [2020;](#page-1252-0) Scheer et al., [2020\)](#page-1252-0), an opensource simulator and research code for flow and transport in porous media. With its modular design, DuMux allows for a flexible choice of physics and discretization methods and solution algorithms, and it facilitates the implementation of new approaches and adaption as in this study. For spatial discretization, we used the Box method for this study; for details, we refer to Scheer et al. [\(2020\)](#page-1252-0). The Box method employs a finite element mesh containing the nodes at which the solution is calculated, while there is a secondary mesh constructed, on which a finite-volume scheme is used. Thus, the Box method guarantees local mass conservation.

2.2.3 Hydraulic properties

Richards' equation describes water flow in partially saturated porous media. Thus, the gas phase is present, though not explicitly modeled. Using the Richards' approach of multiphase flow in the variably saturated zone implies that the relative permeability of the water phase as a function of water saturation is considered as well as the capillary pressure also dependent on water saturation. We aim to reduce the overall uncertainty by measuring relevant parameters (saturated hydraulic conductivity, relationship between capillary pressure and saturation). In situations where measurements are difficult, for example, because of large grain sizes such as in the gravel for this study, the uncertainty due to the parameterisation is considered minor compared to the uncertainty of the overall behavior. For the variably saturated relationships, several models exist in addition to the van Genuchten parameterisation, for example, Brooks and Corey [\(1966\)](#page-1251-0) or Clapp and Hornberger [\(1978\)](#page-1251-0). Working with a different approach can lead to different results, as demonstrated in Yang and Wang [\(2014\)](#page-1252-0). We compared parameters of a Brook–Corey model to van Genuchten's model and saw similar results for both models but a better stability for van Genuchten's model. For this reason, in this study, the relative-permeability-saturation relationship as well as the capillary-pressure-saturation relationship were used according to van Genuchten's model. It **274 Vadose Zone Journal** $\mathbf{z} = \mathbf{0}$

can be written for the capillary pressure as:

$$
p_c = \frac{1}{\alpha} \left(S_e^{-1/m} - 1 \right)^{1/n}
$$
 (3)

with the effective saturation

$$
S_e = \frac{S_{\rm w} - S_{\rm wr}}{1 - S_{\rm wr}} \,. \tag{4}
$$

 α [1/Pa] and *n* are parameters to be determined specifically for the porous material. α is a scaling parameter for the magnitude of the capillary pressure (comparable to the entry pressure in Brooks–Corey's approach), and the parameter n characterizes the uniformity or non-uniformity of the pore-size distribution. A large n expresses a comparatively uniform soil. Commonly, the parameter m is expressed in terms of n by

$$
m = 1 - \frac{1}{n} \tag{5}
$$

The relative-permeability function is, accordingly:

$$
k_{\text{rw}} = \sqrt{S_e} \left[1 - \left(1 - S_e^{1/m} \right)^m \right]^2. \tag{6}
$$

We measured α , *n*, and saturated conductivity for the natural material with the simplified evaporation method as described in Peters and Durner [\(2008\)](#page-1252-0) with a HYPROP device. This method was not suitable for the sandy gravel, as these materials are too coarse for measurements, hence their parameters were calibrated. Asphalt is basically not permeable, thus should receive a close-to-zero permeability. However, in our case, the 1D simplification of the model domain cannot account for effects that occur close to the surface and lead to small lateral inflow of water, which contradicts the idealized 1D assumption of an impermeable boundary. Thus, the apparent permeability of the asphalt in the simplified 1D system is part of the calibration procedure. For the grass surface, it is difficult to perform representative measurements. Consequently, all permeabilities were used as calibration parameters. Another key hydraulic property is the porosity. Again, this could be measured for some soils (e.g., sand), for others, (e.g., grass) it would have been difficult to measure. For this reason, properties were also used as calibration parameters.

2.2.4 Thermal properties

Basically, the thermal conductivity and the heat capacity are material- and temperature-dependent properties of the porous medium. While temperature dependence is small in the range of temperatures encountered in our application (-10 to 40˚C), water content plays a more important role. In reality, we have

an inhomogeneous mixture of soil, water, and air. Thus, the water saturation has an effect on the thermal properties. It is challenging to find an appropriate model or a set of tabulated values for a given material. There are several models available to calculate thermal conductivity based on water content, for example, Ghanbarian and Daigle [\(2015\)](#page-1251-0); Lu et al. [\(2007\)](#page-1252-0); Lu and Dong [\(2015\)](#page-1252-0); Markert et al. [\(2017\)](#page-1252-0); Sadeghi et al. [\(2018\)](#page-1252-0).

The effective heat capacity can be computed as the volumetric average of water, air, and solid heat capacity. The effective thermal conductivity of a wet solid matrix with a given water content depends on the grain sizes. Johansen [\(1975\)](#page-1251-0) introduced a way to determine the effective thermal conductivity by defining the effective thermal conductivity and using K_e , the Kersten number:

$$
\lambda_{\rm eff} = \lambda_{\rm dry} + K_e(S_{\rm w}) \big(\lambda_{\rm wet} - \lambda_{\rm dry} \big). \tag{7}
$$

Thus, the effective conductivity is calculated from the values for the dry conductivity and the wet conductivity. There are several models available based on measured correlations (e.g., Somerton et al., [1974\)](#page-1252-0). For this research, we were working with Lu et al. [\(2007\)](#page-1252-0), as the temperatures and soils used for the underlying correlation fit to our test settings. In their approach, the definition for K_e is

$$
K_e(S_{\rm w}) = \exp \alpha_{Lu} \left[1 - S_{(\alpha_{Lu}-1.33)}^{\rm w} \right],\tag{8}
$$

and the dry conductivity is calculated as

$$
\lambda_{\text{dry}} = -a \cdot \phi + b,\tag{9}
$$

where α_{I_u} is 0.27 for fine-textured soil and 0.96 for coarsetextured soil, a and b are empirical parameters and are determined to be 0.56 and 0.51 with ϕ being the porosity. Typically, the saturated (wet) conductivity is calculated via Johansen [\(1975\)](#page-1251-0):

$$
\lambda_{\text{wet}} = \lambda_{\text{solid}}^{(1-\phi)} \times \lambda_{\text{w}}^{\phi}.
$$
 (10)

2.2.5 | Model domain

The dominant force at our test site in heat and water transport acts in the z-direction (from surface to ground). There is no force for a transport to occur in a horizontal direction. The zones corresponding to various porous media and land-use covers extend horizontally for short distances only (perpendicular to the pipes' length about 0.4 m, parallel to the pipes' length about 3 m) and are thermally insulated with polystyrene. Thus, boundary effects in horizontal directions are very small and can be neglected. For this reason, we chose

$$
\frac{\text{Sensible heat}}{\text{Latent heat}} = \frac{H_{\text{sens}}}{H_{\text{ET}}} = \gamma \times \frac{\frac{\partial \overline{T}}{\partial z}}{\frac{\partial \overline{e}}{\partial z}}.
$$
 (13)

The ratio of latent (H_{ET}) to sensible (H_{sens}) heat flux is equal to the ratio of the vertical gradients of temperature T with vapor pressure e multiplied by a constant, the so-called psychometric constant γ . This is based on the assumption of their similar turbulent diffusion coefficients. Now, we are approximating the gradients of e and T over a height z by the difference of their measured values at 2 m (e_m and T_m) and at surface (e_{surf} and T_{surf}). Then, we solve the equation by H_{ET} :

$$
\frac{H_{\text{sens}}}{H_{\text{ET}}} = \gamma \times \frac{T_{\text{surf}} - T_{\text{m}}}{(e_{\text{surf}} - e_{\text{m}})} \to H_{\text{ET}} = \frac{(e_{\text{surf}} - e_{\text{m}}) \times H_{\text{sens}}}{(T_{\text{surf}} - T_{\text{m}}) \times \gamma}.
$$
\n(14)

This gives us a relation for H_{sens} and H_{ET} , but does not define one of them. H_{ET} is defined as the heat flux at phase transition, this means latent heat of vaporization times density of water $\Delta H_{\text{van}} \times \rho_{\text{w}}$ times the rate of evapotranspiration ET :

$$
H_{\text{ET}} = ET \times \Delta H_{vap} \times \varrho_w \tag{15}
$$

To approximate ET , we use the Penman–Monteith equation, following Allen et al. [\(1998\)](#page-1251-0). Assuming a large surface of constant vegetation height, calculations based on boundary layer theory and Prandtl–Karman's velocity distribution lead to:

$$
ET = \frac{m_{\text{H}_2\text{O}}}{\rho_w RT_{\text{abs}}} \frac{\kappa^2}{\left[\ln\left(\frac{z_{\text{m}} - t_{\text{d}}}{z_0}\right)\right]^2} u_{\text{m}} \times (e_{\text{surf}} - e_{\text{m}})
$$
\n
$$
= \frac{m_{\text{H}_2\text{O}}}{\rho_w RT_{\text{abs}} \times r_{\text{a}}} \times (e_{\text{surf}} - e_{\text{m}}),
$$
\n(16)

with the universal gas constant R, the von Karman constant κ the absolute temperature T_{abs} in K and r_{a} defined as

$$
r_{\rm a} = \frac{\left[\ln\left(\frac{z_{\rm m}-t_{\rm d}}{z_0}\right)\right]^2}{\kappa^2 \times u_{\rm m}},\tag{17}
$$

which represents the aerodynamic resistance, $m_{\text{H}_2\text{O}}$ is the molar mass of water. Combining Equations (14)–(16), we end up with:

$$
\frac{H_{\text{sens}}}{ET \times \Delta H_{\text{vap}} \times \varrho_{\text{w}}} = \frac{H_{\text{sens}} \times RT_{\text{abs}} \times r_{\text{a}}}{m_{\text{H}_2\text{O}} \times \Delta H_{\text{vap}} \times (e_{\text{surf}} - e_{\text{m}})}
$$

$$
= \frac{(T_{\text{surf}} - T_{\text{m}}) \times \gamma}{(e_{\text{surf}} - e_{\text{m}})},\tag{18}
$$

2.2.6 Boundary conditions

The boundary condition at the bottom of the domain at the groundwater table was modeled as a Dirichlet boundary, which means, constant temperature and pressure were set. Temperature was set to the long-term groundwater average of 10˚C; pressure was set to the equivalent of 0.9 m, which is the long-term water table at 5.1 m below ground surface. The boundary conditions at the ground surface for heat and water flow were set as Neumann boundary conditions, implying that both heat and water fluxes needed to be defined. To determine the flux values required for the Neumann condition, we computed heat and water balances at the interface.

Calculating Neumann conditions at the interface

The system behavior is driven by water and heat exchange with the atmosphere. We assume a constant temperature, as we reach lower soil layers, where conditions should be constant over time. Radiation and precipitation measured values of environmental influences, and the resulting heat and water fluxes need to be calculated to impose these influences as Neumann boundary conditions.

Heat

The balance of heat at the boundary can be written as:

Soil heat flux = Net radiation $-$ Latent heat flux

−Sensible heat flux − Heat flux due to net

$$
water flux. \t(11)
$$

The temperature of the infiltrating recharge as well as the contribution stemming from the evapotranspirating water is considered via the specific heat capacity of water and the temperature of air at 2 m height.

Incoming long and short wave radiation values are measured directly. Taking into account the albedo (ratio of incoming and outgoing short wave radiation; ω) of the surface and Boltzmann's law for outgoing long wave radiation, the resulting balance is:

Net radiation =
$$
H_{\text{short,in}} \times (1 - \omega) + H_{\text{long,in}} - \sigma \times T_{\text{abs}}^4
$$
. (12)

well as for asphalt.

Approximating r_a with 208/ u_m (Allen et al., [1998\)](#page-1251-0), with the definition of the specific gas constant for air $R_s = R/m_{air}$, and $\epsilon = m_{\text{H}_2\text{O}}/m_{\text{air}}$ our boundary condition for H_{sens} results in:

$$
H_{\text{sens}} = \frac{\epsilon \times \Delta H_{\text{vap}} \times \gamma}{T_{\text{kv}} \times R_{\text{s}}} \times \frac{u_{\text{m}}}{208} \times (T_{\text{surf}} - T_{\text{m}}),\tag{19}
$$

 T_{kv} is the virtual temperature, defined as $1.01 \times (T)^{\circ}C +$ 273) (Allen et al., [1998\)](#page-1251-0).

The psychrometric constant γ is 66 Pa K⁻¹. As *ET* is defined as the evapotranspiration from a large surface of constant vegetation height, we need to consider the deviation from that idealized assumption at our site. To be pragmatic, we consider it as potential evapotranspiration and use a calibration factor f_{ET} to take into account the different surfaces as a property of the surface. Based on H_{sens} , we are now able to calculate H_{ET} , again using the Bowen ratio (Equation [14\)](#page-1241-0).

$$
H_{\rm ET} = f_{\rm ET} \times \frac{\epsilon \times \Delta H_{\rm vap}}{T_{\rm kv} \times R_{\rm s}} \times \frac{u_{\rm m}}{208} \times (e_{\rm surf} - e_{\rm m}).\tag{20}
$$

While the vapor pressure at ground surface, e_{surf} , is calculated via the Kelvin equation, the vapor pressure at $2 \text{ m}, e_{\text{m}}$, is determined from temperature T and humidity at 2 m, using the ideal gas law, directly in the simulation code.

Using Equations (20) – (12) we can now solve the heat balance [\(11\)](#page-1241-0) and assign meaningful boundary conditions.

Water

With ET defined in Equation [\(15\)](#page-1241-0), the water balance

Incoming water = Precipitation - Evapotranspiration (ET), (21)

can be solved as precipitation intensities were measured.

The basic balances are valid at the transition from air to soil. The asphalt cover can be described with its material parameters. For the land-use cover, we need to add an additional resistance, because the grass acts as an additional resistance. This resistance can be calculated as by Equation (19), but with the bulk surface resistance r_s instead of the aerodynamic resistance. The equation for the resistance λ of the grass cover is given below, where the value of r_s is set to 71 m s⁻¹ according to Allen et al. [\(1998\)](#page-1251-0).

$$
\lambda = \frac{T_{\rm kv} \times R_{\rm s}}{\epsilon \times \Delta H_{\rm vap} \times \gamma} \times r_{\rm s}.
$$
 (22)

2.2.7 | Initial conditions

The system is sensitive to the initial conditions. Sensors were covering depths between 0.45 and 1.15 m, while values between 0 and 6 m were needed. Our initial approach

was to interpolate the measured temperatures from 1.15 m below ground surface linearly down to the groundwater table (where the groundwater temperature is fixed to 10˚C; see Section [2.2.6\)](#page-1241-0) and to assume a constant temperature between 0.45 m below ground surface and ground surface. This did not succeed, as we observed rising temperatures in lower layers at the beginning of calculation time, which seemed to have no physical reason. However, starting simulations with 1 year initialization period solved the problem, as shown in Figure S2. This fits to results presented in Yu et al. [\(2019\)](#page-1252-0), where the uncertainties of initial conditions are well documented.

2.2.8 | Calibration of the model

The required model parameters are known to different degrees and are associated with different uncertainties. As already indicated in the description of the boundary conditions above, some model parameters, such as permeability of the asphalt, must also assume "apparent" values due to the use of a 1D simplification, in order to integrate unconsidered multidimensional effects into the model through the back door, so to speak.

The properties of the fluids water (and air, though not modeled explicitly) are well known and their implementation in the model can be trusted. In particular the hydraulic properties in porous-media flow are usually associated with great uncertainty and variability. In this study, also the parameters for the thermal balance required attention since a sophisticated model was elaborated as explained above.

Some of the required parameters were measured, others are based on literature values (because measuring would have been too complicated). As the simulation time was short (approximately 20 min for one run), we calibrated parameters and finally arrived at parameter sets in good agreement with measurements. We note that the calibration was performed based on expert judgement without sophisticated algorithms or correlation analyses.

Assigning proper values to parameters for heat intake was difficult, as common databases such as Stephan et al. [\(2019\)](#page-1252-0) describe the material as one continuous solid at one fixed water saturation, not considering different saturations. In fact, in the partially saturated soil we encounter varying saturations and need to model property changes with saturation. Bertermann et al. [\(2018\)](#page-1251-0) give an overview on the change of soil properties based on their water contents. Farouki [\(1981\)](#page-1251-0) gives values for "pure" materials. These were used as average values for the parameter calibration for density, heat capacity, and solid thermal conductivity. The set of calibrated parameters, which we finally used as well as their values of variation are listed in Table [1.](#page-1243-0) The capillary pressure–saturation curves for **TABLE 1** Overview over parameter values used in simulations.

the porous media at the pilot site with final calibrated values of Table 1 are shown in Figure S3.

The temperature in the topmost region is highly sensitive to the fitting parameter for evapotranspiration, f_{ET} , and the albedo value ω , while transport parameters become relevant further below. Based on that, we first fitted f_{ET} and ω to the temperature measurements in the topmost 30 cm, upon which heat transport parameters were varied, that is, heat conductivity and heat capacity. Parameters with high sensitivity include the permeability, the moisture content, which is closely coupled to porosity. The used set of parameters can be found in Table 1).

3 RESULTS

3.1 Measured data

The results of the measurements, which we interpret as the meteorological forcing functions for the temperature evolution in the different stations of our experimental pilot site are given in Figure [2.](#page-1244-0) They include (from top to bottom) long wave and short wave radiation, the temperature at a height of 2 m above ground, the wind velocity at 2 m height, the air humidity, and the precipitation. The data show the transition from the summer (i.e, August 2020) to the winter (i.e., December 2020) with associated trends of decreasing temperatures and radiation intensities as well as increasing air humidity.

Figure [3](#page-1244-0) shows the temperature time series for the four stations at the experimental pilot site separated into individual panels for discrete vertical depths. Gaps in the bottom panel indicate missing data. The land-use cover at Stations I and IV is natural vegetation (grass), while it is asphalt at Stations II and III. The subsurface material is naturally occurring silty clay at Stations I and II, and is gravelly material at Stations III and IV. The data show that the type of land-use cover has a larger influence on the evolution of subsurface temperatures than the porous material in the subsurface at our pilot site. This is indicated by small temperature differences between the green curves and the grey curves in Figure [3.](#page-1244-0)

FIGURE 2 Meteorological forcing functions: Data from the nearby weather station.

FIGURE 3 All measured soil temperatures over time.

3.2 The match between measurements and numerical model

The main results of this study is the calibration of the numerical model to the measured data, which is presented below.

Figure [4](#page-1245-0) shows two subplots of subsurface temperature evolutions during a time period in September 2020 at the four stations. Subplot (a) contains the measured temperatures and in comparison to that there is subplot (b) with the calibrated corresponding numerical simulation results. From (a), the dynamics in the evolution of the vertical temperature profiles is visible. The color coding indicates the time from September 1 (purple) to September 21 (yellow). A larger spatial spread can be observed in this figure's central panels (Stations II and III), which represent the asphalt land-use, than in the most left and most right panel (Stations I and IV), which are both with grass as land-use. This indicates that, at least for this design of the experiments, the type of land-use has a stronger influence on temperature evolution than the soil properties. The asphalt-covered gravel section (Station III) spreads even more than the corresponding natural material

FIGURE 4 Evolution of temperature profiles in the subsurface. (a) Temperature over depth for several days, *measured*, (b) Temperature over depth for several days, *simulated*. Legend dates are dd.mm.yyyy.

underneath (Station IV), which proves that there is certainly also an effect due to the different layers below the top cover. Generally, the model is able to reproduce measured soil moisture dynamics at various depths and under various land-use as well.

Figure 4 contrasts precipitation data (top panel) with both measured and simulated water saturations in two different depths, that is, at 60 cm (middle panel) and at 100 cm (bottom subplot). The different stations I–IV are represented in the curves according to their labeling in the legend with grass or asphalt as covers and natural material or gravel underneath underneath. Easy to remember, green curves represent grass land-use and, accordingly, black represents asphalt. Dashed lines indicate simulated curves. Our calibrated model (panel b) is able to reproduce the larger temperature spread under asphalt than under grass with depth and its temporal evolution very similarly at the observations (panel a). The influence of precipitation in the curves observed in the experiments at the pilot site is smaller at deeper regions. Relatively small precipitation events, even where they occur over multiple consecutive days (e.g., during October 2020) are not detected in the saturation data. In contrast to that, the simulated curves show the impact of precipitation on water saturation, that is, soil moisture, is more finely nuanced.

Overall, during the period from August until December 2020, subsurface temperatures decline as expected following the trend of declining seasonal average temperatures in the fall. The deeper below ground surface the more are the amplitudes of temperature fluctuations attenuated. Figure [6](#page-1246-0) shows how the calibrated model reproduces the observed temperatures in good qualitative agreement; the accuracy of the match between calibrated model and observation tends to be better in the deeper regions close to the drinking-water pipes. The curves for the natural material underneath with asphalt/grass on top (Stations I/II) is plotted on the left, accordingly Stations III/IV with gravel underneath are seen on the right. Asphalt cover is plotted in black and grass cover in green.

FIGURE 5 Precipitation data and corresponding measured versus simulated saturations over time. Dashed lines show simulated curves.

FIGURE 6 Simulated (after calibration) and measured temperatures at different depths versus time for all stations.

The amplitudes of temperatures at the interface between soil and the meteorological are dampened with increasing depth and are not detectable anymore at a depth of 115 cm. Up to depths of about about 1 m, the short-term high temperature amplitudes are dampened and occur at the pilot site with a delay of about 1 to 2.5 days.

There is a noticeable mismatch between measured and simulated curves for the asphalt cover at the very early times, which can be attributed to the still hot asphalt after its pouring. This mismatch diminishes by early September. The periods where the temperature data is not plotted continuously, gaps marks those time periods where there are no correct

FIGURE 7 Soil temperatures plotted over time and height with total heat as forcing function (top panels).

FIGURE 8 Contour plots of water saturations over time in the soil with total water influx as forcing function.

boundary conditions available, particularly due to the erroneous long wave sensor. To be able to feed the model with required boundary conditions, we used the long-wave data of the year 2015, which exhibited similar statistical moments in those periods. Obviously, the simulated temperatures mostly

follow the measured trends, but start with a bigger difference after periods with missing "correct" radiation data. Looking at absolute temperature differences, we notice that they are smaller than ≈ 8 K in the beginning at higher positions, while at lower positions they are below 1 K. The temporal

FIGURE 9 Performance plots of measured versus simulated temperatures. Top panels for entire time series, bottom panels for a smaller temperature range exclude the early time when the temperatures were still biased by freshly installed asphalt.

FIGURE 10 Standardized cumulative frequency of simulated temperatures.

evolution of temperature differences between measurements and observations is shown in Figure S6.

Figures [7](#page-1247-0) and [8](#page-1247-0) are intended to give an impression of the dynamics of heat and moisture transport in the soil body as obtained from the numerical model. At the top of these time-map illustrations, we provide the meteorological forcing functions, that is, total heat and total water, respectively. The heat forcing is reflected directly by the temperature response in the model domain. Comparing Stations I and IV (both grass cover, but with different layers underneath), we observe a difference in the temperature distributions: in gravel, the extents of zones of elevated temperatures are larger with smaller gradients. The gravel systems, as expected, react faster than the clayey zones. This difference in heat transport dynamics is smaller for Stations II and III (with asphalt). Regarding the soil moisture dynamics as expressed by the time maps of water saturation in Figure [8,](#page-1247-0) we can see that the gravel

layer acts as an efficient drainage (green zones). While there is no strong difference in soil moisture dynamics between asphalt and grass cover for the back-filled material, the gravel and grass cover combination tends to lead to a drier soil body.

Focusing on the difference of the stations, all simulated temperatures were summed and plotted as standardized cumulative frequency of T in Figure 10. Grass-covered stations (green lines) are to the left of the other lines in this plot, which indicates that grass-covered regions have a tendency to be colder. In the range of to 20˚C, there are differences in slope, which means that the difference of temperatures is not always the same. The type of land-use cover plays an even more important role during hot days (Figure S7).

We conclude this results section with a detailed quantitative assessment of the match between measured and simulated temperatures as given by Figure 9. The performance plot analyses the deviations between model and experiment for different depth zones. The top row considers the entire time period, while in the bottom row the first month was left out. It was mentioned already above that we expect here a bias due to freshly poured asphalt and, thus, elevated temperatures which are not due to meteorological forcing functions. The model can reproduce the trends of the measured data, that is, the order of temperatures, higher temperature spreading under asphalt cover. The match between model results and observation data is generally good, while the deviations vary in a range of $-\pm 2$ K for the entire period (upper panels). When excluding the first month from this analysis, the performance is always better than ± 2 K, while a small consistent overestimation $(\sim+2 K)$ by the model can be noticed.

FIGURE 11 Evaluating the effect of the variably-saturated material properties. (a) Three sets of possible capillary pressuresaturation relationships, (b) Simulated saturations (contoured over depth and time) for the different sets, (c) Simulated temperatures at various depths for the different sets.

4 DISCUSSION

4.1 The agreement between measurement and simulation (part A)

The results show, as summarized, for example, in Figure [9,](#page-1248-0) that calibrated numerical simulations are in satisfactory, partly even excellent agreement with the measured data. Thus, we can claim with confidence that the model is able to not only qualitatively, but also quantitatively is able to describe on a seasonal time scale the governing physical processes that

FIGURE 12 Comparison of time series of simulated and measured temperatures with varying temporal resolution in input data.

lead to temperature evolution in the depths of drinking-water pipes as driven by meteorological forcing functions in terms of temperature curves and precipitation data. We note that the calibration was not globally optimized and no correlation metrics were calculated. For example, the simulated results are sensitive to the van Genuchten parameters for relative permeability and capillary pressure. We have tested several sets of parameters and evaluated them with respect to best fitting the measured data. We have found that for the overall match between data and simulation, it is in particular important to match saturations in the lowest and topmost regions and to adapt the van Genuchten parameters accordingly, as shown in Figure 11a.

4.2 The agreement between measurement and simulation (part B)

The calibrated 1D model had to cope with some effects that are obviously multidimensional, which inevitably means that the calibrated parameter values are not in all instances the "real" values. For example, the asphalt cover is ideally close to impermeable, while there has been some small amount of water infiltrating from laterally, see also the discussion on boundary conditions.

4.3 On initial conditions

Furthermore, the difficulties with a good set of initial conditions has been mentioned, which was addressed here with

a simulated initialization period. The materials were backfilled into the trench and had to equilibrate with ambient conditions for some time. It is also important to note that the measurements started while the asphalt still was fresh and the temperatures were still elevated for a couple of days. This explains why the agreement between simulation results and measurements improves at later time periods.

4.4 On the influence of the boundary conditions

Grass is not sufficiently homogeneous for practical measurements of variably saturated parameters and asphalt is not permeable. Water intrusion does not occur through the asphalt layer itself, but, since it has limited lateral extent, from the sides, so that the situation modeled as 1D must in reality be considered at least 2D. Thus, the material underneath the asphalt is in fact not completely dry, which means, that a representative set of variably saturated parameter values at the top boundary had to be found in order to cope with that situation. Alternatively, the system had to be modeled in 2D, which would have increased computational costs tremendously. Instead of that, we were fitting van Genuchten parameters and permeability such that the resulting saturation fitted to observations.

The boundary conditions in the 1D system strongly affect the moisture distribution in the porous media. The two coarse layers transport all water faster. For achieving a good fit in saturation, it is important to work with values, which have at all times a better conductivity than sand. Parameter studies also showed that the natural material below the test side needs to have saturation values that are different from those of the back-filled material. As the deeper material was not manipulated and might also change in greater depths, this seems reasonable.

4.5 On the length of simulation period

We modeled only the cooling phase, as other values were not available, when we started. This leaves the possibility, a calibration would have led to different parameters with a longer period. For this reason, we are at the moment working on longer time periods.

4.6 On the heat balance solved in this application

The solution of the heat balance coupled to the moisture transport employs a novel and sophisticated approach to consider the soil–atmosphere interface. This involves a number of parameters, which need to be determined, partly based on idealized assumptions, which may not be given as ideal as assumed. This concerns in particular the assumption of a large surface of constant vegetation height, as required for calculating the evapotranspiration, see Equation [\(16\)](#page-1241-0). In order to have a tool to adapt the assumption to the reality, we introduced a calibration parameter, f_{ET} . Thus, one might question the approach with respect to possible over-parameterization or over-sophistication for such a small surface as we have it in our application. On the other hand, the coupled, calibrated solution, gave very satisfactory agreement and the model helps including meteorological parameters to foster the better understanding of processes and parameters, which is considered as important as the reliable prediction of temperatures.

In addition, such sophisticated models require reliable meteorological time series over long periods. As our study showed, sometimes sensors fail and backup sensors, double measurements, etc. would be useful.

4.7 On the temporal resolution of the data as input for the modeling

Regarding the temporal resolution of the measured data in the numerical model, we note that a daily precipitation time series was chosen instead of the available hourly series. The reason for that is that the daily series is more smooth and does not force the model to resolve sudden high peak values, which led to very strong changes in saturation, in particular for the gravel. Relative to the observation time, we assume that a daily average does not introduce significant errors in water content distribution over depth, as can be seen in Figure [12.](#page-1249-0)

5 CONCLUSIONS

The transient dynamics of water saturation and temperatures in the shallow subsurface, dependent on land-use cover and back-fill material were modeled with quite some confidence when meteorological data are available. Beyond this study, the prediction of seepage water temperatures has relevance for further applications as in geothermal heating/cooling systems especially in urban environments. The influence of heat balances coupled to water transport, thus including also evapotranspiration, is in particular high due to the water's very high latent heat of vaporization.

The land-use cover has the biggest effect both on temperatures and saturations in the subsurface, while the soil structure is also relevant, but not as much.

Good initial data at all depths have proven to be crucial. They need to be either known and if (partially) missing be obtained by an initialization period in the model to allow for the conditions in the layer to adapt to meteorological data.

To achieve a good quality of predicted temperature distribution, it is important to have correct saturation data. The model reacts sensitively on saturation parameters and radiation data, while density, heat capacity, and thermal conductivity are more robust.

More calibration and validation efforts are needed. We plan to [calibrate also the spring season, where temperatures tend](https://darus.uni-stuttgart.de/dataset.xhtml?persistentId=doi:10.18419/darus-3554) to increase, and thus to further increase the confidence in the model. This will allow for modeling climate-change scenarios to derive suggestions for resilient design criteria for drinkingwater supply networks.

AUTHOR CONTRIBUTIONS

Elisabeth Nissler: Conceptualization; formal analysis; data curation; investigation; writing—original draft; visualisation. **Samuel Scherrer**: Conceptualization; methodology development; software. **Holger Class**: Conceptualization; methodology; writing—review and editing. **Tanja Müller**: Pilot site construction; investigation; resources; data curation. **Mark Hermannspan**: Pilot site construction; investigation; resources; data curation. **Esad Osmancevic**: Conceptualization; methodology; resources; writing—review and editing; funding acquisition. **Claus Haslauer**: Conceptualization; methodology; resources; writing—review and editing; supervision; project administration; funding acquisition.

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CONFLICT OF INTEREST STATEMENT

The authors declare no conflicts of interest.

DATA AVAILABILITY STATEMENT

At publication time the code and the data will be made available as a FAIR [\(https://www.izus.uni-stuttgart.](https://www.izus.uni-stuttgart.de/en/fokus/darus/) [de/en/fokus/darus/\)](https://www.izus.uni-stuttgart.de/en/fokus/darus/) DaRUS repository. The code and its results are in the dataset [\(https://darus.uni-stuttgart.de/](https://darus.uni-stuttgart.de/dataset.xhtml?persistentId=doi:10.18419/darus-3552) [dataset.xhtml?persistentId=doi:10.18419/darus-3552\)](https://darus.uni-stuttgart.de/dataset.xhtml?persistentId=doi:10.18419/darus-3552) DockerContainer, where dumux can be run via a docker-container and the results are documented for the four stations as tabulated values. Data, used for Boundary Conditions is available at (https://darus.uni-stuttgart.de/dataset.xhtml?persistentId=

doi:10.18419/darus-3554) Meteorologic data, while measured soil moisture and temperature data, used for validation is to be found at [\(https://darus.uni-stuttgart.de/dataset.xhtml?](https://darus.uni-stuttgart.de/dataset.xhtml?persistentId=doi:10.18419/darus-3555) [persistentId=doi:10.18419/darus-3555\)](https://darus.uni-stuttgart.de/dataset.xhtml?persistentId=doi:10.18419/darus-3555) Soil Data.

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SUPPORTING INFORMATION

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